

1 **Inoculum origin and soil legacy can shape plant-soil feedback outcomes for tropical**
2 **grassland restoration**

3 **Running Head:** Grassland restoration: inocula and legacy effects

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22 AMD, MSD, STM, LR, RSO wrote and edited the manuscript.

23 **Abstract**

24 Restoration techniques tailored to grasslands are needed to improve restoration effectiveness
25 in tropical landscapes. In this work, we investigated the joint effects of plant-soil legacies and
26 soil inocula from native and invaded Cerrado grasslands to test whether different microbial
27 origins affect plant-soil feedbacks in ways that can foster restoration techniques. Using two
28 grass species, we measured aboveground biomass, and several plant traits over two growing
29 cycles. Species responded differently to inocula and legacies. The legacy of the invasive
30 *Urochloa eminii* and invaded soil inocula positively affected mycorrhizal colonization. The
31 legacy of *Diectomis fastigiata* resulted in a negative self-feedback discouraging its use in
32 restoration. The invasive species success is in part due to its broad ecological niche, easily
33 coping with soil differences. Our research points to the possibility of using soil inocula and
34 legacies to aid restoration efforts in the tropics; allowing restorers to target desired functional
35 trait stimulation for a given ecosystem.

36 **Key-words:** plant-soil feedback, soil inoculation, legacy effects, invasive species, restoration
37 techniques, Cerrado.

38 **Implications for practice**

- 39
- 40 • Soil inoculation should be used for the restoration of non-forested ecosystems in
the tropics as a soil improvement tool enriching the soil with local microbes
 - 41 • Restoration practitioners need to consider carefully soil feedbacks, especially from
42 annual plants, ensuring they restore the soil as well as the vegetation
 - 43 • Soil legacies must be considered in restoration planning in order to generate long-
44 term restoration success, thus maximizing cost-benefits.
- 45

46 **Introduction**

47 One of the main challenges of current restoration efforts in tropical areas is to
48 develop cost-effective techniques for the restoration of non-forested landscapes (Silveira et
49 al., 2020). Most of the restoration knowledge currently available was designed for forests
50 (Overbeck et al, 2013); and temperate ecosystems (Hartemink 2002) knowledge for
51 grasslands and particularly tropical grasslands is largely lacking (Overbeck et al. 2015). At
52 the present, many techniques are being tested in tropical grasslands, from direct seeding
53 (Coutinho et al. 2018), to top-soil transfer (Pilon et al. 2019), with varying degrees of success
54 and applicability.

55 A relatively novel restoration technique being explored is the use of microbes to
56 foster restoration initiatives (Kardol et al. 2009; Wubs et al. 2016), given they affect plant-
57 performance above and belowground (Kardol & Wardle 2010). Top-soil inoculation is an
58 example of such technique to improve restoration (Wubs et al. 2016); consisting on the
59 transfer of small amounts of top soil, containing microbiota from a natural area, to an area that
60 will be restored. It considers that the transfer of such microbes is capable of stimulating native
61 plant regeneration. Its theoretical framework postulates that microbes present in the soil are
62 capable of improving plant colonization through multiple pathways, from pathogen
63 suppression to growth stimulation (Ortíz-Castro et al. 2009). Soil inoculation has been used as
64 an effective tool to promote growth of target grassland species in the Netherlands (Carbajo et
65 al. 2011), where it has been shown that inoculant origin is capable of directing final desired
66 environments towards grasslands or heathlands, demonstrating the strong role microbes can
67 have in community assembly (Wubs et al. 2016).

68 Associations developed by plants with microbes during their lifetime which can
69 last in the soil even after the focal plant has died, termed legacy effects (Grman & Suding
70 2010), mean the colonization history of a soil patch can influence the future outcomes of plant
71 succession (Kardol et al. 2007). Legacy effects act on soil physical, chemical, and biological
72 properties (Kardol et al. 2007; Corbin & D'Antonio 2012). Soil legacies can be an important
73 factor regulating plant community dynamics, hence, they can be used to target plant
74 community development improving the management of restored ecosystems (Wubs et al.
75 2019). Wubs et al. (2019) have shown that in temperate systems, plant community
76 composition has long-lasting effects on both plant and soil communities.

