

1 Linking land and sea through an ecological-economic model of coral reef  
2 recreation

3 **Authors:** Kirsten L.L. Oleson<sup>1\*</sup>, Kenneth Bagstad<sup>2</sup>, Carlo Fezzi<sup>1</sup>, Megan Barnes<sup>1</sup>, Mary  
4 Donovan<sup>3</sup>, Kim Falinski<sup>1</sup>, Kelvin Gorospe<sup>4</sup>, Hla Htun<sup>1</sup>, Joey Lecky<sup>1</sup>, Ferdinando Villa<sup>5</sup>, Tamara  
5 M. Wong<sup>1</sup>

6  
7 **Affiliations (must be place where author actually did the work):**

8 1 Department of Natural Resources and Environmental Management, University of Hawai'i  
9 Mānoa, 1910 East West Road, Honolulu, HI 96822 [koleson@hawaii.edu](mailto:koleson@hawaii.edu) (\* corresponding  
10 author)

11 2 Geosciences & Environmental Change Science Center, U.S. Geological Survey, Denver, CO  
12 80225

13 3 Hawai'i Institute of Marine Biology, University of Hawai'i at Mānoa, Kaneohe, HI, 96744, USA

14 4 Department of Fisheries, Animal and Veterinary Sciences, University of Rhode Island,  
15 Kingston, RI 02881, USA

16 5 Basque Center for Climate Change, Sede Building 1<sup>st</sup> Floor, Scientific Campus of the  
17 University of the Basque Country, 48940 Leioa, Spain

18

19 Footnote **present address**:

20 CF: Department of Economics and Management, University of Trento, Via Vigilio Inama, 5,  
21 38122 Trento (Italy); LEEP, Exeter Business School, Rennes Drive, Exeter, EX4 4ST (United  
22 Kingdom)

23 MB: Biodiversity and Conservation Science, Department of Biodiversity, Conservation and  
24 Attractions, 17 Dick Perry Avenue, Kensington WA 6151 [megan.barnes@dbca.wa.gov.au](mailto:megan.barnes@dbca.wa.gov.au)

25 MD: Marine Science Institute, University of California Santa Barbara, Santa Barbara, CA 93106,  
26 USA

27 KF: The Nature Conservancy, Hawai'i Program; 923 Nu'uuanu Ave, Honolulu, HI 96817,  
28 kim.falinski@tnc.org

29 HH: EA Engineering, Science, and Technology, Inc., PBC. 615 Piikoi St #515, Honolulu, HI  
30 96814

31 JL: Lynker Technologies LLC. Marine, Ocean and Coastal Science and Information Group, 202  
32 Church Street, SE / Suite 536, Leesburg, VA 20175, USA, www.lynker.com

33 TW: Department of Botany, University of Hawai'i Mānoa, 3190 Maile Way, Honolulu, HI 96822

34

35 Highlights:

- 36 ● Model integrates social values to simulate coastal management outcomes
- 37 ● Coastal recreation benefits and management priorities vary spatially
- 38 ● Land-sea management provides best strategy overall and at most local beach sites
- 39 ● Some beaches require unique strategies to maximize benefit
- 40 ● Snorkelers prefer sites with better visibility, fish abundance, and diversity

## 41 Abstract

42 Coastal zones are popular recreational areas that substantially contribute to social welfare.  
43 Managers can use information about specific environmental features that people treasure, and  
44 how these might change under different management scenarios, to spatially target actions to  
45 areas of high current or potential value. We explored how snorkelers' experience would be  
46 affected by separate and combined land and marine management actions in West Maui,  
47 Hawai'i, using a Bayesian Belief Network (BBN) and a spatially explicit ecosystem services  
48 model. The BBN simulates recreational attractiveness by combining snorkeler preferences for

49 coastal features with expert opinions on ecological dynamics, snorkeler behavior, and  
50 management actions. A choice experiment with snorkelers elucidated their preferences for sites  
51 with better ecological and water-quality conditions. Linking the economic elicitation to the  
52 spatially explicit BBN to evaluate land-sea management scenarios provides specific guidance  
53 on where and how to act in West Maui to maximize ecosystem service returns. Improving  
54 coastal water quality through sediment runoff and cesspool effluent reductions, and enhancing  
55 coral reef ecosystem conditions, positively affected overall snorkeling attractiveness across the  
56 study area, but with differential results at specific sites. The highest improvements were attained  
57 through joint land-sea management, driven by strong efforts to increase fish abundance and  
58 reduced sediment, however, the effects of management at individual beaches varied.

59

60

61 Keywords: Bayesian Belief Network; Recreational ecosystem service; Management scenario  
62 evaluation; Land-sea interactions; Hawai'i

63

## 64 1 Introduction

65 The opportunity for recreation is an important coastal ecosystem service, particularly in places  
66 where coral reefs support thriving tourism and leisure sectors (Brander et al., 2007; Moberg and  
67 Folke, 1999; Spalding et al., 2017). This predominantly non-consumptive service sustains  
68 residents living near coral reefs and fuels a multi-billion-dollar global tourism industry  
69 (Pendleton, 1994; Spalding et al., 2017). People directly enjoy reefs when SCUBA diving,  
70 snorkeling, and fishing, while activities such as swimming, sunbathing, beachcombing, and  
71 surfing at the coast may also be reef-dependent. Particular characteristics of coral reef  
72 ecosystems, like complex structure and diverse fauna, directly impact snorkeling, diving, fishing,  
73 and even surfing user experiences (Brander et al., 2007; Principe et al., 2012). Globally, a  
74 series of studies have documented abiotic, biotic, and social features of reefs that make them  
75 valuable to people for recreation (Beharry-Borg and Scarpa, 2010; Cooper et al., 2009; Inglis et  
76 al., 1999; Pendleton, 1995) including conditions of the reef and fish, presence of charismatic  
77 megafauna, water clarity, pollution, and crowding. While visitation, visitor spending, and  
78 associated economic impacts may be easier to measure, the recreational attractiveness of reefs  
79 may be more difficult to directly measure (Principe et al., 2012).

80

81 Human impacts directly affect the attributes that make reefs most valuable for recreation.  
82 Anthropogenic stressors, both global and local, can cause widespread coral mortality that leads  
83 to rapid and hard-to-reverse shifts away from coral dominated systems (Hughes et al., 2007;  
84 Nyström et al., 2008), with cascading effects on fish abundance and diversity (Pratchett et al.,  
85 2008). Specifically, corals are threatened by extreme sea temperature anomalies that cause  
86 coral bleaching, where corals expel their algal symbionts, and if temperatures stay high for too  
87 long, this can lead to widespread mortality (Brown and Roughgarden, 1997; Hoegh-Guldberg,

88 1999). Pollution can smother corals (in the case of sediment), exacerbate coral disease (in the  
89 case of pathogens from sewage), cause algal outbreaks (in the case of nutrients), have  
90 sublethal effects that alter reef genetics, and kill coral outright (in the case of toxins, including  
91 sunscreen) (Anthony et al., 2015). Further, unsustainable levels of fish harvest can unbalance  
92 the system (Jackson et al., 2001), leading to cascading effects on important ecological  
93 processes such as herbivory (Hughes et al., 2010; Mumby and Steneck, 2008). Given the  
94 multiple and potentially synergistic and cumulative effects of stressors on reef ecosystems (Ban  
95 et al., 2014; Darling and Côté, 2008), research is needed to guide management actions aimed  
96 at understanding the boundaries for success, and the tradeoffs that exist among multiple  
97 stressors for preventing declines and enhancing recovery that leads to delivery of reef-based  
98 recreational ecosystem services (Jouffray et al., 2019; Weijerman et al., 2018).

99

100 A detailed understanding of recreationalists' preferences for coral reef conditions can help  
101 managers focus their efforts to preserve or enhance reefs so they can deliver valued ecosystem  
102 services. The recreational value of coral reefs has been widely researched in the ecological-  
103 economics literature, but, apart from a handful of exceptions where spatial methods were used  
104 (Ghermandi and Nunes, 2013; Ruiz-Frau et al., 2013; Spalding et al., 2017; van Riper et al.,  
105 2012), studies have predominantly used environmental valuation methods that are point-in-time  
106 estimates with no spatial component. Furthermore, these approaches rarely link values to  
107 specific attributes in ways that enable simulation of threats and management scenarios (one  
108 exception is van Beukering and Cesar (2004). Recreational valuation studies have historically  
109 relied on methods like contingent valuation, where respondents were asked to state their  
110 willingness to pay for certain beach attributes (Ahmed et al., 2007; Loomis and Santiago, 2013;  
111 Petrosillo et al., 2007), choice experiments, where respondents were asked to make  
112 hypothetical trade-offs amongst attributes (Beharry-Borg and Scarpa, 2010; Nunes et al., 2015;  
113 Schuhmann et al., 2013), or travel cost, where respondents' actual recreational behavior was

114 used to model willingness to pay (Ahmed et al., 2007; Ariza et al., 2012; Carr and Mendelsohn,  
115 2003; Loomis and Santiago, 2013; Zhang et al., 2015). For a review of valuation studies in  
116 islands see Oleson et al. (2018). Despite this effort, most coral reef valuation studies have not  
117 been contextualized in a manner that enables place-based management scenario analysis.

118

119 Massive efforts are dedicated to coastal management globally. Are these efforts targeting the  
120 conditions and places most valuable to society? Are they addressing stressors in ways that can  
121 support continued delivery of ecosystem services? The aim of this study is to develop an  
122 applied valuation methodology that provides specific and useful management guidance to  
123 coastal managers. Information on the perceived value of specific areas for recreation - and how  
124 these might change under different scenarios - could help communities to ensure persistence of  
125 important values and services. Specifically, we assess the benefits to recreationalists and  
126 recreation-dependent communities of potential land and marine management strategies so that  
127 managers can prioritize which actions to take, and where these actions will yield the greatest  
128 benefits. To be relevant, our approach needs to include features of the nearshore environment  
129 that land and marine management could directly or indirectly affect, as well as physical and  
130 social features that influence the value of a site, such as access and crowding. It has to be  
131 ecologically sound, based on the best scientific understanding of coral reef dynamics, while also  
132 being grounded on the user experience. Our methodology rests on a Bayesian Belief Network  
133 (BBN) to integrate multiple types of information, including expert judgment about ecological  
134 dynamics, management, and snorkeler behavior, and snorkelers' stated preferences elicited  
135 through a choice experiment. While BBNs have been used in studies of coral ecology (Franco et  
136 al., 2016; Graham et al., 2008), this is the first study to use BBNs to assess ecosystem services  
137 in coral reef systems. An ecosystem services approach is relatable to decision makers, visitors,  
138 and residents as it ties ecological conditions to human preferences and wellbeing outcomes  
139 (Tallis and Polasky, 2009; Wainger and Mazzotta, 2011; Wainger and Boyd, 2009). The novel

140 ecological-economic method we developed has the advantages of being able to model and  
141 provide spatially nuanced and policy-grounded information for conservation and resource  
142 management planning. In our spatially explicit case study we identify areas where management  
143 returns are highest, as well as specific management measures that would have the largest pay-  
144 off for popular beaches on the northwest part of the island of Maui, Hawai'i, USA.

