1	The Global Distribution of Biological Nitrogen Fixation in Terrestrial Natural
2	Ecosystems
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10	Key Points:
11	• Evanotranspiration and productivity (e.g. NPP) are unreliable predictors of terrestrial
12	biological nitrogen fixation at the global scale
12	• Free living biological nitrogen fivation makes up at least a third of the terrestrial total
17	 Global terrestrial biological pitrogen fixation is likely in the range of 52 120 TeN.
14	• Global terrestrial biological introgen fixation is likely in the range of $52 - 150$ rgiv
15	year .
10	Abstract
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19	Biological nitrogen fixation is a key contributor to sustaining the terrestrial carbon cycle,
20	providing nitrogen input that plants require. However, the amount and global distribution of
21	this fixation is highly disputed. Using a comprehensive meta-analysis of field measurements
22	we make a new assessment of global biological nitrogen fixation (BNF). We assessed the
23	relationship between BNF in natural terrestrial environments and empirical predictors of BNF
24	commonly used in terrestrial ecosystem and earth system models. We found no evidence for
25	any statistically significant relationship between BNF and evapotranspiration and net or gross

26	primary terrestrial productivity (NPP or GPP). We assessed the relationship between BNF
27	and 11 other climate variables and soil properties at a global scale. We found that all the
28	variables we considered had little predictive power for BNF. Using averaged biome values
29	upscaled we calculated the median global inputs of BNF in natural ecosystems as 88 TgN yr
30	¹ . The range $(52 - 130 \text{ TgN yr}^{-1})$ encompasses most recent estimates and broadly agrees with
31	recent independent top-down estimates of BNF. The global values indicate a significant role
32	for free-living, as opposed to symbiotic, BNF, accounting for at least a third of all BNF. This
33	work provides a new global benchmark and spatial distribution dataset of BNF using a
34	bottom-up methodology.
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37	1 Introduction
38	
39	The terrestrial carbon cycle is an important contributor to the uptake of atmospheric carbon,
40	removing about a third of anthropogenic carbon emissions from the atmosphere (Le Quéré et
41	al., 2018). Carbon fixation in the terrestrial biosphere is dependent on chlorophyll, of which
42	nitrogen is a key component. But while supply of nitrogen (N) is critical, inorganic N is water
43	soluble and therefore is prone to being washed out of soils (Davis, 2014) or lost via gaseous
44	pathways (Lenhart et al., 2015). One of the key questions for future projections of terrestrial
45	carbon uptake is to what extent N will be available to enable increased growth under high
46	carbon dioxide conditions (Davies-Barnard et al., 2015; Zaehle, et al., 2014). Biological
47	nitrogen fixation (BNF) is one source of new N represented in models. These models are
48	dependent on knowing what the current supply of BNF is, how it is spatially distributed, and
49	what are the environmental drivers of BNF. Models not only need process-based
50	understanding, but global datasets to test the model on. This work aims to provide the latter.

52	Two primary classes of BNF can be distinguished: symbiotic (also known as associative, or
53	nodulating fixation) (Granhall, 1981), and free-living (also known as non-symbiotic or
54	asymbiotic) (Reed et al., 2011). Both are found in ecosystems worldwide, to a lesser or
55	greater extent. Symbiotic BNF can be defined as fixation in association with higher plants in
56	the form of root nodules (Granhall, 1981). Fixing plants are mostly part of the legume family
57	(Fabeaceae) and include a full span of sizes and forms, from small plants like clover
58	(Trifolium), shrubs such gorse (Ulex), to trees such as alder (Alnus). Questions exist about to
59	what extent symbiotic N ₂ fixing plants are facultative or obligate fixers, and estimates of
60	fixation can vary accordingly (Menge et al., 2009; Sheffer et al., 2015).
61	
62	Free-living fixation encompasses a huge range of organisms in virtually all parts of terrestrial
63	ecosystems. In vegetated ecosystems free-living fixation can be found in soil, litter, woody
64	debris, and plant canopies, as well as in bryophytes (mosses) and lichen (Reed et al., 2011).
65	Even where vegetation is sparse, fixation is found in cyptogamic covers (Elbert et al., 2012).
66	Valid discussions surround the usefulness of free-living and symbiotic as categorisations, as
67	they are neither consistent nor discrete. BNF associated with bryophytes has been shown to
68	be symbiotic (Adams & Duggan, 2008) but is usually classified as free-living, and free-living
69	fixers are more phylogenetically diverse than makes a logical grouping (Reed et al., 2011).
70	Use of these well-known classifications is a helpful shorthand for us in this context, but there
71	is much more nuance to this issue than we present here. The relative contribution of
72	symbiotic and free-living fixers to global BNF is an ongoing debate that we aim to shed some
73	light on.

75 Existing global terrestrial estimates of BNF from natural sources provide a substantial range 76 with little sign of consensus emerging over time. There are a range of different methods, which can be roughly categorized into three groups. Some global BNF estimates are top-77 78 down estimates using a N budget (e.g. Vitousek et al., (2013)), which takes known global values of carbon to nitrogen ratios and ¹⁵N and uses these to calculate the required BNF. 79 Other top-down methods can be less empirical global budgets (e.g Delwiche, (1970)). There 80 81 are number of 'bottom-up' estimates, of which Cleveland et al., (1999) is the best known. These use a meta-analysis of available field measurements then upscale biome averages to a 82 83 global total. The majority of global BNF values are models or model and data combinations. These include field data in conjunction with models (e.g. Wang & Houlton, (2009)), new 84 data-informed models (e.g. (Xu-Ri & Prentice, 2017)), and existing models with new 85 86 predictive data (e.g Galloway et al., (2004)).