77 Plants grow in the presence of several groups of organisms in the soil, from
78 bacteria to fungi (Compant et al. 2019); altering microbial populations due to associations
79 they develop during their lifetime (Haichar et al. 2008); and developing positive, negative or
80 neutral feedbacks (Hassani et al. 2018). Of particular interest in the tropics are the arbuscular
81 mycorrhizal fungi (AMF), which allow plants to obtain nutrients from the soil, besides
82 helping in soil water flow (Delavaux 2017). Moreover, soil microbial populations are capable
83 of stimulating biomass production (Bailly & Weiskopf 2012) and pathogen suppression
84 (Beneduzi et al. 2012) or build-up (Packer & Clay 2002). Therefore, microbes can be great
85 allies in restoration ecology (Trevelline et al. 2019), for instance, AMF inoculation in restored
86 plots yields a gain in biomass averaging 1.7 times that of non-inoculated plots (Neuenkamp et
87 al. 2018).

88 Soil inocula from beneath invaded and native plant communities have both been
89 demonstrated to be biomass enhancers compared to sterilized soils (Phillips et al. 2020).
90 Invasive species usually tend to develop more positive feedbacks compared to native species
91 (Klironomos 2002), perhaps explaining in-part their success in biological invasions. Many
92 factors influence the interplay between plants and microbes. For instance, late successional
93 species are more likely to be positively affected by feedbacks with soil microbes, whereas
94 early successional plants and annuals tend to be negatively affected (Middleton & Bever
95 2012). This happens most likely due to a trade-off in resource investment between defenses
96 and growth (Lind et al. 2013), making these early-successional species more susceptible to
97 pathogens, (Kardol et al. 2006) which accumulate over-time.

98 It is possible to develop strategies that encompass plant-microbial interactions to
99 reach the desired final ecosystems of interest, specially the non-forested ones. These
100 techniques could improve current common Cerrado grassland restoration techniques, like
101 direct seeding (Coutinho et al. 2018). Within existing direct seeding initiatives in the Cerrado
102 *Diectomis fastigiata* (Sw.) P.Beauv, a cryptogenic pantropical annual species (PWO, 2020;
103 CAD Welker, Universidade Federal de Uberlândia, Uberlândia, MG, personal
104 communication), is commonly used and dominates the plant cover in the beginning of
105 grassland restoration (Coutinho et al. 2019). Restoration plots are also jeopardized by the
106 invasion of African grasses, noteworthy *Urochloa eminii* (Mez) Davidse (Pivello et al. 1999;
107 Coutinho et al. 2019). Currently no studies have investigated the role which soil inoculation

108 and legacy effects could have on promoting the successful growth of native over non-native
109 grass species.

110 Given the importance of understanding plant-microbial interactions, we
111 investigated the effects of native and invasive soil inocula and of sterilized soil, upon *U.*
112 *emini* and *D. fastigiata* analyzing functional traits and using aboveground biomass as a proxy
113 for performance. We also investigated both species legacy effects in order to understand the
114 different outcomes the species can develop in terms of soil conditioning, and if their presence
115 can be detrimental or beneficial for restoration efforts. We hypothesized that the native
116 inoculum would be beneficial for the species used in restoration, and that the invasive species
117 would be indifferent to inoculum origin, both species should be negatively affected by plain
118 sterilized soil due to a lack of microbial diversity in them. The annual *D. fastigiata* should be
119 negatively affected by its own-legacy due to a build-up of pathogens and annual species
120 generally presenting trade-offs between defenses and growth. The invasive *U. emini* species
121 should be favored by its own-legacy, as invasive species usually present self-positive-
122 feedback. Regarding inoculum-legacy interactions, we believe the interactions will lead to
123 multiple different effects depending on treatment combination, with an overall less
124 pronounced effect on the invasive *U. emini*.