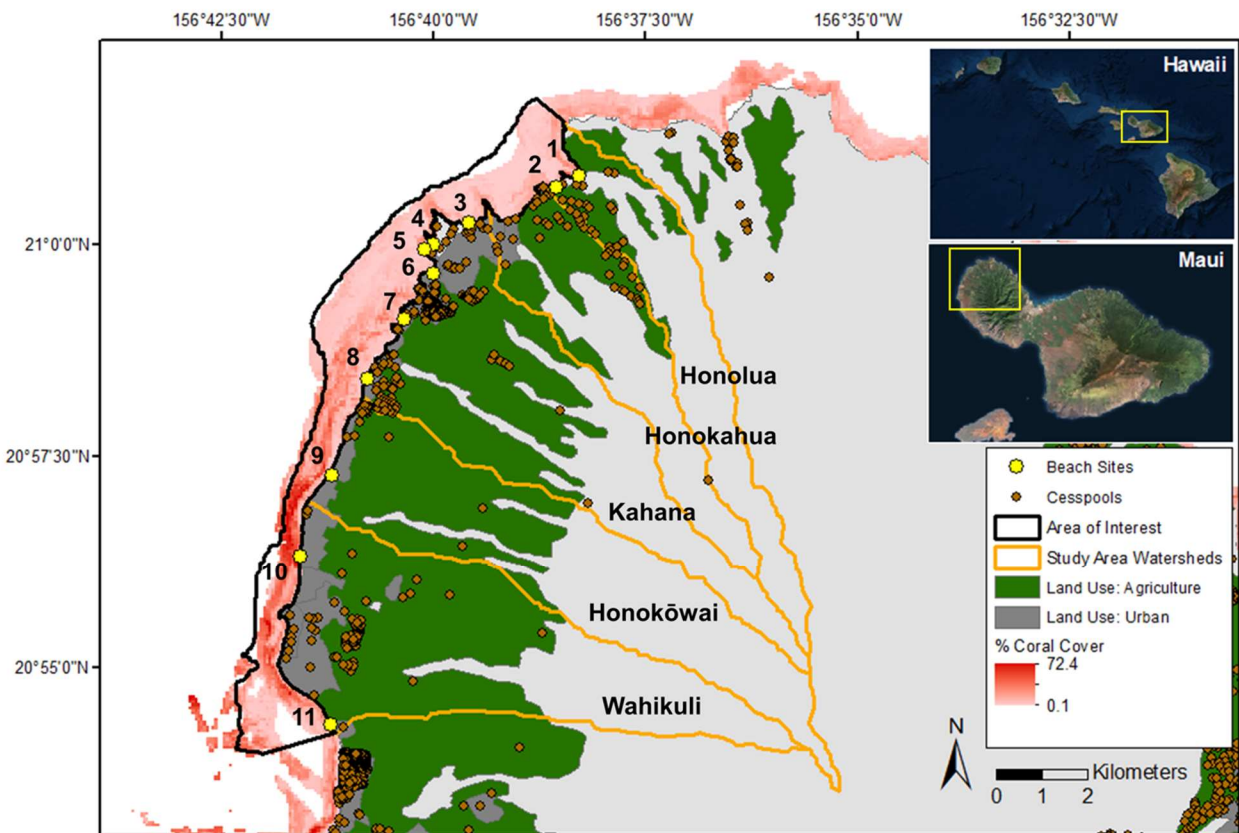
145  
146 The rest of the paper is organized as follows. In a models and methods section, we describe the  
147 site, then step through our approach, which integrates different methods and datasets, and  
148 builds scenarios. We detail the survey instrument, choice experiment, Bayesian Belief Network  
149 modeling, and scenario modeling. In each of these sub-sections we detail the method and the  
150 results, as the results are then used as inputs to the subsequent sub-section (i.e. the choice  
151 experiment results inform the BBN, which underpin the scenarios). Our discussion section  
152 focuses on the management implications, modeling innovations, and study limitations.

## 153 2 Models and Methods

### 154 *2.1 Site characteristics*

155 Over 167,000 people are residents of Maui island, in the state of Hawai'i, USA (U.S. Census  
156 Bureau, 2017). Nearly three million (2.7 million) tourists visited Maui in 2017, spending \$4.68  
157 billion (Hawaii Tourism Authority, 2016). Our case-study focuses on West Maui (Figure 1). West  
158 Maui's coasts are a popular recreation destination for tourists and residents, many of whom are  
159 attracted to the calm, clear waters and historically high-quality coral reefs. World-famous  
160 beaches in the West Maui region serve as launching spots for recreation. Today, land  
161 previously farmed as sugar or pineapple plantations for over a century is kept as fallow or being  
162 converted for residential use, while resort development continues along the coast.

163 Unfortunately, West Maui's coral reefs have declined in the past fifteen years as a result of  
164 fishing and pollution from land (Sparks et al., 2015).  
165



166  
167 *Figure 1 Map of study site with beaches, land use, cesspools, and coral reef cover depicted. Watershed boundary*  
168 *and land use from (State of Hawai'i Office of Planning, 2019)) and predicted coral cover from (Weijerman et al.,*  
169 *2018).*

## 170 2.2 Survey instrument

171 We used a tablet-based survey to collect responses from 290 recreational snorkelers in West  
172 Maui between August and September 2015. We intercepted resident and tourist snorkelers at  
173 beaches and in resort areas (Figure 1), distributing our sampling effort across five watersheds  
174 running north to south (Honolua (5% of respondents), Honokahua (8%), Kahana (22%),  
175 Honokōwai (8%), and Wahikuli (57%) based on visitation, which we estimated using a crowding



176 model based on social media photo uploads (Wood, Guerry et al. 2013). The survey instrument,  
177 approved by University of Hawai'i's Institutional Review Board (2016-31181), was tested on  
178 beach goers on a nearby island. The survey included questions related to demographics,  
179 knowledge, values, experience, and preferences for attributes of snorkeling sites. We focused  
180 on snorkelers, as snorkeling is a common activity for both residents and tourists, and snorkelers  
181 tend to be aware of environmental conditions. The design enabled us to explore possible  
182 differences between residents and tourists. The survey instrument is included as supplementary  
183 information (SI\_S1).

184

185 Full descriptive statistics are provided in Table SI\_T1. Just over half (53%) of the respondents  
186 were female. Eighty-one were permanent Maui residents, twenty were seasonal residents, and  
187 180 were visitors. The median respondent age was 45, higher than the median in the county  
188 (37), the median annual household income was \$87,500, also higher than the average in Maui  
189 County (\$72,762), and the sample was more educated than average (26.3% of 167,000  
190 residents have a college degree vs. 60% in the sample) (U.S. Census Bureau, 2017). While  
191 Maui residents are ethnically diverse, the sample was skewed towards Caucasians (65% vs.  
192 35% in Maui (U.S. Census Bureau, 2017)), likely reflecting both the tourists and the  
193 demographic who snorkels at the beaches surveyed. Most respondents reported additional  
194 snorkeling experience in locations other than Maui (240), and 40 said they had experience  
195 snorkeling on Maui. Ten noted they had no snorkeling experience and were planning on going.  
196 Snorkelers with experience had a median of 20 events. Nearly a third of all respondents (92)  
197 were also SCUBA divers.

### 198 *2.3 Choice experiment*

199 Following examples such as Schuhmann et al. (2013), we used a discrete choice experiment to  
200 determine snorkeler preferences for environmental attributes that may be affected by

201 management and/or climate change. Snorkelers were asked to choose among three different  
 202 beaches characterized by different travel costs and attributes. These attributes represent a  
 203 subset of those important for snorkeler satisfaction that were cited during interviews with experts  
 204 and local snorkelers, and reported in the literature (Beharry-Borg and Scarpa, 2010; Loomis and  
 205 Santiago, 2013; Peng and Oleson, 2017). Due to known cognitive limitations when evaluating  
 206 trade-offs in choice experiments (Johnston et al., 2017), we restricted the number of  
 207 environmental attributes included in our choice experiment to: water quality, visibility, fish  
 208 abundance and diversity, coral cover, and chance of seeing sea turtles, as well as price, which  
 209 represents both transportation costs to access the beach and the opportunity value of time  
 210 (Fezzi et al., 2014). We set three levels for each environmental attribute (Table 1), while travel  
 211 cost had six levels. The levels of all attributes were depicted in photos (Figure SI\_F1). Each  
 212 respondent faced 10 choice tasks. We validated these levels by asking respondents about their  
 213 perceptions of snorkeling on Maui.

214  
 215 A complete factorial design for our choice experiments includes all possible combinations of  
 216 attributes and levels and would use 4,374 choice tasks ( $3 \times 3 \times 3 \times 3 \times 3 \times 6 = 4,374$ ). From the total  
 217 possible combinations, 100 choice tasks with two alternative combinations of attributes and one  
 218 fixed status quo were generated in a series of ten different choice set versions (ten choice tasks  
 219 per version) in SSI Web 10.0 Sawtooth Software. Snorkelers were asked to decide between a  
 220 baseline site that represented the lowest conditions at zero cost (considered the opt-out), and  
 221 two alternative sites with improved conditions.

222  
 223 *Table 1 Attributes and levels for choice experiment*

<b>Attribute</b>	<b>Low</b>	<b>Moderate</b>	<b>High</b>	<b>Citation/Justification</b>
	<b>(Base condition)</b>			

Bacterial warning	12 days/year	6 days/year	0 days/year	(Hawai'i Department of Health, 2019) and DOH experts
Visibility	15 feet	30 feet	60 feet	NOAA experts
Coral cover	<15%	26%	>45%	(Sparks et al., 2015)
Fish abundance	75/125m <sup>2</sup>	115/125m <sup>2</sup>	150/125m <sup>2</sup>	(Friedlander et al., 2005; Williams et al., 2008)
Fish diversity	8 species	17 species	28 species	
Turtle sighting	P(sighting) = 0%	<50%	>50%	NOAA experts
Price	\$0, \$10, \$50, \$100, \$175, \$250			Estimate of cost for extra time and transportation

---

DOH = Department of Health

NOAA = National Oceanic and Atmospheric Administration

224

225 We analyzed the choice experiment data by specifying a random utility model (RUM), following  
 226 the method established by McFadden (1974). Under this framework, the utility that respondent  $j$   
 227 receives from visiting option  $i$  can be written as:

228

$$229 \quad (1) U_{ij} = \sum_{k=1}^5 \theta_{ki} + \gamma \text{cost}_i + \beta \text{SQ}_i + \varepsilon_{ij}, \quad (1)$$

230

231 Where  $\theta_{ki}$  indicates the part of utility for each of the five attributes ( $k$ ) characterizing option  $i$ ,  
 232  $\text{cost}_i$  is the cost of access,  $\gamma$  is the marginal utility of money,  $\text{SQ}_i$  is a dummy variable indicating  
 233 whether the option is the status quo,  $\beta$  is the parameter allowing for "status-quo bias," and  $\varepsilon_{ij}$  is  
 234 the random component encompassing the unobserved (to the researcher) part of the utility that  
 235 person  $i$  associates to option  $j$ . The  $\theta$  coefficients illustrate the relative importance of attributes  
 236 and their levels, and the willingness of respondents to trade one attribute level for another. To

237 allow for maximum modelling flexibility, we model each attribute via dummy variables, with the  
238 worst level for each attribute selected as the baseline (for example, for the attribute “bacterial  
239 warnings” the baseline level is 12 days per year).