87

The seminal meta-analysis of BNF done by Cleveland et al., (1999) established an empirical 88 89 relationship between evapotranspiration (ET) or net primary productivity (NPP) and nonagricultural BNF. This relationship has been used by many terrestrial carbon-nitrogen models 90 91 (Bloh et al., 2018; Goll et al., 2017; Koven et al., 2013; Smith et al., 2014) as well as other estimates of total global BNF (Cleveland et al., 2013; Galloway et al., 2004). A fuller 92 93 description of the range of common model calculations of BNF can be found in Zaehle, et al., 94 (2014) or Meyerholt et al., (2016). As the most data-based BNF estimate available, the global 95 spatial distribution of BNF based on Cleveland's model of ET is sometimes used as 'observations' for comparison with model output (e.g. Meyerholt et al., (2016)). Therefore, 96 97 the reliability of the relationship between NPP and ET and BNF is important to accurate modelling. Given that 20 years of new field measurements are now available and there is 98

99 continued uncertainty about the global total and spatial pattern of BNF, a new 'bottom-up'100 assessment is timely.

102	This paper aims to give a new comprehensive insight into the empirical relationship between
103	BNF in natural ecosystems and a range of related variables. We consider linear modelling to
104	establish the relationship between BNF and soil and climate variables. We also use a
105	upscaled biome approach using land cover groupings to provide global total and spatial
106	distribution BNF estimates based only on measured data. We conclude by comparing our
107	new global calculations to previous assessments of global BNF.
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110	2 Methods
111	
112	We reviewed over 300 papers and books and collected information about the N_2 fixation,
113	fixer type, latitude and longitude, and vegetation type. This gave over 550 entries. We
114	exclude some measurements, including some used in previous studies, because they do not
115	meet our standard for reliability. Our overarching principles for inclusion are:
116	
117	- The measurement must be stated by the author in annual units. Therefore, values that
118	are per hour or per day or representative of short-scale measurements, are not scaled
119	up to annual estimates and used here. We do include scaled-up measurements if the
120	author has themselves calculated an annual value, as we assume the author judges the
121	measurements to be sufficiently representative. Where a measurement is given for the
122	entire growth season, we include it as representative of the whole year.

Values must be in comparable units of N. Values only given in C₂H₄ (ethylene) or 123 124 C_2H_2 (acetylene) are therefore excluded because conversion between C_2H_4 and N is 125 variable (Ley & D'Antonio, 1998; Nohrstedt, 1985; Saiz et al., 2019). However, 126 where the author has made the conversion, we accept their scientific judgement. 127 The measurements must be representative and not anomalous. Measurements that 128 specify that they are the maximum, represent an uncommon soil or vegetation type, 129 are noted by the authors as being unreliable, or similar provisos, are excluded from 130 the analysis.

Values must be from the primary source. The practice of using numbers cited in
reviews, other secondary material, or from unpublished data increases the risk of
transcription errors. Therefore, we only include values verifiable in the primary
source. For that reason, we include in our dataset the precise location (e.g. page
number) of the data within the source.

Measurements must distinguish the source of BNF to some extent. Being unable to
specify the source of the BNF is suggestive of unreliable methods, for instance

budgets that assume an amount of BNF. Where the BNF sources are not differentiated

139 it is possible that issues such as including non-biological nitrogen fixation (e.g.

140 weathering), could be present, overestimating the amount of BNF.

141 - The values must be field measurements, not 'guesses', 'estimates', or values deduced

142 from carbon or nitrogen budgets. Though we understand the useful role these

143 estimates had in previous work, it is difficult to be sure that they are accurate,

144 particularly as the methods of reaching the estimates are often opaque.

Values must be weighted by the presence or cover of the plant or organism. Some
measurements assume 100% coverage of say, a fixing legume, but do not specify

147 what the level of coverage of that fixing legume is in the environment. This then

requires a highly uncertain assumption of the cover. We could not be sure that any resulting relationship was genuine or due to error in the cover percentage.

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149

This resulted in 253 usable values. A reference list of all sources used in this paper isavailable in Supporting Information section 1.

153

154 These criteria are more stringent than those used by Cleveland et al. (1999), because with 155 increased data availability comes the opportunity to discard less reliable data. The most 156 notable difference between our inclusion criteria and that of Cleveland et al. (1999) is our 157 exclusion of unweighted symbiotic measurements. Cleveland et al. (1999) includes unweighted values by averaging the available coverage percent in that biome. We could have 158 done similarly, however we found only 16 estimates of symbiotic fixing coverage across all 159 160 biomes, from 12 sources (Baker et al., 1986; Bauters et al., 2016; Blundon & Dale, 1990; Bowman et al., 1996; Cech et al., 2010; Fahey et al., 1985; Grove & Malajczuk, 1992; 161 162 Johnson & Mayeux, 1990; Kummerow et al., 1978; Menge & Chazdon, 2016; Permar & 163 Fisher, 1983; Rundel et al., 1982), which range from 0.3% (Cech et al., 2010) to 34% (Rundel et al., 1982). Given this range and the small sample size, we could not be certain that 164 wrong assumptions of symbiotic coverage would not skew the results. Therefore we only 165 166 include values where the ecosystem average is given. 167

168 The method of measurement is also a significant issue for BNF. We exclude all 'budget' type 169 estimates of BNF, where the BNF is extrapolated from measurements of say, large scale 170 deposition, nitrogen uptake, and nitrogen leaching. In principle, we include all direct

171 measurements, however the method of measurement may have some effect on the resultant

values. There are two main methods of measuring BNF: the acetylene-ethene reduction assay
method (Hardy et al., 1968) (ARA) and the ¹⁵N₂ method.