125 **Methods**

126 *Soil sampling*

127 Soil was collected at the Estação Ecológica de Itirapina (EEI), a protected area
128 where scientific experimentation is encouraged. The area is a sandy-soil Cerrado remnant
129 located in the middle of the state of São Paulo, Brazil, in November 2018. In total 570 kg of
130 soil was collected from the upper 20 cm of a firebreak, which are areas around the park
131 devoid of vegetation, which inhibit the spread of fire. No vegetation grows on firebreaks due
132 to constant soil upturning and management. Native soil was collected at the EEI, from a
133 native grassland, containing several grass species, although no *D. fastigiata* were readily
134 observed at the collection site. The invaded soil was collected from a highly disturbed patch at
135 EEI, heavily colonized by *U. emini*.

136 The soil was put in five liter pots, lined with bidim® blanket to avoid soil
137 leaching, and finally transferred to six liter plastic bags. Pots were then transported to the

138 Instituto de Pesquisas Energéticas e Nucleares (IPEN), at the Universidade de São Paulo
139 (USP), where they were sterilized with gamma radiation, in a Cobalt-60 Irradiator, with
140 effective minimal radiation of 40 kGy according to Buchan et al. (2012). One day prior to the
141 beginning of the experiment, soil beneath native grassland and beneath a grassland invaded by
142 *Urochloa eminii* was collected in order to inoculate pots.

143 *First phase of the experiment: inoculation and soil conditioning*

144 A total of 96 pots were prepared with 6 kg of gamma-irradiated soil each, to each
145 32 pots we inoculated 120 g of native soil, 120 g of invaded soil and finally 120 g of sterilized
146 soil, accounting for soil manipulation. After inoculation half the numbers of pots with each
147 inoculum (16) were sown with *Diectomis fastigiata* seeds, a species used in direct seeding
148 Cerrado restoration efforts (Coutinho et al. 2018). The other half were sown with *Urochloa*
149 *eminii* seeds, an African invasive species widespread in Brazil (Pivello et al. 1999). Pots were
150 watered daily and thinned out to six individuals. Pot position was randomized in the
151 greenhouse every three weeks. Plants were grown at the greenhouse under natural light and
152 mean daily temperature conditions - for Campinas this is 22.1° C in January and 15.7 in July,
153 the shortest days of light in winter have less than 12 hours of light and the longest light days
154 in the summer have more than 13 hours of light. After 14 weeks, plants were collected and the
155 two tallest individuals from each pot were selected for subsequent functional attributes and
156 biomass analyses.

157 *Second phase of the experiment: soil legacies*

158 After removing all individuals from each pot, each treatment was divided in two
159 and each half was sown with either *U. eminii* or *D. fastigiata* in 8 pots for each treatment in
160 May 2019. In the second week after sowing, when seedlings emerged, a small wire was
161 placed next to the tallest individual in the pot, to mark the individual for continuous growth
162 follow-up, thinning of plants occurred as previously cited, collecting the marked individual
163 and the tallest individual left in the pot. The experimental design is shown in figure 1.

164 *Aboveground biomass and functional traits measurements*

165 Two individuals from each pot were used for analyses, thus, each pot had two
166 repeated measurements. Functional traits were measured according to the standard methods
167 described in Pérez-Harguindeguy et al. (2013).

168 *Aboveground biomass*

169 Aboveground biomass was measured collecting the aboveground part of the
170 selected individuals (leaves and rhizome) after harvesting the plants. Leaves were scanned to
171 measure leaf mass per area and then dried. Rhizomes and excess leaves were stored in a
172 paper bag and left to oven dry at 60 °C for three days, when they were weighed. Leaves and
173 rhizome weight were then added to compound total aboveground biomass.

174 *Leaf mass per area (LMA)*

175 In the first phase, all leaves of the two tallest individuals in each pot were
176 collected, and in the second phase of the experiments, LMA was obtained using four leaves
177 from each of the two individuals measured in each pot.

178 *Specific root length (SRL)*

179 Root sampling happened through gently excavating the soil beneath the same two
180 selected individuals following the roots in order to avoid breaking them, and keeping a high
181 degree of certainty regarding root origin. For scanning, roots were placed in a transparent
182 plastic tray covered with a water film and then scanned and analyzed with WinRHIZO™ Pro
183 2017a software.