240

241 Again following (McFadden 1974), by assuming the random error  $\varepsilon_{ij}$  to be identically and  
242 independently distributed as a type I extreme value (i.e., Gumbel), and indicating with  $V_{ij}$  the  
243 observed portion of the utility (i.e.,  $V_{ij} = U_{ij} - \varepsilon_{ij}$ ), we can write the probability of choosing  
244 alternative  $i$  as:

245

$$246 \quad P_{ij} = \frac{\exp(V_{ij})}{\sum_{h=1}^3 \exp(V_{ih})} \quad (2)$$

247

248 This conditional logit specification includes all the parameters in (1) and can be estimated via  
249 maximum likelihood.

## 250 Results of the choice experiment

251 We used the results of the choice experiment (below) to construct/parameterize the BBN model  
252 described below. Results of the choice experiment are summarized in Table 2. All attribute  
253 coefficients are significant. Interviewed snorkelers preferred sites with better ecological and  
254 water quality conditions, especially high and moderate visibility (coefficients 0.747 and 0.615),  
255 followed by high coral cover (0.497), high chance of sighting turtles (0.469), high bacteriological  
256 quality (0.465), and finally high fish diversity (0.379) and abundance (0.344). In many cases,  
257 most of the value to snorkelers lay in improving conditions to the moderate level from the base  
258 level; any additional improvement to the high level was less valued. This diminishing return is  
259 particularly strong in the visibility characteristic, suggesting that people were happy with being  
260 able to see 30 feet (+0.615) but the additional gains from visibility up to 60 feet were less valued

261 (+0.132). In contrast, fish diversity and abundance showed roughly linear preferences from base  
 262 conditions through moderate to high. Notably, there were few differences amongst groups.  
 263 Residents had similar preferences as tourists and seasonal residents, with one exception  
 264 (residents prioritized visibility more), although the low sample size of residents prevents  
 265 comparison of many of the attributes (Table SI\_T2).

266

267 *Table 2 Choice experiment results. Z-value is the number of standard deviations from the mean value.*

Attribute	Estimate	Std. error	z-value	
Bacteria: 0 days	0.465	0.066	7.046	***
Bacteria: 6 days	0.243	0.063	3.834	***
Visibility: 30 feet (9.14 m)	0.615	0.063	9.707	***
Visibility: 60 feet (18.29 m)	0.747	0.065	11.378	***
Coral cover: high	0.497	0.065	7.628	***
Coral cover: medium	0.304	0.061	4.962	***
Fish number: high	0.344	0.062	5.478	***
Fish number: medium	0.149	0.065	2.27	*
Fish diversity: high	0.379	0.065	5.849	***
Fish diversity: medium	0.144	0.063	2.282	*
Turtles: high	0.469	0.064	7.369	***
Turtles: low	0.234	0.066	3.543	***
Cost	-0.006	0.000	-19.164	***
Status quo	-0.658	0.112	-5.868	***
pseudo R <sup>2</sup>	0.27			
Log likelihood	-2281.83			

268 Notes: parameters need to be interpreted as differences with the baseline category, which is  
269 omitted from the model. For example, for bacteria the baseline category is 12 days in which  
270 bathing is unsafe because of potential contamination, for visibility it is 15 feet. All attributes  
271 are in Table 1.

## 272 *2.4 Bayesian Belief Network*

273 A BBN graphs the causal structure of variables in an inference or modeling problem, and uses  
274 conditional probability distributions to define relationships between variables (Aguilera et al.,  
275 2011; Ames et al., 2005). Combining diverse sources of information within a BBN is particularly  
276 important when one cannot include all attributes characterizing choices within a stated  
277 preference exercise, for well-known issues of cognitive burden (Johnston et al., 2017). BBNs  
278 have been used *inter alia* to model ecosystem services (Dee et al., 2017; Landuyt et al., 2013);  
279 and as a tool for planning (Gonzalez-Redin et al., 2016); pollution impact assessment (Spence  
280 and Jordan, 2013); guiding adaptive management (Nyberg et al., 2006); and assessing  
281 ecological water quality (Forio et al., 2015).

282  
283 Our BBN model estimates spatially explicit relative snorkeling attractiveness in the West Maui  
284 study area by integrating attributes of ecological, water, and social quality such as coral cover,  
285 fish richness, pollution, depth, and accessibility. The model's area of interest (AOI) consisted of  
286 West Maui shoreline from Honolua Bay to south of Black Rock Point, extending to 30m depth  
287 (Figure 1). The model variables, structure, and strength of relationships between variables were  
288 informed by a literature review, experts (Kuhnert et al., 2010), and the choice experiment  
289 described in the section above. Past valuation studies were useful in identifying important  
290 attributes for beach users, particularly divers and snorkelers (Grafeld et al., 2016; Parsons and  
291 Thur, 2008; Pendleton, 1994; Schuhmann et al., 2013; Wielgus et al., 2002).

292

293 Ultimately, the BBN had 11 attribute parent nodes that interact, as illustrated by the arrows, in  
 294 order to determine snorkeling attractiveness (“Snorkeling Quality” in Figure 2). Each of these  
 295 parent nodes have spatial data associated with them (Table 3) (SI, Figure SI\_F2A-K). The  
 296 current status of each attribute (i.e., prior probabilities) in West Maui is represented by the  
 297 colored bars within the parent nodes; these represent the average status across the entire AOI  
 298 and are divided into bins (Table 3, Figure 2). Parent nodes are aggregated into four  
 299 intermediate nodes (social quality, water quality, visibility, and ecological quality) that determine  
 300 snorkeling quality. The grouping of parent nodes into intermediate nodes simplifies the  
 301 conditional probabilities of the BBN model and thus reduces the cognitive load required to  
 302 determine the relationships. The selection of parent nodes and arrangement of intermediate  
 303 nodes constitutes the causal structure of the model. We tested a number of model structures via  
 304 interviews with 15 experts, including marine scientists with two Division of Aquatic Resources  
 305 staff (DAR, the state agency charged with coral reef management), a lifeguard working in the  
 306 area, ten avid snorkelers, and two snorkel tour operators.

307

308 *Table 3 Attributes in the Bayesian Belief Network (BBN)*

<b>Attributes</b>	<b>Data source</b>	<b>Measurement &amp; Bins in BBN</b>	<b>Data resolution</b>
Access	(Hawai'i Mapping Research Group, 2016; Wedding et al., 2018)	1-4 (classification)	10m
Exposure	(Wedding et al., 2018)	<5,300, >5,300 (wave energy, J*s/m)	500m
Crowding	(Wood et al., 2013)	<3, 3-6, >6 (Photograph user days)	60m
Cesspool discharge	data from (Barnes et al., 2019) using methods from (Wedding et al., 2018)	0-0.004, 0.004-0.008, >0.008 (kg N/m2)	500m

	updated, using methods from Wedding		
Sediment dispersion	et al., (2018)	0-3, 3-10, >10 (ton/ha)	30m
	(Hawai'i Mapping Research Group,		
Bathymetry	2016)	0-10, >10 (m depth)	5m
Coral cover	(Weijerman et al., 2018)	<20, 20-35, >35 (% cover)	60m
		<0.76, 0.76-1.06, >1.06	
Fish abundance	(Weijerman et al., 2018)	(count/m <sup>2</sup> )	60m
Fish species richness	(Weijerman et al., 2018)	<8, 8-17, >17 (count/grid cell)	60m
		<0.37, 0.37-0.74, >0.74	
Habitat diversity	(Friedlander and Kendall, 2006)	(ranking)	60m
Turtle chance as a	(National Centers for Coastal Ocean	0-0.35, 0.35-0.99, 0.99-1 (%)	
function of habitat	Science, 2007)	likelihood of viewing)	50m

---

Note: Probability of spotting turtles calculated as a function of habitat. High probability - coral dominated hard bottom habitat; Medium probability - algal dominated habitat (including macroalgae, turf, and crustose coralline algae (CCA)), both hard and soft bottom; Low probability - everything else - primarily uncolonized soft bottom or unknown/unclassified.

309

310 The next step was to set the relative importance of each variable via conditional probability  
311 tables. The conditional probability distribution defines the relative importance of each parent  
312 node. For instance, the intermediate node “water quality” is determined based on the value of  
313 two parent nodes, cesspool discharge and sediment dispersion. The water quality outcome is  
314 determined by specifying the likelihood that water quality is high, moderate, or low, given levels  
315 of cesspool discharge and sediment dispersion (the values of each column always sum to 1).  
316 An example conditional probability table for the water quality node is presented in Table 4. The  
317 thickness of the arrows in Figure 2, which illustrate each variable’s relative importance to the  
318 outcome, denoting average Euclidian influence, are based on the conditional probabilities  
319 (Koiter, 2006). Water quality is a relatively simple intermediate node, with only two



320 determinants; as the relationships become more complicated, the number of columns in the  
 321 tables expand very rapidly.

322

323 *Table 4 Water quality (intermediate node) conditional probability table given parent nodes Cesspool discharge and*  
 324 *Sediment dispersion.*

<b>Water Quality</b>									
Cesspool Discharge	High			Moderate			Low		
Sediment Dispersion	High	Moderate	Low	High	Moderate	Low	High	Moderate	Low
High	0	0	0.1	0	0.2	0.3	0.4	0.8	0.9
Moderate	0.05	0.1	0.1	0.6	0.6	0.6	0.4	0.2	0.1
Low	0.95	0.9	0.8	0.4	0.2	0.1	0.2	0	0

325

326 We populated the conditional probability tables based on our data from the choice experiment  
 327 and additional survey questions, as well as through consultation with coral reef managers and  
 328 experts. The choice experiment focused on a limited number of the variables (six) in the BBN to  
 329 elicit their relative importance for snorkelers in West Maui. For instance, from the choice  
 330 experiment results we understand that snorkelers in West Maui highly valued improved visibility,  
 331 more than reductions in the probability of bacteriological water quality below recreational water  
 332 standards. Features of social quality (like access and crowding) were assessed in the survey.  
 333 Interviews with experts elicited the relative importance of the other variables. Conditional  
 334 probability tables for all variables are in Table SI\_T4a and strength of influence in Table SI\_T4b.