174

175 ARA works on the basis of the enzyme mainly responsible for fixation having a preference for acetylene (C_2H_2) over N₂. The amount of resultant ethylene (C_2H_4) indicates fixation, and 176 can be converted to the equivalent amount of N₂ fixed. The conversion factor of 3:1 C₂H₄:N₂ 177 178 is commonly assumed (Hardy et al., 1968), and was used by Cleveland et al., (1999) to 179 convert values not already reported in units of N. But this conversion factor varies 180 considerably, with studies suggesting anything from 1.6:1 to 5.6:1 (Nohrstedt, 1985) and 181 1:1000 to 5.363:1 (Saiz et al., 2019) with variation over space, time, and species. This disparity is the reason we do not convert C₂H₄ or C₂H₂ measurements to N. Since ARA 182 183 method measurements makes up a significant proportion of the measurements available and 184 many studies have site-specific conversion factors or use more than one method for verification, we include them. 185 186 The ¹⁵N method involves measurements based on known ratios of the stable isotope ¹⁵N to 187 the more common ¹⁴N. Measurements can use enrichment of soil with ¹⁵N or naturally 188 occurring differences. Compared to ARA, ¹⁵N methods are more expensive, but thought to be 189 190 more reliable. Other methods, such as N accumulation within plants, generally don't provide 191 estimates that are acceptable given the other restrictions listed above, but have been included 192 where appropriate.

193

194 From the stated vegetation type we matched to the most appropriate IGBP (International195 Geosphere-Biosphere Programme) Land Cover Type Classification, as used in the MODIS

- 196 land cover product from Friedl et al., (2010), shown in Table 1. We acknowledge that
- allocation of a vegetation type to an ecosystem is unavoidably a normative judgement.
- 198

Abbreviation	Name	Area (km ²)	Number of values. Notes
ENF	Evergreen Needleleaf Forest	3,849,855	65.
EBF	Evergreen Broadleaf Forest	14,136,082	57.
DNF	Deciduous Needleleaf Forest	1,516,648	3. For BNF types where
			no values for DNF are
			available the BNF value
			of ENF is allocated to
			DNF.
DBF	Deciduous Broadleaf Forest	1,195,671	22.
MF	Mixed Forest	10,233,122	8.
Shrub Cl	Closed shrublands	47,447	0. No BNF values for
			Shrub Cl are available so
			the BNF value of Shrub
			Op is allocated to Shrub
			Cl.
Shrub Op	Open shrublands	21,312,930	22.
Sav Wood	Woody savannas	10,187,798	8.
Savanna	Savannas	9,649,685	14.
Grass	Grasslands	18,449,115	18.
Wetland	Permanent wetlands	709,907	34.

Table 1. The IGBP Land Cover Type Classifications and corresponding abbreviations used.

-	Croplands	11,804,307	Allocated BNF values of
			Grass
-	Urban and built up	86,447	Allocated BNF values of
			Grass
-	Cropland/Natural vegetation	6,200,218	Allocated BNF values of
	mosaic		Grass
Barren	Barren or sparsely vegetated	19,047,032	2.
-	Snow and ice	2,974,617	Excluded from analysis.

201 Each BNF measurement is categorized to only one of the BNF types described in Table 2. 202 These types cover all the major categories frequently found in the literature. Measurements 203 are allocated to the most granular appropriate category and are not duplicated. For instance, if 204 a field study gives 3 values: for free-living, symbiotic, and the free-living and symbiotic BNF 205 combined, the value for combined free-living and symbiotic would be disregarded and only 206 the separate symbiotic and free-living values would be used. Some types of free-living BNF 207 have been grouped for simplicity, particularly the blue-green algae values which are allocated 208 into soil BNF (see Table 2).

209

Unless otherwise specified, we use a single representative value of BNF from each physical
location and BNF type. Where a range rather than a single value is given, the middle value of
the range is used. If a range and a 'best estimate' is given and the best estimate is not the
middle value, we use the 'best estimate'.

214

Table 2. Description and abbreviations for different BNF type categories and the number ofmeasurements in each category.

Abbreviation	Description	Number
		of values
S	Symbiotic values of BNF.	47
FL-ud	Measurements of BNF from an undifferentiated source or mix	71
	of sources of all free-living sources in that environment. In	
	these values there is no distinction between the amount of	
	BNF from different sources of free-living BNF. Where	
	different types are measured separately, they are allocated to	
	one of the below appropriate categories. This category	
	includes soil crusts or cryptogamic covers.	
FL-sl	Measurements of free-living fixing within soil, including	29
	green-blue algae and cyanobacteria (unless otherwise	
	specified as being associated with, for instance, moss).	
FL-lr	Measurements of free-living fixing within leaf litter.	26
FL-ln	Measurements of free-living fixation associated with lichens.	33
	This is assumed to be weighted by the area covered by lichens	
	in the environment measured. Where the values are stated or	
	believed to be unrepresentative of the average lichen cover in	
	the environment, these values are excluded.	
FL-ms	Measurements of free-living fixation in association with	26
	bryophytes. This is assumed to be weighted by the area	
	covered by bryophytes in the environment measured. Where	
	the values are stated or believed to be unrepresentative of the	
	average bryophyte cover in the environment, these values are	
	excluded.	

FL-cy	Measurements of free-living fixation within the canopy,	14
	including epiphytes, leaves, tree trunks, and stems.	
FL-wd	Measurements of free-living fixing within wood on the	7
	ground or other woody debris, excluding leaf litter and stems.	