184 *Plant size*

185 Plant size was measured as number of leaves, final plant height and number of
186 clones. All leaves were counted in each one of the two selected individuals; plant height,
187 which was measured just before plant harvesting, was done from the base of the plant in the
188 soil to the highest point of the plant; and number of clones, which were counted as plants
189 were thinned.

190 *Arbuscular Mycorrhizal Fungi colonization (AMF%)*

191 Root sampling for mycorrhizal colonization analysis was conducted after
192 sampling for SRL. Fine roots were collected in each pot, regardless of which individual they
193 belonged to, to a total of at least 30 cm of roots per pot whenever possible. Six samples from
194 each treatment in the first phase and four samples from each treatment in the second phase
195 were collected. Methodology for staining is available in the supplementary material.

196 ***Soil analyses***

197 Soil samples were sent to the Department of Soil of the Universidade Federal de
198 Viçosa (UFV) for analysis. Air-dried soil samples from three pots per treatment in the first
199 phase of the experiment and mixed soil samples from three pots for the second phase of the
200 experiment were collected and sent for analysis at the soil laboratory at UFV. Methodologies
201 used for soil measurements are available in the supplementary material

202 ***Statistical analyses***

203 In order to understand general plant responses to the treatments, we performed
204 Principal component analyses (PCA) jointly and separately for each. In order to compare
205 variables among treatments, we performed bootstrap analyses. All variables were compared
206 only between pairs of the same species, within the same experimental phase (conditioning or
207 legacy phase). The metric used consisted in the mean group difference, repeated 10.000 times,
208 creating a normal distribution, which can be compared to the measured values.

209 Data were analyzed using R software version 3.6.1. (R Core Team 2013).
210 Packages ggplot2 (Wickham 2016) and FactoMineR (Le et al. 2008) were used for data
211 visualization. Packages Rsampling (Prado et al. 2016), shiny (Chang et al. 2020) and
212 PerformanceAnalytics (Peterson & Carl 2020) were used for performing bootstrap analysis in
213 order to make pair-pair comparisons.

214 **Results**

215 ***Conditioning phase***

216 A PCA with all the treatments and all traits from the conditioning phase (Figure 2)
217 generated a first axis of (PC1) explaining 56.2% of the variation and PC2 accounting for
218 25.1%, together they represented 81.3% of the variation. *Diectomis fastigiata* and *U. eminii*
219 were separated along the second axis of the PCA. *Urochloa* was mostly dispersed in the
220 positive half whereas *Diectomis* dispersed mostly in the negative half. *Diectomis* tended
221 towards higher height and biomass values, and *U. eminii* tended towards higher LMA,
222 especially in the sterilized-soil treatment. *Diectomis fastigiata* in sterile soil showed the
223 highest SRL values, segregating from other *D. fastigiata* treatments, towards negative PC1

224 values, meaning a lesser investment in biomass production. The PCA shows the two species
225 allocate resources in different ways; hence, each species was hereafter treated independently.

226 We observed that *D. fastigiata* performed better when inoculated with living soil
227 than when grown on sterilized soil, presenting higher biomass and height values ($p < 0.0001$)
228 (Figure 3A; Table S1). *Diectomis fastigiata* grown in the sterilized treatment showed smaller
229 values for all the measured traits, except values, when compared to both invaded and native
230 soil inoculum treatments (Figure 3-E). Mycorrhizal colonization could not be compared to the
231 sterile treatment due to insufficient root production.

232 Regarding the growth and functional traits of *U. eminii*, LMA values were
233 different among all three treatments (Table S1), being highest for the sterile treatment
234 ($p < 0.0001$), followed by plants grown with the invaded ($p = 0.011$) and then the native soil
235 inoculum ($p = 0.017$). *Urochloa eminii* grown with native soil inoculum showed higher SRL
236 values compared to the other treatments ($p = 0.04$). Although grown in completely sterilized
237 soil, *U. eminii* had much lower mycorrhizal colonization than the other two treatments ($p <$
238 0.0001). No other differences were observed amongst *Urochloa eminii* treatments

239 Regarding soils, after the first phase, they tended to be grouped by inocula rather
240 than species, the exception being the invaded soil inoculum treatment which dispersed along
241 the second axis of the plot, with a high variation in cation exchange capacity (T) (Figure 4).
242 PC1 accounted for 64.1% of the variation whereas PC2 accounted for 19.4%, 83.5% of the
243 variation in soil could be explained by these two axes. Sterilized soils had the highest levels
244 of phosphorus and tended at the same time to have lower values of sum of bases (SB), base
245 saturation (v) and pH measured with water (pH.H₂O), opposing native soil inoculum
246 treatments, values are shown in table S4.