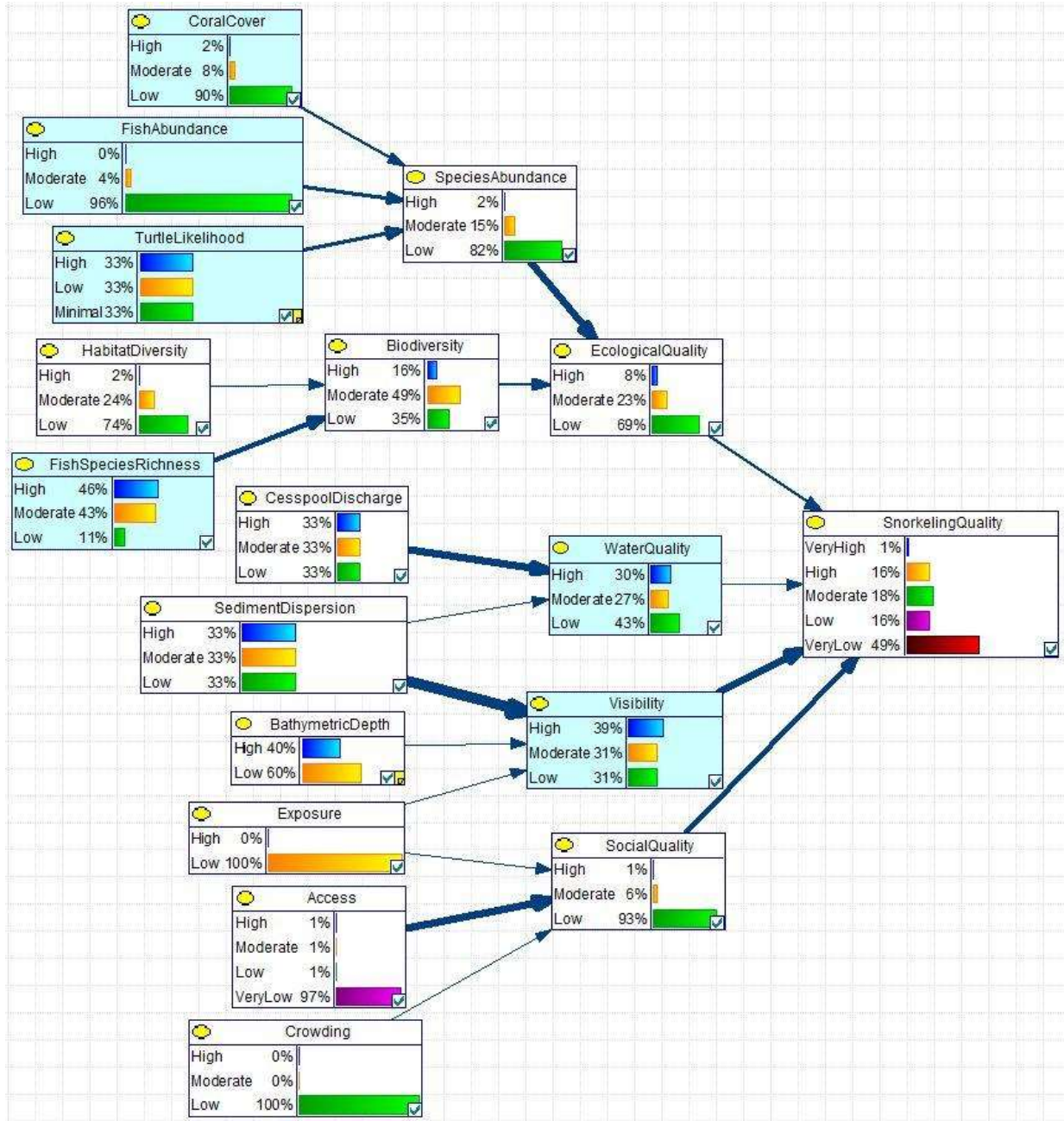
335

336 The model's output is a score (from 0 to 100) of the quality or attractiveness of each grid cell for  
 337 recreational snorkelers. A score of 100 indicates a very high-quality snorkeling site within the  
 338 study area, and 0 very poor. This score range is specific to the AOI and normalized to the range  
 339 of outcomes and scenarios in this analysis. The score is binned into five levels (0-20 very low;

340 21-40 low; 41-60 moderate; 61-80 high; and 81-100 very high). To explore assumptions of the  
341 model, we ran various hypothetical scenarios to see if the results were consistent with  
342 expectations. For instance, we set the value of model inputs that the choice experiment or  
343 experts told us were highly important (e.g., turtle-sighting likelihood, fish species richness, or  
344 visibility) to the highest possible values and evaluated the model's sensitivity to changes in  
345 these inputs, as opposed to those deemed to be less important (e.g., crowding or habitat  
346 diversity). We generated results for the entire study area, as well as for subsetting areas within  
347 the highly and moderately accessible areas surrounding popular beaches. We ran models for  
348 current conditions and a set of management scenarios (described below) at 50 m resolution  
349 using the Artificial Intelligence for Ecosystem Services (ARIES) modeling platform (Villa et al.,  
350 2014).

351

352



353

354 *Figure 2 Bayesian Belief Network describing a site's snorkeling quality. Nodes shaded in light blue indicate variables*  
 355 *included in the choice experiment. Arrow thickness denotes average Euclidian influence per the conditional*  
 356 *probability tables (strength of influence for each relationship is included in SI Table SI\_T4b). The most influential*  
 357 *relationship (Sediment Dispersion on Visibility) is about 10 times the value of the weakest relationship (Crowding on*  
 358 *Social Quality). The colored bars indicate current conditions across all pixels in the AOI in Figure 1. Note that this*  
 359 *means that most pixels are far away from the coast or near a rocky shoreline, causing the access to be very low for*  
 360 *most of them.*

### 361 3 Scenario modeling

362 A primary objective of this paper is to determine what management actions would be most  
363 effective and where their implementation would have the strongest effects. Therefore, we  
364 modeled a number of land and marine management scenarios. Land management options  
365 target sediment and effluent reduction from cesspools. Marine-based management included  
366 reducing fishing, and the effect of changes in coral cover and associated fish abundance and  
367 richness. Target levels for these reductions were based on the goals stated in official watershed  
368 management plans (Group 70, 2015a, 2015b; Sustainable Resources Group International,  
369 2012a, 2012b) and telephone, email, and in-person interviews with the watershed management  
370 coordinator, environmental consultants who prepared the watershed management plans, the  
371 State aquatic resource manager, and a Federal coral reef ecologist familiar with the area. We  
372 used four different levels for each scenario to represent increasing levels of investment in each  
373 type of management.

#### 374 Land-based management

375 In the watersheds upstream of West Maui's coral reefs, former agricultural lands currently  
376 remain fallow and access roads unfixed, stream banks continue to erode, and no cesspools are  
377 upgraded (Oleson et al., 2017; Stock et al., 2016; Whittier and El-Kadi, 2014). Land-based  
378 management scenarios represent realistic and aspirational levels of local pollution abatement.  
379 We modeled the following individually and in combination: reduce sediment input by 10%, 15%,  
380 20%, and 25%; reduce cesspool input by 10%, 25%, 50%, and 100%. Notably, we did not  
381 adjust input layers for known cesspool upgrades, and we ignored discharge from the Kahekili  
382 wastewater treatment plant.

383 Marine-based management

384 We also constructed a second set of management scenarios based on improvements to coral  
385 reef benthic habitat and associated changes in coral reef fish communities. Local coral reef  
386 experts agreed that increasing coral cover by 5%, 10%, 15%, and 20% above current levels  
387 were reasonable aspirations in this area, particularly given historical coral cover levels and  
388 improvements in managed areas (Williams et al., 2016). To estimate how fish biomass would  
389 change under different marine management scenarios, we draw upon a previously published  
390 hierarchical, linear Bayesian model of how multiple biophysical and human population drivers  
391 influence fish biomass throughout the main Hawaiian Islands (Gorospe et al., 2018). Data from  
392 the same study show that increases in coral cover would also result in increases in reef  
393 complexity (Figure SI\_F3). Therefore, although reef complexity was not a component of our  
394 snorkeler choice experiments, we use both coral cover and complexity to estimate changes in  
395 reef fish biomass. Finally, applying a linear model to data from West Maui fish surveys, we  
396 translate modeled fish biomass into the more snorkeler-relevant metrics of fish abundance  
397 (Figure SI\_F4A) and fish species richness (Figure SI\_F4B). Overall, this allowed us to derive a  
398 complete picture of how the reef attributes in the BBN (coral cover, fish abundance, and fish  
399 species richness) collectively changed (Table 5). All data for the above analyses came from fish  
400 and benthic surveys conducted by the NOAA Pacific Islands Fisheries Science Center's  
401 Ecosystem Science Division in 2012, 2013, and 2015 (Pacific Islands Fisheries Science Center,  
402 2019).

403

404 *Table 5 Model-predicted fish biomass, abundance, and species richness based on hypothetical, absolute increases in*  
405 *percent coral cover achievable with management. Using field data from throughout the main Hawaiian Islands, a*  
406 *hierarchical, linear Bayesian model (Gorospe et al. 2018) was used to predict fish biomass based on increases in*  
407 *coral cover and associated increases in reef complexity. Modeled fish abundance and richness outcomes are*  
408 *presented for different levels of absolute coral cover change over baseline, where the baseline is the current mean for*

409 *the Maui-Lahaina area. When coral reef cover increases over the baseline, the model predicts coral reef complexity*  
 410 *increase (Figure SI\_3), fish biomass, fish abundance, and fish richness. For instance, moving from baseline coral*  
 411 *cover and complexity to a scenario where coral cover increases to baseline+5%, fish biomass would increase from*  
 412 *5.89g/m<sup>2</sup> to 7.10g/m<sup>2</sup>, fish abundance from 0.028 fish/m<sup>2</sup> to 0.039 fish/m<sup>2</sup> (scenario is 139% of baseline), and fish*  
 413 *richness from 6.13 to 6.97 species (scenario is 114% of baseline).*

Coral Cover (% absolute change over baseline at a site)	Model-linked Fish Biomass	Fish Abundance		Fish Richness	
	(g/m <sup>2</sup> )	(# fish/m <sup>2</sup> )	(% of baseline)	(# species)	(% of baseline)
Baseline	5.89	0.028	NA	6.13	NA
+5	7.10	0.039	139%	6.97	114%
+10	8.33	0.050	178%	7.83	128%
+15	9.63	0.062	220%	8.74	143%
+20	10.97	0.074	263%	9.68	158%

414

#### 415 Combined marine-land management

416 As a third set of management scenarios, we combined all management outcomes into a single  
 417 scenario, where both land-based pollution was reduced and benthic habitat and fish  
 418 communities were rehabilitated at increasing levels.

#### 419 Scenario results

420 Baseline snorkeling attractiveness was estimated using the BBN under current conditions and is  
 421 mapped in Figure 3. Popular snorkeling destinations such as Ka'anapali Beach have high  
 422 snorkeling attractiveness, as expected, due to low exposure, sediment, and cesspool effluent,

423 and good ecological quality. But not all popular beaches score high. For instance Honolulu Bay  
424 has a lower than expected score, explained by high sediment, exposure, and crowding, which  
425 reduce its attractiveness, despite low cesspool discharge, high fish richness and abundance,  
426 and high probability of viewing turtles.