218	To create a dataset where the relationships between BNF and climate and soil variables can
219	be explored, we take the latitude and longitude associated with each value. Where the
220	location is specified in latitude and longitude in the source, this is used, and where it is absent
221	the closest point from the description is used. From this we extrapolate the following for each
222	location of a BNF value:
223	
224	Mean annual Gross Primary Productivity (GPP) based on FLUXCOM RS+METEO
225	with CRUNCEPv6 climate, average of 2000 - 2010 (Jung et al., 2017; Tramontana et
226	al., 2016)
227	• Mean annual Net Primary Productivity (NPP) from the Coupled Model
228	Intercomparison Project Phase 6 (CMIP6) model UKESM1 (United Kingdom Earth
229	System Model 1) historical simulation, r8i1p1f2, 2000 – 2010. (Available from
230	ESGF@CEDA, https://esgf-index1.ceda.ac.uk/search/cmip6-ceda/)
231	• Mean annual temperature from WATCH, average of 1980 – 1999 (Weedon et al.,
232	2011)
233	• Total annual precipitation from WATCH, average of 1980 – 1999 (Weedon et al.,
234	2011)
235	• Mean annual incoming solar radiation from WATCH, average of 1980 – 1999
236	(Weedon et al., 2011)
237	• Mean annual humidity from WATCH, average of 1980 – 1999 (Weedon et al., 2011)

238	•	Mean annual pressure from WATCH, average of 1980 – 1999 (Weedon et al., 2011)
239	•	Mean annual ET from LandFlux, average of 1989 – 2005 (Mueller et al., 2013)
240	•	Global phosphorus soil distribution (total including inorganic and organic) from
241		ORNL DAAC, NASA Earth Data. (Yang et al., 2014)
242	•	Soil bulk density from ORNL DAAC, NASA Earth Data Regridded Harmonized
243		World Soil Database v1.2 (CLM resolution) (Saxton et al., 1986; Wieder et al., 2014)
244	•	Soil Organic Content (SOC) of the dominant mapping unit ID from HWSD from
245		ORNL DAAC, NASA Earth Data Regridded Harmonized World Soil Database v1.2
246		(CLM resolution) (Saxton et al., 1986; Wieder et al., 2014)
247	•	Soil clay fraction by percent weight from HWSD from ORNL DAAC, NASA Earth
248		Data Regridded Harmonized World Soil Database v1.2 (CLM resolution) (Saxton et
249		al., 1986; Wieder et al., 2014)
250	•	Soil sand fraction by percent weight from HWSD from ORNL DAAC, NASA Earth
251		Data Regridded Harmonized World Soil Database v1.2 (CLM resolution) (Saxton et
252		al., 1986; Wieder et al., 2014)
253	•	Soil pH in water for the dominant mapping unit from HWSD from ORNL DAAC,
254		NASA Earth Data Regridded Harmonized World Soil Database v1.2 (CLM
255		resolution) (Saxton et al., 1986; Wieder et al., 2014)
256		
257	Clevel	and et al. (1999) used NPP and ET from the Century ecosystem model. We aim to use
258	satellit	e or measured data wherever possible. For ET, we use the observations-based
259	Landfl	ux data (Mueller et al., 2013). However, for NPP the situation is hampered by data
260	unavai	lability. At time of writing, the only satellite derived product of NPP, from MODIS, is
261	unavai	lable due to errors caused by persistent cloud cover biases. Pragmatically, gross
262	primar	ry productivity (GPP) is a very good proxy for NPP, as NPP is GPP minus plant

263	respiration. In models, NPP is generally approximately half of GPP, so there is reason to
264	believe the two are interchangeable for the purpose of relationship with BNF. There are well
265	established observation-based products for GPP, including from Fluxcom, which is the
266	dataset we use here. However, we appreciate the need for some direct comparison with NPP,
267	the most common variable used for BNF empirical relationships. Therefore, in the initial part
268	of our analysis we also assess NPP from one of the CMIP6 earth system models, UKESM1.
269	
270	
271	3.0 Results Overall Assessment
272	
273	The global distribution of BNF measurements (Figure 1) immediately shows the paucity of
274	data available. There is a bias towards north America and north-west Europe, with most of
275	values coming from these gross Control Eurosia, south Asia, and north Africa are particularly

values coming from these areas. Central Eurasia, south Asia, and north Africa are particularlypoorly represented.



Figure 1. All the data locations mapped, with the colour indicating the type of BNF. Since
some regions have multiple measurements associated, less values are shown than are used in
other parts of the analysis, as they are over-plotted. The numbers shown in the map relate to
the number of values within each category found within the delineated region.

283

The challenge of this dataset is further revealed by considering the range of reported BNF 284 285 values. The measurements have been separated here into either values with an upper and 286 lower limit (i.e. a range), or with only one central estimate (Figure 2 a). If the measured 287 values were evenly spread across biomes, one would expect the central single values to fall 288 somewhere between the upper and lower range values. This is not the case in all the BNF 289 types, indicating that the data amount is too low for this pattern to emerge or that the data is 290 very heterogeneous. Only FL-ln (lichens) follows the expected pattern consistently (for 291 range, inter quartile range, and median values). This suggests that any further data acquired 292 could not be entirely relied upon to conform to the pattern of current data, especially in 293 categories with small sample sizes.







297 Figure 2. a): Fixation values for each of the BNF types, grouped by lower and upper bounded ranges or a single central value (note a measurement is either represented as a range 298 299 or as a single value, but not both). For each boxplot, the midline is the median, the upper line third quartile, lower line the first quartile, and the whiskers extend up to 1.5 times the 300 301 interquartile range from the top of the box to the furthest datum within that distance. Datum 302 beyond 1.5 times the interquartile range are represented as individual points. Overlaid on the boxplots are all the individual points as a 'beeswarm' scatter. b): BNF values by latitude and 303 304 separated by type. These are the central values as described in the methods, i.e. the mean of a range or the single most representative value. BNF units for all cases are kg ha⁻¹ year⁻¹. 305

We can see that all the individual free-living BNF categories are relatively small (Figure 2a)
compared to symbiotic or FL-ud. The FL-wd category has the smallest range overall and the
largest ranges of the free-living categories are FL-cy and FL-ms, which are skewed by a
small number of outliers. The FL-ud category is not much higher in mean or median than the
individual categories of free-living BNF.