247 ***Conditioned phase – Legacies***

248 Survival of individuals in the legacy phase is given in table 1. A PCA regarding
249 soil chemistry explained 71.2% of the variation with axes 1 and 2 only. *Diectomis fastigiata*
250 and *U. eminii* soils segregated mostly along the first axis of the PCA, *D. fastigiata* towards
251 the negative side of PC1 and *U. eminii* towards the positive side (Figure 5). Soils where *U.*
252 *eminii* grew in the second phase were concentrated towards positive PC1 and PC2 axes, the
253 exception being *U. eminii* soil with *D. fastigiata* legacy, which sided with soils where *D.*

254 *fastigiata* had grown, towards negative PC1 values. Soil P and pH.H₂O were the variables that
255 segregated soils the most, phosphorus was negatively associated with pots where *U. eminii*
256 was grown in the second phase. All the three treatments located within extremes of the figure
257 had *D. fastigiata* grown in the second phase, the treatments were *Diectomis* legacy with
258 sterilized and invaded soil inocula, and *Urochloa* legacy with invaded soil inoculum. We can
259 see that *U. eminii* legacies scatter around lower base saturation (v) and sum of bases values.
260 Values are shown in Table S5.

261 *Diectomis fastigiata*

262 *Diectomis fastigiata* with self-legacy and native soil inoculum had lower trait
263 values compared to all other treatments for: aboveground biomass, height, number of leaves
264 and AMF% (Table S2). *Diectomis fastigiata* grown in un-inoculated sterilized soil with *U.*
265 *eminii* legacy showed the highest values of aboveground biomass, LMA and number of leaves
266 compared to all other treatments (Figures 7; S1). Height of *D. fastigiata* was also higher in
267 sterilized soil inoculum with *U. eminii* legacy than in all treatments except for *D. fastigiata*
268 legacy with invaded soil inoculum (Figure S1). The sterile inoculum with *U. eminii* legacy did
269 not differ in mycorrhizal colonization compared to the other *U. eminii*-legacy treatments
270 (Table S2).

271 Native soil inoculum was associated with lower mycorrhizal colonization (Table
272 S2) of *D. fastigiata* roots ($p < 0.0001$). *Diectomis fastigiata* legacies seem to be associated with
273 higher SRL values when compared to *Urochloa* legacies with living soil inocula (Table S2;
274 Figure 6). Growing with native soil inoculum and self-legacy, *D. fastigiata* presented the
275 smallest values of AMF%, whereas with self-legacy and invaded soil inoculum, AMF% was
276 higher than for *Urochloa* legacy with native inoculum and equal to the other *Urochloa*-
277 legacies treatments (Figure 6).

278 *Urochloa eminii*

279 In *Urochloa* most differences were seen in comparison to the sterilized soil
280 inoculum treatments. Self-legacy with sterilized soil rendered higher aboveground biomass,
281 LMA and height values compared to all pots with *D. fastigiata* legacy, and at the same time,
282 also presented lower SRL values (Table S3). *Urochloa eminii* grown with *D. fastigiata* legacy
283 in sterilized soil presented lower aboveground biomass than all other *U. eminii* legacies

284 treatments, as well as lower LMA and height than *U. eminii*-legacies with invaded and
285 sterilized soil inocula (Figures 7, S1). The *U. eminii* treatment with *D. fastigiata* legacy and
286 sterilized soil inoculum also showed similar AMF% compared to *U. eminii* legacy in
287 sterilized soil and *D. fastigiata* legacy with native soil inoculum.