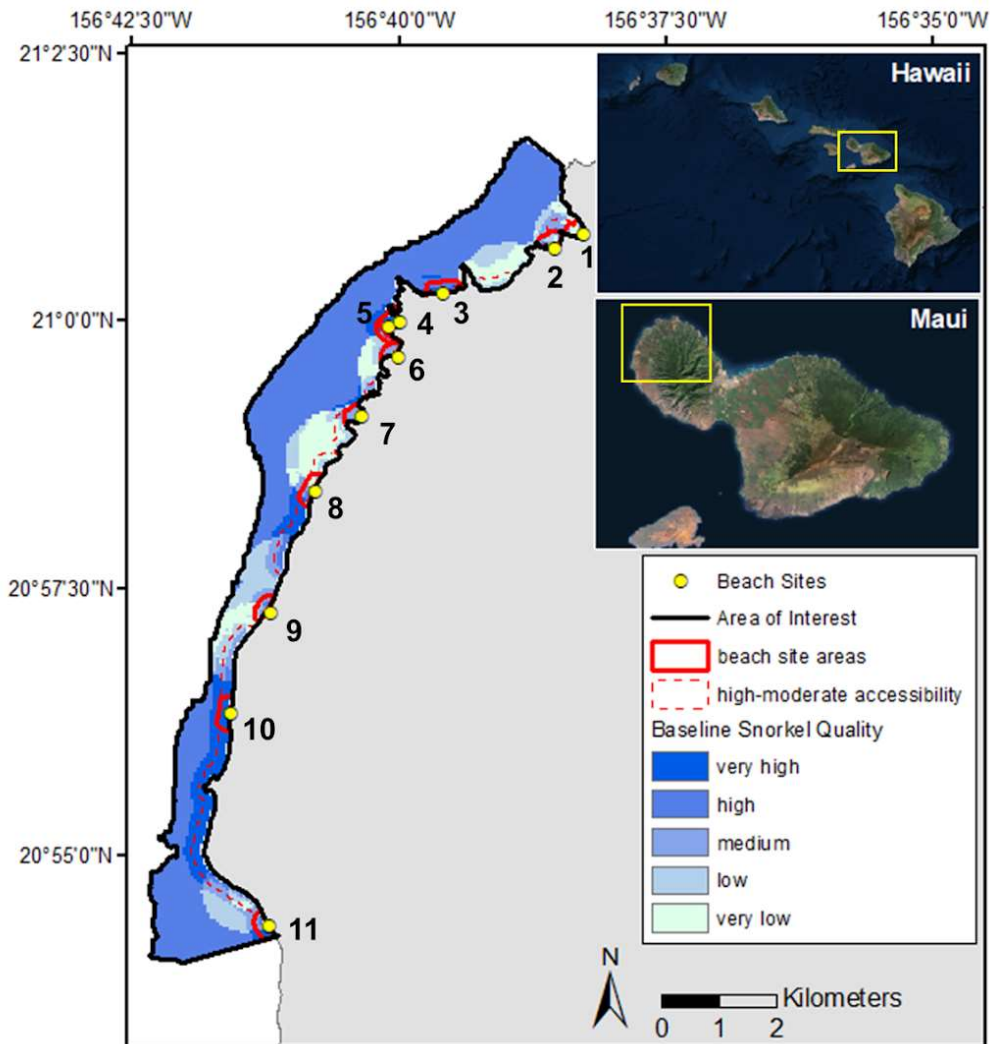
427

428 Using the BBN to estimate the effects of 20 management scenarios on recreation for the entire  
429 AOI and a subsetted area of high and moderate accessibility, we found that improving local  
430 water quality through controlling sediment and cesspool effluent and enhancing coral reef  
431 conditions (i.e., coral cover, fish abundance, fish diversity as “combined marine”) positively  
432 affected snorkeling attractiveness across our study AOI (Figure 4; Table SI\_T5). Reducing  
433 sediment alone had stronger effects on overall attractiveness than cesspool-related pollution  
434 reductions. Increasing fish abundance had the strongest effects on snorkeling quality of all  
435 ocean-related actions, while combined marine management (coral, fish abundance, and fish  
436 richness improvements) resulted in slightly larger quality improvements than combined land  
437 management (sediment and cesspool pollution reduction). Results of coral reef restoration  
438 scenarios cannot be evaluated independently, as fish abundance and richness estimates are  
439 directly tied to coral cover improvements, though we present the 12 decomposed results in  
440 Figure 4 to illustrate the relative benefits. The greatest improvements across the entire AOI and  
441 the accessible areas came from combining both land- and marine-based management.

442

443 Results of land-based scenarios suggest that sediment reductions have the most value to  
444 people, more so than cesspool effluent reductions. Reducing sediment by 25% - the highest-  
445 level erosion reduction scenario - improved the recreational value more than completely  
446 removing cesspools (7.1% vs. 4.3% improvement in the snorkeling attractiveness score for the  
447 highly and moderately accessible areas). A coordinated effort to control both sediment and  
448 cesspool effluent at the highest levels can improve the value by 11.4% in accessible areas.

449 Increasing coral cover to baseline plus 20%, fish abundance to 263% of baseline, and richness  
 450 by 158% of baseline in a combined strategy would increase snorkeling quality by 15.7% in  
 451 accessible areas. Combining all land and marine-based management activities at the highest  
 452 levels resulted in a 27.7% improvement in snorkeling quality in more accessible areas, 15.7%  
 453 from marine management and 11.5% from land management.  
 454

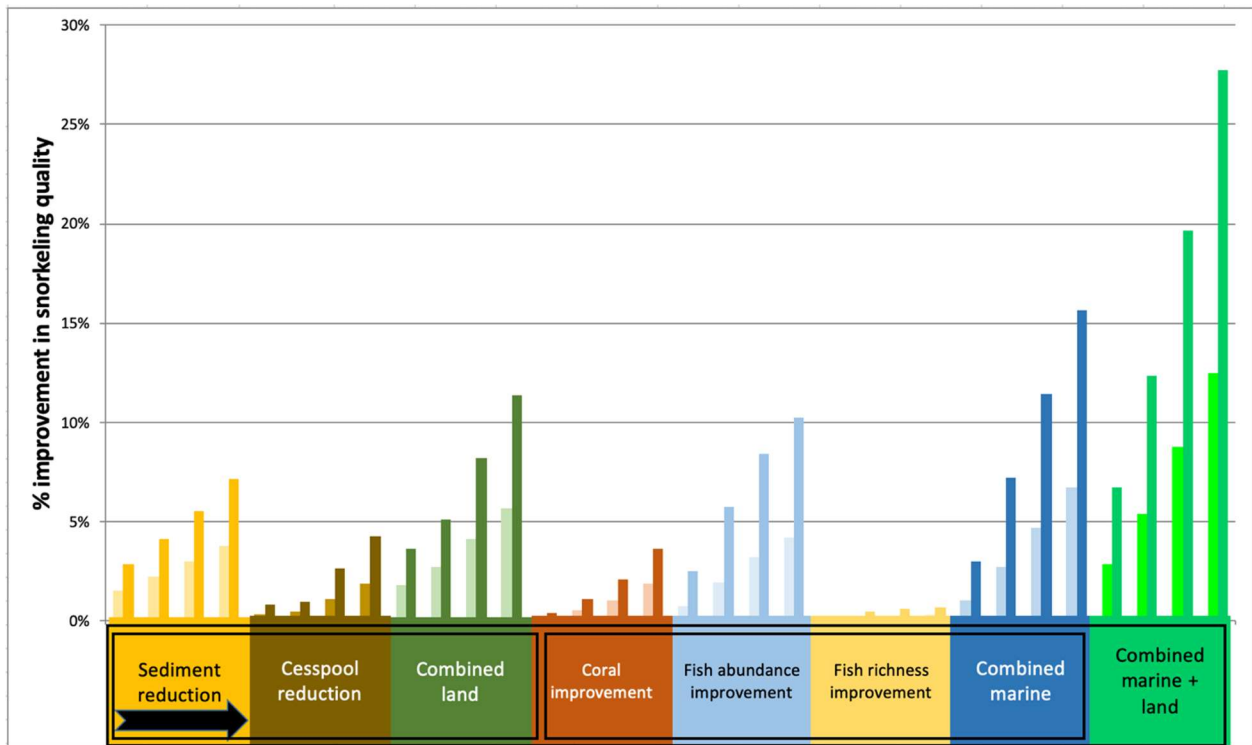


455  
 456 *Figure 3 Baseline snorkeling quality at current conditions (initial data inputs), binned as 0-20 very low; 21-40 low; 41-*  
 457 *60 moderate; 61-80 high; and 81—100 very high. Area of interest (AOI), high-moderate access area, and beach site*  
 458 *areas depicted. Beach sites indicated by yellow dots and numbers (see beach names in Table 6).*



459

460



461

462 *Figure 4 Improvement in snorkeling quality by management action/combination. Results show improvements across*  
463 *the entire area of interest (AOI) in lighter shading, and nearshore areas with high to moderate accessibility in darker*  
464 *shading. The sequence of four sets of bars for each management action shows progressively greater improvements*  
465 *for that activity, as described in the methods and Supplemental Information.*

466 Zooming in on popular local beaches illustrates how site-specific conditions determine the  
467 effects of management outcomes within the most accessible areas around those beaches.  
468 While results across the entire AOI and the most accessible areas suggest that reducing  
469 sediment is more impactful than cesspool-related action (Figure 4), this is not always true when  
470 we look at the area around popular beaches individually (Figure 3). The current recreation value  
471 of each beach area, along with results for five of the management scenarios with the largest  
472 improvements in outcomes are summarized in Table 6 for the high-access areas within 300m  
473 around eleven key beaches (see Table SI\_T6 for details and Figure SI\_F5A-FF for maps). In

474 some beaches, reducing cesspool effluent has more value than reducing sediment, and in  
 475 others, land management has no effect on recreation. As expected from the overall results,  
 476 marine management has the highest outcomes for the majority of examined beaches, higher  
 477 even than both land management actions together.

478

479 *Table 6 Snorkeling attractiveness score in highly accessible areas around each beach (listed in order north to south)*  
 480 *under baseline conditions, and relative improvements due to high-impact management scenarios: 1. reduce sediment*  
 481 *by 25%; 2. eliminate cesspools; 3. do both ["Land"]; 4. improve coral cover to baseline + 20%, fish abundance to*  
 482 *263% of baseline, and fish species richness to 158% of baseline ["Marine"]; and 5. do both "Land" and "Marine"*  
 483 *simultaneously ["Combined"]).*

Map	Beach	Baseline snorkeling attractiveness score	Snorkeling attractiveness score improvement due to management scenario				
			Sediment	Cesspool	Land	Marine	Combine d
1	Honolua Bay	25.5	1.1	0.0	1.1	7.2	8.3
2	Mokulē'ia Beach	32.5	0.0	0.0	0.0	3.7	3.7
3	Oneloa Bay	66.2	3.3	0.0	3.3	11.3	14.9
4	Hanaka'ō'ō Beach	75.4	3.1	4.2	6.4	5.1	10.8
5	Kapalua Beach	65.4	6.6	0.0	6.6	13.9	20.7
6	Nāpili Bay	36.3	6.9	5.0	10.9	4.3	14.9
7	Keonenui	36.9	6.8	11.1	17.6	16.2	33.3
8	Kahana Beach	39.7	3.7	0.0	3.7	6.3	9.1
9	Honokōwai Beach Park	34.6	0.0	1.3	1.3	7.4	8.8
10	Kā'anapali Beach	78.8	6.0	3.7	9.7	10.9	20.8
11	Wahikuli State Wayside Park	57.0	0.0	10.3	10.3	14.9	26.5

## 484 4 Discussion

### 485 *Management implications*

486 State agencies charged with protecting the environment often focus on ecological outcomes, but  
 487 the ecosystem services approach used here translates ecological conditions into terms more

488 relatable to decision makers, visitors, and residents by tying them to human wellbeing and  
489 preferences (Tallis and Polasky, 2009; Wainger and Mazzotta, 2011; Wainger and Boyd, 2009).  
490 In an era of increasingly scarce management resources and compounding threats, it is all the  
491 more important to ensure that management has net benefits. Hawai'i's economy and the  
492 Hawaiian lifestyle are tightly linked to ocean recreation, and people have positive willingness to  
493 pay for improvements to coastal amenities (Peng and Oleson, 2017; Penn et al., 2016, 2014).  
494 Our results underscore and add to the current trend globally to integrate science and  
495 management across the land-marine interface to address stressors to the ocean more  
496 holistically (Alvarez-Romero et al., 2011; Halpern et al., 2009; Pressey et al., 2007; Tallis et al.,  
497 2008; Toft et al., 2013) and efficiently (Klein et al., 2010). We introduce the human dimension to  
498 this trend: the benefits of integrated management also apply to maximizing returns to society  
499 through recreational ecosystem services.