Nearly all the categories have a skew to lower values (the median is lower than the mean) and 313 314 a high tail. But overall, there is perhaps less difference between the categories than might be expected. The paradigm of more fixation by symbiosis (Cleveland et al., 1999) is difficult to 315 316 justify looking at these values, as FL-ud and S are both the categories that are highest and 317 have the largest range in values. 318 319 Looking at the values of BNF by type across latitudes (Figure 2b), there is a lack of 320 latitudinal pattern that we would expect if productivity or ET were a driver of BNF. There are 321 latitudinal clusters of measurements in the mid latitudes and near the equator, but little 322 evidence that BNF increases with decreasing latitude. Free-living BNF, that might be thought 323 to be higher or more prevalent in cooler climates given high carbon uptake by lichens and 324 bryophytes at high latitudes (Porada et al., 2014), also shows little sign of that trend. S and 325 FL-ud appear to have a peak around 40 °N, but this could be sampling error because of the 326 higher number of measurements around this latitude. 327 328

329 3.1 Statistical Modelling

330

To assess the relationship between ET and productivity versus BNF we use linear modelling which shows how well correlated two datasets are. If there were a relationship between terrestrial productivity or ET and BNF as strong as Cleveland et al. (1999) found (r² = 0.63 for ET), it should be evident in a plot of each value we have versus the GPP (or NPP) or the ET for the nearest grid cell (see methods and Figure 3). However, instead of the positive relationship we might expect, neither ET, NPP, nor GPP show any obvious relationship with BNF overall (Figure 3). For each individual BNF type the pattern is contradictory, with some

- showing positive, flat, or negative relationships. The r^2 for NPP, GPP and ET versus BNF
- relationships are low (around zero or negative) and the p-values high (>0.5). Given this, it
- 340 makes sense to widen the analysis to consider other variables.
- 341



Figure 3. BNF plotted against a) ET (mm day⁻¹), b) GPP (gC m⁻² day⁻¹) and, c) NPP (gC m⁻²
day⁻¹), colour coded by BNF type (Table 2). Dotted lines represent the linear fit of the BNF
type of the same colour.

- 347 We expanded our assessment to other variables (listed in the methods) to see which best
- 348 predict the individual types of BNF (Table 2). We can see the best predicting variable (the
- highest r^2) of each type of BNF in Figure 4. The category of BNF with the best
- environmental predictor is FL-wd (r^2 of 0.46, p-value 0.055), although this category has the

smallest sample size (7 data, see Table 2). The BNF types with the highest number of
measurements have some of the lowest r² values (e.g. FL-ud and S). This suggests that we
cannot safely assume that if there were a similar number of measurements available for FLwd the same relationship would be maintained.



356

Figure 4 The strongest relationship (highest r²) between each of the BNF types and the
climate or soil variables listed in methods. The linear fit is shown as a line in each of the plots
and the corresponding r² value is in the top left-hand corner. SOC (kg m⁻²), ET (mm day⁻¹),
Precipitation (mm yr⁻¹), Clay or Sand content (percent), Surface pressure (Pa).

Climate variables in general are not the best predictors of BNF, with only FL-wd (wood) and
 FL-ms (moss) having the highest r² from a directly climate related variable. However, with

only 7 data points, FL-wd is not compelling evidence that climate is a key driver. In contrast
to the BNF model by Cleveland et al., (1999), none of the types of BNF are best correlated
with ET or GPP.

367

368 Six of the eight BNF types are best predicted by a soil variable (Figure 4). None of these have 369 a strong predictive power and have correspondingly low p-values. However, in the context of 370 the complete lack of evidence for climatic or productivity controls on BNF, three-quarters of 371 the BNF types being best correlated with soil variables show the most promise for further 372 research.

373

374

375 3.2 Averaged Modelling

376

377 Taking the raw values and corresponding soil or climate value does not produce a clear 378 correlation, as seen in section 3.1. Cleveland et al., (1999)'s approach was to bin values using 379 biomes, then build an empirical model from these biome averages. However, there are some 380 limitations with this approach we wanted to avoid, whilst still considering if underlying patterns can be revealed. The subjectivity in the allocation of vegetation types to biomes and 381 382 the small sample sizes in some biomes are both undesirable. To avoid this, we average across 383 10 equal sized bins, bounded by the minimum and maximum for the predictive variable range of that BNF type. i.e. the predictive variable or each BNF type the bins are one tenth of the 384 385 range of the predictive variable. This provides bin averages in a similar way to biome 386 averages, but without the uncertainty and normative judgements about allocations to biome groups. However, because of the low number of values in the free-living categories 387 388 (excluding FL-ud), we combine these to make a second single free-living value, in a similar

389	approach to Cleveland et al. (1999)'s. We use the sum of all the average of free-living BNF
390	types within each bin to give an alternate FL-ud. We call this FL-sum.

392	The binning and averaging helps smooth out the variability seen in Figures 1-3 but also
393	reduces the number of values (Figure 5). None of the climate or productivity variables
394	perform well here, though the soil variables are more promising. Sand has the highest r^2
395	(0.53) for symbiotic (S) BNF but a high p-value (p=0.061). SOC has a lower r^2 (0.38) and
396	higher p-value (p=0.12). The FL-ud category also has a high r^2 for Sand content (0.79) and a
397	correspondingly low p-value (0.0044). Clay also has good predictive power for FL-ud and for
398	FL-sum and S, but the p-values are >0.01 for all.



Figure 5. The values and linear model of the binned BNF for S, FL-ud, and FL-sum (see
above) is plotted against a range of predictive variables. SOC (kg m⁻²), GPP (gC m⁻¹ day⁻¹),
ET (mm day⁻¹), specific humidity (kg kg⁻¹), precipitation (mm yr⁻¹), surface pressure (Pa),

- 404 surface downwelling shortwave (W m⁻²), mean annual surface (2m) air temperature (Celsius),
- 405 total soil phosphorus (gP m⁻²), soil bulk density (kg dm⁻³), clay or sand content (percent).