288 *Urochloa eminii* with *Diectomis* legacy and invasive soil inoculum had higher
289 SRL than all *Urochloa*-legacy treatments (Figure 7), and the only difference when compared
290 to other *D. fastigiata* legacies was in mycorrhizal colonization, which was higher ($p < 0.0001$)
291 (Table S3). *Diectomis fastigiata* with both native and sterilized inocula showed lower AMF%
292 levels than all other *Urochloa* legacy treatments, including sterile *U. eminii* and *D. fastigiata*
293 legacy with invaded soil inoculum. The latter and the *U. eminii* legacies with both native and
294 invaded soil inocula (i.e. living soil inocula), responded in a similar way in respect to AMF%,
295 the three being equal among themselves, presenting higher values of AMF% than the other
296 treatments (Figure 6).

297 **Discussion**

298 Soil legacies and inoculum origin are capable of altering plants responses to the
299 environment. *Diectomis fastigiata* was benefitted by soil inoculation. Each species legacy was
300 different, and stimulated different traits. *D. fastigiata* legacies stimulated finer root production
301 while *U. eminii* legacies stimulated mycorrhizal colonization, pointing to the possibility of
302 using plant legacies as drivers for desired functional trait goals in restored ecosystems.
303 *Urochloa eminii* and *D. fastigiata* grown in sterilized soil with *U. eminii* legacy presented the
304 highest average aboveground biomass values, indicating some grasses might be benefitted by
305 lower diversity inocula. The origin of the inoculum affected the net feedback response; *D.*
306 *fastigiata* was negatively affected by its own legacy; an effect that was overcome by the
307 presence of invaded soil inoculum, with roughly the same aboveground biomass production
308 compared to those grown with *U. eminii* legacies. Overall results indicate that the presence of
309 *U. eminii* with its subsequent removal should not constitute an impediment for the
310 establishment of fast-growing grasses in ecological restoration regarding its microbial legacy
311 effects. Legacies and inocula are possible tools to be used in tropical settings aiding grassland
312 restoration techniques. However, further studies are needed to evaluate the legacy effects of
313 exotic grasses on the establishment and growth of native slow-growing grasses.

314 *Soil inoculation effects and mycorrhizal responses*

315 In the conditioning phase, we aimed to test if the origin and presence of the
316 inocula affected plants, more specifically, if native soil inoculum would benefit *D. fastigiata*,
317 a species used in ecological restoration projects in Neotropical savannas (Coutinho et al.
318 2018; Sampaio et al. 2019). *Diectomis fastigiata* was benefitted by the presence of soil
319 inocula regardless of their origin, with soil sterilization hindering its development. This is in
320 line with recent evidence provided by Phillips et al. (2020), where a native Chaparral shrub
321 was indifferent to inoculum origin, although they do not propose mechanisms for this. We
322 believe that the presence of microorganisms positively affects *D. fastigiata*, allowing it to
323 obtain nutrients more effectively, the presence of one mycorrhizal species alone (*Glomus*
324 *intraradices*) is capable of transferring organic N and stimulate plant growth (White et al.
325 2015). Soil inocula here likely had thousands of different microbial and fungal strains, which
326 likely helped plant nutrient acquisition and development.

327 Negative effects from soil sterilization were amplified in the legacy phase, leading
328 to a strong negative feedback, which induced high plant mortality. *Diectomis fastigiata*
329 probably managed to grow more in the conditioning phase because of the nutrients released
330 from the microbial biomass during the sterilization process (McLaren 1969), or nutrients
331 available in the soil. Lacking microbes to cycle and make nutrients available, nutrient
332 concentration availability of the sterilized treatment diminished with time due to plant uptake,
333 leading to nutrient depletion (Bennett & Klironomos 2019) and poor plant development. It is
334 likely that, as other plants, *D. fastigiata* needs to associate with microbes in order to obtain a
335 satisfactory mineral nutrition from recalcitrant sources (Jacoby et al. 2017).