500

501 Our approach identifies and prioritizes the many opportunities to conserve, improve, and restore  
502 recreation quality along West Maui's coast, including which actions yield the greatest  
503 improvements in snorkeling attractiveness and where these benefits will occur. Combined  
504 efforts to address land and marine problems achieve the best outcomes overall and for most  
505 beaches (Figure 4, Table 6). This aligns with recent studies in Hawai'i that have shown that  
506 addressing just one or the other (i.e., either land- or marine-based) stressors leads to sub-  
507 optimal ecological outcomes, and may even threaten ecological regime shifts (Jouffray et al.,  
508 2019; Weijerman et al., 2018). Focusing on particular beaches adds specificity to our  
509 management recommendations, highlighting the crucial need for tools to be applied at an  
510 appropriate scale. Guided by the broader scale analysis, management recommendations for  
511 West Maui as a whole are different than those coming from the local scale analysis. For  
512 instance, at some of the beaches, controlling effluent from cesspools would be more impactful  
513 than mitigating sediment (Table 6). Fortunately, recent evidence suggests that many of

514 cesspools in West Maui were upgraded by homeowners over the ensuing years since the data  
515 were collected (Barnes et al., 2019), but the importance of effluent for recreational quality, and  
516 the link between wastewater and coral degradation (Wear and Thurber, 2015), raises the need  
517 for future analysis to also consider the effects of various wastewater treatment plants along the  
518 coast.

519  
520 While the best results will generally come from integrated management, it is notable that marine  
521 management had higher payoffs overall than land management (Figure 4), driven by strong  
522 preferences for improvements in the various marine attributes, but mainly the modeled  
523 improvements in fish abundance (Table 2). The fact that fish abundance can greatly improve the  
524 delivery of recreational ecosystem services may help coastal managers, who face challenges  
525 managing for coral cover, given bleaching and other hard-to-mitigate threats, while the tools to  
526 manage fishes can be easier to implement. Further, in many places, the jurisdiction of a  
527 resource management agency may not cover both land and sea, as in the case of Hawai'i,  
528 where the Division of Aquatic Resources has jurisdiction over fisheries but not watershed and  
529 land management, which is the responsibility of other divisions within the Department of Land  
530 and Natural Resources, as well as other government departments, and water quality is the  
531 purview of the Department of Health.

532  
533 The benefits of the various management actions should ideally be weighed against their costs to  
534 determine whether action is justified, and which are the most cost-effective. These benefits may  
535 extend well beyond the recreational benefits measured here, and a full cost-benefit analysis  
536 would need to consider all costs and benefits (De Groot et al., 2013). Our results show positive  
537 preferences for improving ecosystem services, and given the scale of recreational users in  
538 Hawai'i, willingness to pay is likely more than sufficient to justify taking action, but we do not  
539 attempt to estimate the magnitude of social benefit. Different management actions will have

540 variable costs, and implementing the most cost effective (i.e., most benefit per cost) actions first  
541 will generate the greatest economic return on investment. Cesspool upgrades in the area could  
542 costs millions of dollars, while sediment reduction efforts could entail tens of millions of dollars  
543 of land restoration and infrastructure investments (Group 70, 2015a, 2015b; Sustainable  
544 Resources Group International, 2012b, 2012a). Fisheries management could have high  
545 enforcement expenses and opportunity costs for fishers and related businesses. Importantly,  
546 these costs could differ depending upon the watershed in question. Spatially explicit cost  
547 estimates to couple with the ecosystem services benefits modeled here would help decision-  
548 makers prioritize the most cost-effective actions (Naidoo et al., 2006).

549

#### 550 *Modeling innovations and limitations*

551 Our efforts contribute to an ongoing research program to evaluate ecosystem services spatially  
552 through time using big data techniques and artificial intelligence to inform management (Villa et  
553 al., 2014). An increasing number of tools use BBNs in ecosystem services modeling, including  
554 plug-ins to GIS (Landuyt et al., 2015) and stand-alone modeling platforms like ARIES, used  
555 here (Villa et al., 2014). Our innovation of linking an economic elicitation method to inform the  
556 BBN provides additional rigor to the model structure and parameterization. Specifically, we  
557 embedded the results of a choice experiment along with an expert elicitation into the BBN's  
558 structure and conditional probability tables. This enabled us to model how recreational  
559 attractiveness changes with improvements in specific, interrelated conditions. We grounded our  
560 management scenarios by eliciting reasonable outcomes for sediment and cesspool reduction  
561 and coral reef restoration from land and reef managers, and building an ecological model,  
562 based on a Hawaiian archipelago-wide dataset, to evaluate how fish conditions would change  
563 given improvements in coral cover.

564

565 The approach has some limitations. Preferences elicited from the choice experiment helped  
566 inform the conditional probabilities in the BBN. There was a design flaw that forced answers in  
567 the choice experiment, which affected the absolute, but not relative, value of the various  
568 attributes. For this reason, we do not report willingness to pay results. Our survey sample likely  
569 underrepresented residents and younger snorkelers, although no demographics exist compare.  
570 If managers are interested in examining how different management scenarios would affect  
571 different groups (e.g., tourists vs. residents; younger vs. older), then a broader survey could be  
572 conducted to build conditional probabilities (and perhaps alternate BBN structures) for these  
573 groups. Within a BBN's structure, intermediate nodes can temper or enhance the strength of  
574 influence of any given parent node on a subsequent node. For instance, in the choice  
575 experiment, snorkelers preferred fish abundance and fish species richness about the same, but  
576 in the end, fish abundance had much greater effect on overall snorkeler quality. Examining the  
577 arrows in Figure 2 that represent the strength of influence (also Table SI\_4b, fish species  
578 richness has a strong influence on the biodiversity intermediate node, but the biodiversity node's  
579 smaller contribution to the ecological quality diminishes the contribution of fish species richness  
580 to the overall snorkeling quality. Intermediate nodes are important for keeping conditional  
581 probability tables tractable, but they can have side effects of amplifying or diminishing the  
582 importance of other variables. The aim is that the combined structure and conditional  
583 probabilities are a faithful representation of the system; validation is important for ensuring this  
584 (Marcot et al., 2006). While we used expert opinion and our own intuition to validate and test  
585 assumptions of the model based on the chosen conditional probabilities, new capabilities within  
586 ARIES for BBN structural learning algorithms would be a useful, additional step (Willcock et al.,  
587 2018).

## 588 5 Conclusion

589 Natural resource managers need to know how potential management strategies are likely to  
590 impact people's wellbeing. Ecological-economic models such as the one developed here can  
591 help managers choose what actions to take where, based on the outcome's societal value. For  
592 recreational ecosystem services, the use of a BBN to combine survey-based data of the relative  
593 value of important environmental and socioeconomic features with expert opinion and spatial  
594 modeling to enable scenario analysis can provide a new path forward for integrating social and  
595 natural science with management. Such integrated modeling of coupled nature-human systems  
596 can benefit the management of recreational resources, particularly in settings with complex  
597 combinations of stressors and human uses, such as recreation and management at the land-  
598 sea interface.

## 599 6 Acknowledgements

600 Many thanks to survey team members (Lindsay Veazey, Marcus Peng), Michele Barnes for  
601 research assistance, and Derek Ford for figures. Funding was provided by Pacific Islands  
602 Climate Science Center (PICSC) award G13AC00361; USDA NIFA grants: Hatch HAW01125-  
603 H, McIntire-Stennis HAW01120-M; NOAA CRCP award NA15NOS4820209; and the National  
604 Socio-Environmental Synthesis Center (SESYNC) NSF DBI-1052875. Support for Ken  
605 Bagstad's time was provided by the USGS Land Change Science Program. The use any of  
606 trade, firm, or product names is for descriptive purposes only and does not imply endorsement  
607 by the U.S. Government.

## 608 7 Citations

- 609 Aguilera, P.A., Fernández, A., Fernández, R., 2011. Bayesian networks in environmental  
610 modelling. *Environmental Modelling & Software* 26, 1376–1388.
- 611 Ahmed, M., Umali, G.M., Chong, C.K., Rull, M.F., Garcia, M.C., 2007. Valuing recreational and  
612 conservation benefits of coral reefs - The case of Bolinao, Philippines. *Ocean & Coastal*  
613 *Management* 50, 103–118. <https://doi.org/10.1016/j.ocecoaman.2006.08.010>
- 614 Alvarez-Romero, J.G., Pressey, R.L., Ban, N.C., Vance-Borland, K., Willer, C., Klein, C.J.,  
615 Gaines, S.D., 2011. Integrated land-sea conservation planning: the missing links. *Annual*  
616 *Review of Ecology, Evolution, and Systematics* 42, 381–409.
- 617 Ames, D.P., Neilson, B.T., Stevens, D.K., Lall, U., 2005. Using Bayesian networks to model  
618 watershed management decisions: an East Canyon Creek case study. *Journal of*  
619 *Hydroinformatics* | 07, 267.
- 620 Anthony, K.R.N., Marshall, P.A., Abdulla, A., Beeden, R., Bergh, C., Black, R., Eakin, C.M.,  
621 Game, E.T., Gooch, M., Graham, N.A.J., 2015. Operationalizing resilience for adaptive  
622 coral reef management under global environmental change. *Global Change Biology* 21,  
623 48–61.
- 624 Ariza, E., Ballester, R., Rigall-I-Torrent, R., Saló, A., Roca, E., Villares, M., Jiménez, J.A.,  
625 Sardá, R., 2012. On the relationship between quality, users' perception and economic  
626 valuation in NW Mediterranean beaches. *Ocean & coastal management* 63, 55–66.
- 627 Ban, S.S., Graham, N.A.J., Connolly, S.R., 2014. Evidence for multiple stressor interactions and  
628 effects on coral reefs. *Global Change Biology* 20, 681–697.  
629 <https://doi.org/10.1111/gcb.12453>
- 630 Barnes, M.D., Goodell, W., Whittier, R., Falinski, K.A., Callender, T., Htun, H., LeViol, C., Slay,  
631 H., Oleson, K.L.L., 2019. Decision analysis to support wastewater management in coral



632 reef priority area. *Marine Pollution Bulletin* 148, 16–29.  
633 <https://doi.org/10.1016/j.marpolbul.2019.07.045>

634 Beharry-Borg, N., Scarpa, R., 2010. Valuing quality changes in Caribbean coastal waters for  
635 heterogeneous beach visitors. *Ecological Economics* 69, 1124–1139.

636 Brander, L.M., Van Beukering, P., Cesar, H.S., 2007. The recreational value of coral reefs: a  
637 meta-analysis. *Ecological Economics* 63, 209–218.