406	Note that not all-of the 10 bins have BNF values, thus for most variables there are less than
407	10 data points. The numbers on each plot represent the r^2 for the corresponding colour.

For an empirical relationship to be sound, it seems rational to expect that FL-sum and FL-ud would show the same sign of relationship. Soil Phosphorus, SOC, and Near Surface Specific Humidity are the only variables with consistency of sign and all have a negative relationship (Figure 5). In all these at least one of FL-ud and FL-sum have a negative r² value and high pvalues. So the relationships are weak, but suggestive that soil properties are more likely to be useful to predicting BNF than productivity or climate.

- 415
- 416

417 **3.2 Results Mapping**

418

As an alternative to the linear model approach just presented, we also consider a land cover 419 420 type approach for upscaling BNF similar to that used by Cleveland et al., (1999). We used the 421 allocated IGBP land cover types (see methods section) and upscaled the averaged values 422 (Table 3) to the MODIS map using the same scheme. For this we only consider the FL-ud and S categories. We chose not to attempt (as Cleveland et al. (1999) did) to sum all the 423 424 different sources of BNF by assuming all sources of BNF in all land cover types for the 425 upscaled measurements in order to increase the robustness of the results. Whereas Cleveland 426 et al., (1999) adds up the average symbiotic and free-living types into a single BNF value, we 427 keep these two separate. This enables us to see which aspects of BNF are contributing to any 428 overall pattern and establish whether the drivers could be different. Excluding FL-sum and using only S and FL-ud also means that standard statistical methods can be used. 429

431 Upscaling the biome level mean values shown in Table 3 gives a global total BNF of 102 TgN yr⁻¹ (Table 4), with 45% from free-living and 55% from symbiotic. Using the geometric 432 mean that accounts for the lognormally distributed data (Parkin and Robinson, 1993) as used 433 by Cleveland et al. (1999) the global total is 67 TgN yr⁻¹, with 33% free-living BNF. Due to 434 the large range and small number of values available, the mean +/- one standard deviation 435 gives negative values in some cases, particularly the lower free-living estimates. An 436 alternative way of looking at the spread and average of the values is to use the median and 437 interquartile range (see Table 4 and Figure 6 and 7), and this is the approach we focus on. 438 439

440 Table 3 Mean, median, and geometric mean of BNF (kg m⁻² yr⁻¹) for each IGBP land cover
441 type and the number of measured values used for each.

	ENF	EBF	DBF	MF	Shrub	Sav	Sav	Grass	Wetland	Barren
					Ор	Wood				
S mean	3.83	5.12	8.30	7.47	0.55	11.63	0.35	6.71	6.92	-
S Geo.	1.27	4.32	8.12	7.47	0.55	8.99	0.26	4.85	3.27	-
mean										
S	0.55	5.75	8.30	7.47	0.55	7.90	0.34	8.10	1.50	-
median										
FL-ud	1.28	4.25	0.73	0.50	5.89	0.65	0.59	6.06	2.40	0.68
mean										
FL-ud	0.69	2.60	0.47	0.45	2.69	0.05	0.20	2.92	1.06	0.68
Geo.										
mean										
FL-ud	1.10	3.11	0.3	0.50	2.77	0.02	0.13	5.00	1.45	0.68
median										

S	3	13	2	1	1	3	8	5	11	0
(number										
of obs.)										
FL-ud	10	16	3	2	16	4	6	5	8	1
(number										
of obs.)										

Table 4. Global totals of BNF (in TgN yr⁻¹) from symbiotic and free-living sources and in

444	percent	of	the	total.

Global total	l Symbiotic (S)		Free-living (FL-ud)		S + FL-ud
Mean	56	56%	45	44%	101
Geometric mean	44	66%	22	34%	67
Median	57	64%	31	36%	88
25% Quartile	31	59%	21	41%	52
75% Quartile	66	51%	66	49%	130

445

446 The range of values within categories varies and how much influence this has depends on the extent of that land cover globally. The Grass category is variable for both S and FL-ud but 447 important, as large areas of Eurasia and North America are categorized here as Grass as a 448 proxy for cropland (which we do not directly consider). Open Shrub and Woody Savanna 449 also have disjoints between the mean, geometric mean, and median because of their large 450 451 range and small sample size. Because of their large coverage Grass and Shrub Open 452 contribute the most to the differences between global totals for the statistical methods. Conversely, Wetland has the largest range, with a substantially higher arithmetic mean than 453

454 median for symbiotic BNF. But since Wetland covers a small area (see Table 1), this does not significantly affect the global values. 455

456



457

Figure 6. Maps of BNF using median BNF allocated to IGBP land cover types. Top row: 458 Symbiotic (S) and free-living (FL-ud) BNF combined (left) and relative proportion of free-459 460 living (FL-ud) BNF (right). Bottom row: Free-living (FL-ud) BNF (left) and Symbiotic (S) BNF (right). BNF in kg ha⁻¹ yr⁻¹. The proportion of free-living BNF is 0 (all symbiotic) to 1 461 (all free-living). 462

463

Globally, free-living is consistently smaller than symbiotic BNF, but still a major contributor. 464 465 The proportion of free-living BNF is between 34 - 48% in any of the statistical values. The 466 balance of symbiotic to free-living BNF also varies regionally (Figure 6) and similar 467 differences lie between symbiotic and free-living BNF as between different biomes. In broad

468 terms, areas absent of symbiotic BNF have higher free-living BNF and vice versa. The 469 exception is EBF and Grass which have relatively high levels of both symbiotic and free-470 living fixation, resulting in the highest BNF areas. In Figure 6 and Table 3 we can see that 471 more arid and/or cold areas tend to have higher proportions of free-living fixation. Barren 472 land has the lowest BNF as it has no symbiotic BNF but has free-living fixation in the form 473 of cryptogamic crusts. Conversely, we can see that temperate and tropical forested areas 474 generally have a higher proportion of symbiotic BNF. This would explain why NPP could be 475 a good proxy for BNF if symbiotic is assumed to be the major contribution to BNF. The low 476 amount of BNF for Savanna is difficult to confirm and the subjectivity of allocation between 477 Evergreen Broadleaf Forest, Savanna, and Open Shrubland increases the uncertainty.