336 *Urochloa eminii* was largely indifferent to inoculum origin, as predicted, yet
337 surprisingly, sterilized-soil self-legacy benefitted *U. eminii* plants the most, even with lower
338 AMF% compared to living inocula treatments. Mycorrhizal colonization in the sterilized
339 treatment by itself indicates that contamination occurred, as it was expected to have no
340 colonization. Contamination by mycorrhizal fungi likely originated in fine dust present in the
341 purchased seed mix, but most likely contamination also occurred for both species from
342 microbes present in the seed surface as well.. Searching the literature we found three mutually
343 synergistic possibilities to explain high biomass production for sterilized soils where *U.*
344 *eminii* grew: 1) AMF spores introduced with seeds probably had a tight association developed

345 along the years between mycorrhizal strains and *U. eminii* in the seed production site, as
346 specialization occurs as time goes by, with selection of preferred strains, leading to more
347 efficient associations (Rúa et al. 2016); 2) invasive species, usually associate with generalist
348 widespread mycorrhizal fungi (Moora et al. 2011) establishing said associations with ease;
349 and 3) Wagg et al. (2014) found that in mixed species assemblages, after manipulating
350 microbial diversity levels, grasses seemed to be benefitted by low diversity microbial
351 communities. Put together, this can explain, although not mechanistically, why both species
352 produced more biomass with sterilized *U. eminii* legacies, which contained a small fraction of
353 an inoculum. In this experiment we showed that a simpler microbiota is capable of
354 stimulating higher biomass production than a putatively more diverse inoculum, in a
355 monoculture. If proven, this relationship between lower microbe diversity and grass-biomass
356 production could have direct implications for the restoration of heavily degraded areas, such
357 as former mining sites; it would be advisable to initiate restoration of those areas with grasses.
358 Benefits provided by low diversity inoculum for grasses should be investigated possibly
359 affecting the way we understand grass physiology as well.

360 *Urochloa eminii* legacies and invaded inocula rendered high levels of root
361 mycorrhizal colonization for both species, in line with evidence that *U. eminii* is a very
362 efficient AMF propagator, effectively changing the soil microbiota; it is even used in
363 laboratories for mycorrhizal spore propagation (Banuelos et al. 2013). Recent evidence also
364 points to *U. eminii* ability to accumulate spores in invaded Cerrado areas (Leite et al. 2019);
365 hence, any living soil in which *U. eminii* has grown should harbor high amounts of
366 mycorrhizal spores, ultimately affecting future colonizers of this soil patch.

367 *Diectomis fastigiata* improved development with *U. eminii*-associated legacies
368 and soil inocula could have been caused by an effect of higher availability of mycorrhizal
369 spores, or even other microorganisms. Mycorrhizal colonization is capable of overcoming
370 negative plant-soil feedbacks, *via* plant immune system stimulation (Wang et al. 2019;
371 Cameron et al. 2013). Spore diversity can be an important factor positively influencing
372 mycorrhizal colonization (Fitzsimons & Miller 2010) and high spore loads lead to elevated
373 mycorrhizal colonization (Khakpour & Khara 2012). Ultimately, the origin of inoculated soil,
374 can affect net plant response (Ma et al. 2018), as demonstrated here, directly affecting plants
375 ability to recruit available microorganisms. We also acknowledge that although our results

376 point to a strong effect caused by mycorrhizal fungi, we cannot attest for sure observed effects
377 are definitely caused by mycorrhizal fungi and not by other organisms present in the soil.
378 Understanding the biological influences of inocula, and especially of mycorrhiza, in species
379 used for restoration is necessary to better predict outcomes and influence the system in
380 desired ways.

381 ***Legacy effects***

382 *Diectomis fastigiata* had a broader range of responses regarding treatment
383 combinations, especially with self-legacy. *Diectomis* legacies stimulated the production of
384 lower LMA and higher SRL compared to *U. eminii* legacies, this however, had little effect in
385 either species biomass production, although both traits are considered to be acquisitive. This
386 indicates the responses in terms of biomass production cannot be attributed to variation in
387 these traits alone, indicating, microorganisms probably exert more influence towards biomass
388 production than variation in morphological traits alone. Other plant size traits measured were
389 not good predictors of biomass investment, and therefore are not discussed in detail.