638 Brown, G., Roughgarden, J., 1997. A metapopulation model with private property and a  
639 common pool. *Ecological Economics* 22, 65–71.

640 Carr, L., Mendelsohn, R., 2003. Valuing coral reefs: A travel cost analysis of the Great Barrier  
641 Reef. *Ambio* 32, 353–357. [https://doi.org/10.1639/0044-  
642 7447\(2003\)032\[0353:vcratc\]2.0.co;2](https://doi.org/10.1639/0044-7447(2003)032[0353:vcratc]2.0.co;2)

643 Cooper, E., Burke, L.M., Bood, N.D., 2009. Coastal capital, Belize: The economic contribution of  
644 Belize's coral reefs and mangroves. World Resources Institute.

645 Darling, E.S., Côté, I.M., 2008. Quantifying the evidence for ecological synergies. *Ecology*  
646 *Letters* 11, 1278–1286. <https://doi.org/10.1111/j.1461-0248.2008.01243.x>

647 De Groot, R.S., Blignaut, J., Van Der Ploeg, S., Aronson, J., Elmqvist, T., Farley, J., 2013.  
648 Benefits of investing in ecosystem restoration. *Conservation Biology* 27, 1286–1293.  
649 <https://doi.org/10.1111/cobi.12158>

650 Dee, L.E., Allesina, S., Bonn, A., Eklöf, A., Gaines, S.D., Hines, J., Jacob, U., McDonald-  
651 Madden, E., Possingham, H., Schröter, M., 2017. Operationalizing network theory for  
652 ecosystem service assessments. *Trends in Ecology & Evolution* 32, 118–130.

653 Fezzi, C., Bateman, I.J., Ferrini, S., 2014. Using revealed preferences to estimate the value of  
654 travel time to recreation sites. *Journal of Environmental Economics and Management*  
655 67, 58–70. <https://doi.org/10.1016/j.jeem.2013.10.003>

656 Forio, M.A.E., Landuyt, D., Bennetsen, E., Lock, K., Nguyen, T.H.T., Ambarita, M.N.D.,  
657 Musonge, P.L.S., Boets, P., Everaert, G., Dominguez-Granda, L., 2015. Bayesian belief

658 network models to analyse and predict ecological water quality in rivers. *Ecological*  
659 *Modelling* 312, 222–238.

660 Franco, C., Hepburn, L.A., Smith, D.J., Nimrod, S., Tucker, A., 2016. A Bayesian Belief Network  
661 to assess rate of changes in coral reef ecosystems. *Environmental modelling & software*  
662 80, 132–142.

663 Friedlander, A., Aeby, G., Brown, E., Clark, A., Coles, S., Dollar, S., Hunter, C., Jokiel, P.,  
664 Smith, J., Walsh, B., others, 2005. The state of coral reef ecosystems of the main  
665 Hawaiian Islands. *The state of coral reef ecosystems of the United States and Pacific*  
666 *freely associated states* 222–269.

667 Friedlander, A., Kendall, M., 2006. Fishes - Reef Fish, in: Costa, B., Kendall, M. (Eds.), *Marine*  
668 *Biogeographic Assessment of the Main Hawaiian Islands*, OCS Study BOEM 2016-035  
669 and NOAA Technical Memorandum NOS NCCOS 214. Bureau of Ocean Energy  
670 Management National Oceanic and Atmospheric Administration, pp. 156–196.

671 Ghermandi, A., Nunes, P.A.L.D., 2013. A global map of coastal recreation values: Results from  
672 a spatially explicit meta-analysis. *Ecological Economics* 86, 1–15.

673 Gonzalez-Redin, J., Luque, S., Poggio, L., Smith, R., Gimona, A., 2016. Spatial Bayesian belief  
674 networks as a planning decision tool for mapping ecosystem services trade-offs on  
675 forested landscapes. *Environmental Research* 144, 15–26.

676 Gorospe, K.D., Donahue, M.J., Heenan, A., Gove, J.M., Williams, I.D., Brainard, R.E., 2018.  
677 Local biomass baselines and the recovery potential for Hawaiian coral reef fish  
678 communities. *Frontiers in Marine Science* 5, 1–13.

679 Grafeld, S., Oleson, K., Barnes, M., Peng, M., Chan, C., Weijerman, M., 2016. Divers'  
680 willingness to pay for improved coral reef conditions in Guam: An untapped source of  
681 funding for management and conservation? *Ecological Economics* 128, 202–213.

682 Graham, N.A.J., McClanahan, T.R., MacNeil, M.A., Wilson, S.K., Polunin, N.V.C., Jennings, S.,  
683 Chabanet, P., Clark, S., Spalding, M.D., Letourneur, Y., 2008. Climate warming, marine

684 protected areas and the ocean-scale integrity of coral reef ecosystems. PLoS One 3,  
685 e3039.

686 Group 70, 2015a. West Maui watershed plan: Kahana, Honokahua, and Honolua watersheds,  
687 Strategies and Implementation Report.

688 Group 70, 2015b. Kahana, Honokahua and Honolua watersheds characterization report.

689 Halpern, B.S., Ebert, C.M., Kappel, C.V., Madin, E.M.P., Micheli, F., Perry, M., Selkoe, K.A.,  
690 Walbridge, S., 2009. Global priority areas for incorporating land-sea connections in  
691 marine conservation. Conservation Letters 1–8. [https://doi.org/10.1111/j.1755-](https://doi.org/10.1111/j.1755-263X.2009.00060.x)  
692 [263X.2009.00060.x](https://doi.org/10.1111/j.1755-263X.2009.00060.x)

693 Hawai'i Department of Health, 2019. Hawai'i Clean Water Branch (CWB) Beach Water Quality  
694 Data.

695 Hawai'i Mapping Research Group, 2016. 5 Meter Bathymetry Synthesis Grid.

696 Hawaii Tourism Authority, 2016. Fact Sheet: Benefits of Hawaii's Tourism Economy. State of  
697 Hawaii, Honolulu, HI.

698 Hoegh-Guldberg, O., 1999. Climate change, coral bleaching and the future of the world's coral  
699 reefs. Marine and Freshwater Research 50, 839–866. <https://doi.org/10.1071/mf99078>

700 Hughes, T.P., Graham, N.A.J., Jackson, J.B.C., Mumby, P.J., Steneck, R.S., 2010. Rising to the  
701 challenge of sustaining coral reef resilience. Trends in Ecology & Evolution 25, 633–642.  
702 <https://doi.org/10.1016/j.tree.2010.07.011>

703 Hughes, T.P., Rodrigues, M.J., Bellwood, D.R., Ceccarelli, D., Hoegh-Guldberg, O., McCook, L.,  
704 Moltschanowskyj, N., Pratchett, M.S., Steneck, R.S., Willis, B., 2007. Phase shifts,  
705 herbivory, and the resilience of coral reefs to climate change. Current Biology 17, 360–5.  
706 <https://doi.org/10.1016/j.cub.2006.12.049>

707 Inglis, G.J., Johnson, V.I., Ponte, F., 1999. Crowding norms in marine settings: A case study of  
708 snorkeling on the Great Barrier Reef. Environmental Management 24, 369–381.

709 Jackson, J.B.C., Kirby, M.X., Berger, W.H., Bjorndal, K.A., Botsford, L.W., Bourque, B.J.,  
710 Bradbury, R.H., Cooke, R., Erlandson, J., Estes, J.A., 2001. Historical overfishing and  
711 the recent collapse of coastal ecosystems. *Science* 293, 629–637.

712 Johnston, R.J., Boyle, K.J., Adamowicz, W., Bennett, J., Brouwer, R., Cameron, T.A.,  
713 Hanemann, W.M., Hanley, N., Ryan, M., Scarpa, R., 2017. Contemporary guidance for  
714 stated preference studies. *Journal of the Association of Environmental and Resource*  
715 *Economists* 4, 319–405.

716 Jouffray, J.-B., Wedding, L.M., Norström, A.V., Donovan, M.K., Williams, G.J., Crowder, L.B.,  
717 Erickson, A.L., Friedlander, A.M., Graham, N.A.J., Gove, J.M., 2019. Parsing human and  
718 biophysical drivers of coral reef regimes. *Proceedings of the Royal Society B* 286,  
719 20182544.

720 Klein, C.J., Ban, N.C., Halpern, B.S., Beger, M., Game, E.T., Grantham, H.S., Green, A., Klein,  
721 T.J., Kininmonth, S., Treml, E., 2010. Prioritizing land and sea conservation investments  
722 to protect coral reefs. *PLoS One* 5, e12431.

723 Koiter, J., 2006. Visualizing inference in Bayesian networks. Faculty of Electrical Engineering,  
724 Mathematics, and Computer Science, Department of Man-Machine Interaction, Delft  
725 University of Technology.

726 Kuhnert, P.M., Martin, T.G., Griffiths, S.P., 2010. A guide to eliciting and using expert  
727 knowledge in Bayesian ecological models. *Ecology Letters* 13, 900–914.

728 Landuyt, D., Broekx, S., D’Hondt, R., Engelen, G., Aertsens, J., Goethals, P.L.M., 2013. A  
729 review of Bayesian belief networks in ecosystem service modelling. *Environmental*  
730 *Modelling & Software* 46, 1–11.

731 Landuyt, D., Van der Biest, K., Broekx, S., Staes, J., Meire, P., Goethals, P.L., 2015. A GIS  
732 plug-in for Bayesian belief networks: towards a transparent software framework to  
733 assess and visualise uncertainties in ecosystem service mapping. *Environmental*  
734 *Modelling & Software* 71, 30–38.

735 Loomis, J., Santiago, L., 2013. Economic valuation of beach quality improvements: comparing  
736 incremental attribute values estimated from two stated preference valuation methods.  
737 Coastal Management 41, 75–86.

738 Marcot, B.G., Steventon, J.D., Sutherland, G.D., McCann, R.K., 2006. Guidelines for developing  
739 and updating Bayesian belief networks applied to ecological modeling and conservation.  
740 Can. J. For. Res. 36, 3063–3074. <https://doi.org/10.1139/x06-135>

741 McFadden, D., 1974. Conditional Logit Analysis of Qualitative Choice Behaviour, in: P.  
742 Zarembka (Ed.), *Frontiers in Econometrics*. Academic Press, New York.

743 Moberg, F., Folke, C., 1999. Ecological goods and services of coral reef ecosystems. *Ecological*  
744 *economics* 29, 215–233.

745 Mumby, P.J., Steneck, R.S., 2008. Coral reef management and conservation in light of rapidly  
746 evolving ecological paradigms. *Trends in Ecology & Evolution* 23, 555–563.  
747 <https://doi.org/10.1016/j.tree.2008.06.011>

748 Naidoo, R., Balmford, A., Ferraro, P.J., Polasky, S., Ricketts, T.H., Rouget, M., 2006. Integrating  
749 economic costs into conservation planning. *Trends in ecology & evolution* 21, 681–687.

750 National Centers for Coastal Ocean Science, 2007. Benthic Habitat Map - Main Hawaiian  
751 Islands.

752 Nunes, P.A.L.D., Loureiro, M.L., Piñol, L., Sastre, S., Voltaire, L., Canepa, A., 2015. Analyzing  
753 beach recreationists' preferences for the reduction of jellyfish blooms: Economic results  
754 from a stated-choice experiment in Catalonia, Spain. *PLoS One* 10, e0126681.