479

480 Figure 7. The symbiotic (S) and free-living (FL-ud) values categorised by biome type. For481 each boxplot, the midline is the median, the upper line third quartile, lower line the first

482	quartile, and the whiskers extend up to 1.5 times the interquartile range from the top of the
483	box to the furthest datum within that distance. Datum beyond 1.5 times the interquartile range
484	are represented as individual points. Overlaid on the boxplots are all the individual points as a
485	'beeswarm' scatter. The grey x on each set of data represents the arithmetic mean.
486	
487	
488	4 Discussion
489	
490	Our central global estimate of 89 TgN yr ⁻¹ is only a little lower than most recent estimates
491	(with the notable exception of Xu-Ri and Prentice, (2017)) (Figure 8). However, ours is one
492	of only three estimates below 100 TgN yr ⁻¹ (the others being Delwiche, (1970) and Vitousek
493	et al., (2013)). This represents a distinct lowering of the likely value of global BNF. The fact
494	that Vitousek et al. (2013), using a completely independent top-down budget-method,
495	proposes a low-end-range similar to ours gives more validity to our results. The upper end of
496	our range encompasses most papers this century, but is vastly lower than Xu-Ri and Prentice,
497	(2017) and Cleveland et al., (1999), which are well outside our range as well as being
498	anomalous compared to all other global estimates. The large range emphasises the continuing
499	uncertainty of global BNF values.
500	



Figure 8. Global estimates of BNF, ordered by the publication date of the paper, plus the 502 503 results from this study. Some model estimates are included for completeness, but most are a combination of measured data and modelling. Papers cited are: Burns & Hardy, (2012); 504 505 Cleveland et al., (1999), (2013); Delwiche, (1970); Galloway et al., (2004); Hutchinson, 506 (1944); Meyerholt et al., (2016); Peng et al., (2018); Robinson & Robbins, (1970); Vitousek 507 et al., (2013); Wang Ying-Ping & Houlton Benjamin Z., (2009); Xu-Ri & Prentice, (2017). 508 The symbols and categories relate to those identified in the introduction. The 'Model' 509 category encompasses model only and a variety of model-data combinations. This figure shows all the observation-based values found in the literature and a representative selection 510 511 of the modelled BNF values.

Few previous studies' estimates consider the relative contribution of free-living BNF, but one
meta-analysis of BNF from cryptogamic crusts estimates it to be as much as 49 TgN yr⁻¹
(Elbert et al., 2012). Our calculation of all free-living BNF (which encompasses cryptogamic

crusts as well as other free-living BNF) is more modest at just 31 TgN yr⁻¹ (Table 4), but still 516 517 accounts for 36% of global BNF. Wang and Houlton, (2009) estimate 17 – 44% free-living in the tropical and extra-tropical regions respectively, broadly in line with what we found. 518 519 Cleveland et al., (1999) does not explicitly state a ratio of free-living to symbiotic fixation, 520 but their numbers suggest the free-living percentage is low. A later paper using similar data 521 found free-living BNF accounted for only 18% of global BNF (Cleveland et al., 2013). Our 522 study therefore suggests free-living fixation is a substantial contributor to BNF, possibly 523 higher than previously thought. And while it remains dubious how helpful the symbiotic -524 free-living dichotomy is, there do appear to be important spatial and process differences 525 between BNF types.

526

527 There is a slight indication from our statistical modelling that soil properties could be a 528 determinant of BNF. However, without many more field measurements the number of values 529 is not enough to do multivariate modelling. SOC, Soil Phosphorus, and Clay and Sand could 530 all help predict BNF according to our analysis. Soil properties are known to be important to 531 nitrogen limitation, as for instance young tropical soils are more nitrogen limited than old tropical soils (LeBauer & Treseder, 2008). If BNF were related to N limitation as is 532 hypothesised in some models, the relative global homogeneity of N limitation (LeBauer & 533 534 Treseder, 2008) would be consistent with our finding of BNF not having any strong global 535 pattern.

536

Soil properties provide a possible predictor of BNF with some theoretical basis. Molybdenum
and phosphorus availability are both known to affect BNF (Barron et al., 2009; Reed et al.,
2007). Similarly, older soils tend to have higher carbon content and thus soil N also increases
in accordance with a well-constrained global soil C:N ratio of 186:13 (Cleveland & Liptzin,

541 2007). Although organic N is not bioavailable to plants, mineralisation (organic to inorganic) 542 of N is a significant contributor of N in modelled simulations of the N cycle (Zaehle, et al., 2014), even though as a proportion of total soil N it is low. N mineralisation makes up more 543 544 than half of the N inputs into the global terrestrial N system according to an analysis done by 545 Cleveland et al., (2013). There has also been research suggesting the role of mycorrhizal fungi in recycling N could be larger than previously thought (Terrer et al., 2016). Therefore, 546 547 SOC could be a proxy for potentially available N. Because of the relative energetic costs, it 548 stands to reason that as available soil N increases, N₂ fixation would decrease. It also is 549 consistent with the theory of BNF being primarily an early succession feature of biomes, as 550 SOC and other soil nutrients would be low at that point.