755 Nyberg, J.B., Marcot, B.G., Sulyma, R., 2006. Using Bayesian belief networks in adaptive  
756 management. *Canadian Journal of Forest Research* 36, 3104–3116.  
757 <https://doi.org/10.1139/x06-108>

758 Nyström, M., Graham, N.A., J., Lokrantz, J., Norström, A.V., 2008. Capturing the cornerstones  
759 of coral reef resilience: linking theory to practice 27, 795–809.

760 Oleson, K.L.L., Falinski, K.A., Lecky, J., Rowe, C., Kappel, C.V., Selkoe, K.A., White, C., 2017.  
761 Upstream solutions to coral reef conservation: The payoffs of smart and cooperative  
762 decision-making. *Journal of Environmental Management* 191, 8–18.  
763 <https://doi.org/10.1016/j.jenvman.2016.12.067>

764 Oleson, K.L.L., Grafeld, S., Van Beukering, P., Brander, L., James, P.A.S., Wolfs, E., 2018.  
765 Charting progress towards system-scale ecosystem service valuation in islands.  
766 *Environmental Conservation* 45, 212–226.

767 Pacific Islands Fisheries Science Center, 2019. National Coral Reef Monitoring Program:  
768 Stratified random surveys (StRS) of reef fish, including benthic estimate data of the  
769 Hawaiian Archipelago since 2013.

770 Parsons, G.R., Thur, S.M., 2008. Valuing changes in the quality of coral reef ecosystems: a  
771 stated preference study of SCUBA diving in the Bonaire National Marine Park.  
772 *Environmental and Resource Economics* 40, 593–608.

773 Pendleton, L.H., 1995. Valuing coral reef protection. *Ocean & Coastal Management* 26, 119–  
774 131.

775 Pendleton, L.H., 1994. Environmental quality and recreation demand in a Caribbean coral reef.  
776 *Coastal Management* 22, 399–404.

777 Peng, M., Oleson, K.L., 2017. Beach Recreationalists' Willingness to Pay and Economic  
778 Implications of Coastal Water Quality Problems in Hawaii. *Ecological Economics* 136,  
779 41–52.

780 Penn, J., Hu, W., Cox, L., Kozloff, L., 2016. Values for recreational beach quality in O'ahu,  
781 Hawai'i. *Marine Resource Economics* 31, 47–62. <https://doi.org/10.1086/683795>

782 Penn, J., Hu, W., Cox, L., Kozloff, L., 2014. Resident and tourist preferences for stormwater  
783 management strategies in O'ahu, Hawai'i. *Ocean & Coastal Management* 98, 79–85.  
784 <https://doi.org/10.1016/j.ocecoaman.2014.06.002>

785 Petrosillo, I., Zurlini, G., Corliano, M.E., Zaccarelli, N., Dadamo, M., 2007. Tourist perception of  
786 recreational environment and management in a marine protected area. *Landscape and*  
787 *Urban Planning* 79, 29–37.

788 Pratchett, M.S., Munday, P.L., Wilson, S.K., Graham, N.A.J., Cinner, J.E., Bellwood, D.R.,  
789 Jones, G.P., Polunin, N.V.C., McClanahan, T.R., 2008. Effects of climate-induced coral  
790 bleaching on coral-reef fishes - Ecological and economic consequences. *Oceanography*  
791 *and Marine Biology: an Annual Review*, Vol 46 46, 251–296.  
792 <https://doi.org/10.1201/9781420065756.ch6>

793 Pressey, R.L., Cabeza, M., Watts, M.E., Cowling, R.M., Wilson, K.A., 2007. Conservation  
794 planning in a changing world. *Trends in Ecology & Evolution* 22, 583–592.  
795 <https://doi.org/10.1016/j.tree.2007.10.001>

796 Principe, P.P., Bradley, P., Yee, S.H., Fisher, W.S., Johnson, E.D., Allen, P., Campbell, D.E.,  
797 2012. Quantifying coral reef ecosystem services. U.S. Environmental Protection Agency,  
798 Washington, DC.

799 Ruiz-Frau, A., Hinz, H., Edwards-Jones, G., Kaiser, M.J., 2013. Spatially explicit economic  
800 assessment of cultural ecosystem services: Non-extractive recreational uses of the  
801 coastal environment related to marine biodiversity. *Marine Policy* 38, 90–98.

802 Schuhmann, P.W., Casey, J.F., Horrocks, J.A., Oxenford, H.A., 2013. Recreational SCUBA  
803 divers' willingness to pay for marine biodiversity in Barbados. *Journal of Environmental*  
804 *Management* 121, 29–36.

805 Spalding, M., Burke, L., Wood, S.A., Ashpole, J., Hutchison, J., zu Ermgassen, P., 2017.  
806 Mapping the global value and distribution of coral reef tourism. *Marine Policy* 82, 104–  
807 113.

808 Sparks, R., Stone, K., White, D., Ross, M., 2015. Maui and Lāna'i monitoring report. Hawaii  
809 Division of Aquatic Resources, Maui Office, Wailuku, HI.

810 Spence, P.L., Jordan, S.J., 2013. Effects of nitrogen inputs on freshwater wetland ecosystem  
811 services—A Bayesian network analysis. *Journal of Environmental Management* 124, 91–  
812 99.

813 Stock, J.D., Falinski, K.A., Callender, T., 2016. Reconnaissance Sediment Budget for Selected  
814 Watersheds of West Maui, Hawai'i: U.S. Geological Survey Open-File Report 2015–  
815 1190. United States Geological Survey, Reston, VA. <https://doi.org/10.3133/ofr20151190>

816 Sustainable Resources Group International, 2012a. Wahikuli-Honokōwai watershed  
817 management plan volume 1: Watershed characterization, Ridge to Reef Initiative.

818 Sustainable Resources Group International, 2012b. Wahikuli-Honokōwai watershed  
819 management plan volume 2: Strategies and implementation.

820 Tallis, H., Ferdana, Z., Gray, E., 2008. Linking terrestrial and marine conservation planning and  
821 threats analysis. *Conservation Biology* 22, 120–130.

822 Tallis, H.T., Polasky, S., 2009. Mapping and valuing ecosystem services as an approach for  
823 conservation and natural-resource management. *The Year in Ecology and Conservation*  
824 *Biology* 1162, 265–283.

825 Toft, J.E., Burke, J.L., Carey, M.P., Kim, C.K., Marsik, M., Sutherland, D.A., Arkema, K.K.,  
826 Guerry, A.D., Levin, P.S., Minello, T.J., 2013. From mountains to sound: modelling the  
827 sensitivity of Dungeness crab and Pacific oyster to land–sea interactions in Hood Canal,  
828 WA. *ICES Journal of Marine Science* 71, 725–738.

829 U.S. Census Bureau, 2017. QuickFacts Maui County, Hawai'i [WWW Document]. URL  
830 <https://www.census.gov/quickfacts/fact/table/mauicountyhawaii/BZA210216> (accessed  
831 6.19.19).

832 van Beukering, P., Cesar, H.S., 2004. Ecological economic modeling of coral reefs: Evaluating  
833 tourist overuse at Hanauma Bay and algae blooms at the Kihei Coast, Hawai'i. *Pacific*  
834 *Science* 58, 243–260.



835 van Riper, C.J., Kyle, G.T., Sutton, S.G., Barnes, M., Sherrouse, B.C., 2012. Mapping outdoor  
836 recreationists' perceived social values for ecosystem services at Hinchinbrook Island  
837 National Park, Australia. *Applied Geography* 35, 164–173.

838 Villa, F., Bagstad, K.J., Voigt, B., Johnson, G.W., Portela, R., Honzák, M., Batker, D., 2014. A  
839 methodology for adaptable and robust ecosystem services assessment. *PLoS One* 9,  
840 e91001. <https://doi.org/10.1371/journal.pone.0091001>

841 Wainger, L., Mazzotta, M., 2011. Realizing the potential of ecosystem services: a framework for  
842 relating ecological changes to economic benefits. *Environmental Management*.

843 Wainger, L.A., Boyd, J.W., 2009. Valuing ecosystem services, in: McLeod, K.L., Leslie, H.M.  
844 (Eds.), *Ecosystem-Based Management for the Oceans*. Island Press, Washington D.C.,  
845 pp. 92–111.

846 Wear, S.L., Thurber, R.V., 2015. Sewage pollution: mitigation is key for coral reef stewardship.  
847 *Annals of the New York Academy of Sciences* 1355, 15–30.

848 Wedding, L.M., Lecky, J., Gove, J.M., Walecka, H.R., Donovan, M.K., Williams, G.J., Jouffray,  
849 J.-B., Crowder, L.B., Erickson, A., Falinski, K., 2018. Advancing the integration of spatial  
850 data to map human and natural drivers on coral reefs. *PLoS One* 13, e0189792.

851 Weijerman, M., Veazey, L., Yee, S., Vaché, K., Delevaux, J., Donovan, M., Lecky, J., Oleson,  
852 K.L.L., 2018. Managing local stressors for coral reef condition and ecosystem services  
853 delivery under climate scenarios. *Frontiers in Marine Science* 5, 425.

854 Whittier, R.B., El-Kadi, A.I., 2014. Human health and environmental risk ranking of on-site  
855 sewage disposal systems for the Hawaiian islands of Kaua'i, Moloka'i, Maui, and  
856 Hawai'i. University of Hawaii.

857 Wielgus, J., Chadwick-Furman, N.E., Dubinsky, Z., Shechter, M., Zeitouni, N., 2002. Dose-  
858 response modeling of recreationally important coral-reef attributes: a review and  
859 potential application to the economic valuation of damage. *Coral Reefs* 21, 253–259.

860 Willcock, S., Martínez-López, J., Hooftman, D.A.P., Bagstad, K.J., Balbi, S., Marzo, A., Prato,  
861 C., Sciandrello, S., Signorello, G., Voigt, B., Villa, F., Bullock, J.M., Athanasiadis, I.,  
862 2018. Machine learning for ecosystem services. *Ecosystem services* 33, 165–174.

863 Williams, I.D., Walsh, W.J., Schroeder, R.E., Friedlander, A.M., Richards, B.L., Stamoulis, K.A.,  
864 2008. Assessing the importance of fishing impacts on Hawaiian coral reef fish  
865 assemblages along regional-scale human population gradients. *Environmental*  
866 *Conservation* 35, 261–272.

867 Williams, I.D., White, D.J., Sparks, R.T., Lino, K.C., Zamzow, J.P., Kelly, E.L.A., Ramey, H.L.,  
868 2016. Responses of herbivorous fishes and benthos to 6 Years of protection at the  
869 Kahekili Herbivore Fisheries Management Area, Maui. *PLoS One* 11.  
870 <https://doi.org/10.1371/journal.pone.0159100>

871 Wood, S.A., Guerry, A.D., Silver, J.M., Lacayo, M., 2013. Using social media to quantify nature-  
872 based tourism and recreation. *Scientific Reports* 3. <https://doi.org/10.1038/srep02976>

873 Zhang, F., Wang, X.H., Nunes, P.A.L.D., Ma, C., 2015. The recreational value of gold coast  
874 beaches, Australia: An application of the travel cost method. *Ecosystem Services* 11,  
875 106–114.

876