551

The question naturally arises why our results are at odds with the neat outcomes of Cleveland et al., (1999), who found a strong positive relationship with ET and NPP. There are two key differences that account for the discrepancy: the increase in available data, and methodology differences (particularly the separation of modelling and land cover type averaged upscaling in this study). A potentially useful context for the discrepancy of results would be how robust Cleveland et al., (1999)'s analysis is and thus how similar additional results would be expected to be. For further analysis of this, see Supporting Information, section 2.

559

The lack of a relationship between BNF and productivity at a macro scale shown by these results is in contrast to agricultural systems, where legume productivity is related to total BNF (Herridge et al., 2008). However, the difference between plant scale processes and ecosystem processes can be significant. Natural ecosystems would be expected to respond differently to the single plant scale or an agricultural system because the amount of fixers is variable and determined by natural selection and competition, rather than agricultural choice. 566 Since fixation has high energetic requirements, in most ecosystems non-fixers are more 567 competitive and thus cover more area. In high productivity environments, such as tropical forest, non-fixers are therefore the main source of that high productivity. Even if symbiotic 568 569 fixation at the plant level were higher in high than low productivity environments, they might 570 still be out competed by non-fixers. In addition, it's not clear that the multitude of different types of free-living BNF organisms have the same relationship between fixation and 571 572 productivity as seen in symbiotic agricultural species. Therefore, that natural ecosystem BNF 573 does not scale with productivity simply shows the differences between natural and 574 agricultural ecosystems.

575

576 One of the issues of low BNF estimations is difficulty in closing the global N budget. We 577 know that new productivity requires nitrogen to maintain carbon-nitrogen ratios, and since 578 inorganic N is soluble there are losses from the terrestrial biosphere. High estimates of BNF 579 have sometimes been used as a convenient way to reconcile the apparent N shortage. This 580 issue is muddled by the fact that many BNF estimations, especially from models, group together all non-deposition sources of new terrestrial N. Houlton et al., (2008) estimate N 581 from weathering between 14 - 40 TgN yr⁻¹ from denudation and 3 - 23 TgN yr⁻¹ from 582 chemical. Agricultural BNF has been calculated as 50 - 70 Tg year⁻¹ (Herridge et al., 2008) 583 584 and analysis done by Vitousek et al., (2013) found that fertiliser from the Haber-Bosch process accounts for 120 Tg N yr⁻¹. The contribution of lightning to the nitrogen budget is 585 thought to be small, around 7 Tg N yr-1 (Tie et al., 2002). Cumulatively, these could make a 586 587 terrestrial N inputs large. However, the spatial distribution is very different between BNF and 588 other terrestrial N sources and model developers need to be wary of assuming a single simple equation can capture the heterogeneity of either BNF or total terrestrial N inputs. 589

591 An alternative hypothesis for how low N input from BNF could be possible is that the N is 592 not new but recycled. This could explain the incorrect paradigm of high BNF in the tropics by high nutrient recycling in the tropics. Terrer et al., (2016) suggests that mycorrhizal fungi 593 594 could be responsible for much higher levels of N cycling than previously thought, and that 595 ectomycorrhizal fungi and arbuscular mycorrhizal fungi could have different abilities to acquire N. The sample size of this study is small and has attracted considerable critique, 596 597 however mycorrhizal fungi in principle could be a process that accounts for low levels of 598 BNF where N limitation is also low.

599

600 The limitations of this analysis are mainly in the quantity of measurements available. 601 However, quality and reliability are also key. Since N₂ is the most common gas in the 602 atmosphere, small changes are difficult to measure accurately and precisely. As discussed in 603 the methods, the measurement of N uptake via the acetylene-ethene assay reduction method 604 (Hardy et al., 1968) is still prevalent despite margin for error in the conversion ratio 605 (Nohrstedt, 1985; Saiz et al., 2019). Until better methodologies are cheaply and widely 606 available, and enough new measurements are available to give reliable sample sizes and good 607 global coverage, this is the best analysis possible.

608

609

610 5 Conclusions

611

612 Upscaling available symbiotic and free-living measured values obtained from natural

ecosystems, by land cover, we estimate a median global value of 88 TgN yr⁻¹ (52 - 130 TgN

614 yr⁻¹) for BNF in natural terrestrial ecosystems. Our bottom-up estimate supports previous top-

615 down methods that show a similarly low estimate of BNF. We found that at least a third of

616 BNF comes from free-living sources. In contradiction to previous work by Cleveland et al.,

617 (1999), we found no evidence for any relationship between BNF and either

618 evapotranspiration or terrestrial productivity (NPP or GPP).

619

620 More field measurements are critical to progression of our understanding of BNF. The 621 enormous heterogeneity of BNF at every level, especially in poorly represented areas such as 622 Russia, Australia, Africa, and South East Asia make estimates uncertain. Multi-year field 623 studies across several biomes are particularly rare at present. There is also a risk of null 624 results (of no BNF being found) being left unpublished, even though absence of BNF is an 625 important result. We urge the nitrogen community to continue to make BNF measurements, 626 despite the seemingly large number already available, because without many more 627 measurements with improved spatial and temporal distribution we cannot establish a more 628 precise benchmark of BNF.

629

630 These datasets have a range of potential usages. The separate BNF type datasets and 631 empirical models presented here open the possibility for modelling of free-living and symbiotic BNF in a more nuanced way than is presently done. Statistical modelling suggests 632 soil characteristics show the most potential for an empirical relationship with BNF, which 633 634 could theoretically be useful for models. The S and FL-ud BNF single categories are most 635 important in terms of usefulness in projections and have the most measurements available. 636 However, they are also poorest predicted. This presents a considerable challenge for modelling efforts. Therefore, the spatially identified maps of BNF provide the most 637 638 opportunity by opening the possibility of comparing models to direct observational data. 639

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648	
649	
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