390 Regarding *D. fastigiata* grown with self-legacy and native inoculum, initial
391 positive effects from soil inoculation in the first phase disappeared, and a self-negative
392 feedback emerged. This effect could have been caused by a build-up of host-specific
393 pathogens (Mills & Bever 1998), or other natural enemies (Kardol et al. 2007), as is expected
394 that native inocula harbor more pathogenic organisms because native soils usually have
395 higher plant diversity. Pathogens could even be targeting soil mycorrhiza (Bennett &
396 Klironomos 2019) as we observed a considerable smaller mycorrhizal presence in this
397 treatment. Kulmatiski et al. (2008) in a meta-analysis found that there is a trade-off between
398 fast growth rates and enemy defenses for annual species. Pathogen build-up through one or
399 more plant cycles could explain the observed negative feedback for the annual *D. fastigiata*.
400 This is in accordance to initial expectations the annual *D. fastigiata* would be more impaired
401 by its own legacy, although we punctuate that impairment depends on inoculum origin.

402 *Urochloa eminii* legacies were no different from *D. fastigiata* legacies in terms of
403 biomass production for *U. eminii*; making it difficult to assume positive feedbacks were
404 happening. We also acknowledge that given that species grew in warm weather during the
405 first phase and cooler weather during the second, this likely affected overall biomass
406 production, for all plants. That is to say, the absence of a noticeable positive feedback might

407 lie in external causes; hence, we consider *U. eminii* to present neutral-to-positive feedbacks,
408 irrespective of soil legacy, with effects being strengthened by low diversity inocula. This is
409 not in accordance with previous hypothesis that the invasive *U. eminii* would benefit itself; it
410 is rather indifferent to inoculum origin.

411 ***Ecological considerations and applications in restoration***

412 Soil chemistry indicates species interactions with soils leave some sort of
413 signature in terms of soil characteristics, probably associated to resource-uptake strategies
414 (Waring et al. 2015) and interactions with microbes. There seems to be a complex interplay
415 between inoculation, legacies and species identity, but due to the circumstantial nature of the
416 evidence provided here, generalizations are complicated, and further investigations are
417 needed.

418 The combination of distinct inocula and legacies can result in different outcomes.
419 These outcomes are directly associated to each species ability to recruit and interact with soil
420 microorganisms. *Diectomis fastigiata*, as an annual species, is more sensitive to negative
421 effects of different inoculum-legacy combinations than the invader *U. eminii*, as initially
422 hypothesized, showing a wider range of variation in trait responses to different treatments.
423 *Urochloa eminii* on the other hand, had little variation in functional traits, presenting overall a
424 neutral-to-positive feedback. In the legacy phase of the experiment, we found that the species
425 has a trade-off in investments on LMA and SRL with no associated effects in biomass
426 production. This helps demonstrate that part of the reason *U. eminii* is a successful invader
427 lies in its ability to tolerate differences in soil history and changes in its functional traits,
428 presenting a satisfactory performance in a range of conditions, meaning it has a broad
429 ecological niche. Although it is a widespread invasive species, this study indicates the soil
430 changes posed by *U. eminii* do not cause any sort of deficit for *D. fastigiata*, they actually
431 benefit the species. Future studies should test if *U. eminii* presents negative effects for
432 combinations of multiple native species.

433 The negative self-feedback observed for *D. fastigiata*, indicates it would probably
434 not dominate the system for the next growth cycle. We highly recommend divesting from
435 using *D. fastigiata* in ecological restoration, owing to the uncertainty regarding the species
436 origin (Filgueiras 1988) as well as its self-negative feedback. We urgently need to invest in
437 research on perennial grass species which have more chances of presenting no negative self-

438 legacy (Koziol et al. 2018) and are more likely to be benefitted by inoculation (Middleton and
439 Bever 2010), improving restoration efforts. Our study shows the importance of considering
440 plant-soil feedbacks when planning for restoration of native systems.

441 The effect of *D. fastigiata* legacies inducing higher SRL indicates there is a
442 window of possibility to use plant legacies in restoration efforts, as suggested by Brinkman et
443 al. (2017) for fen meadows. It can be possible to target desired functional traits of a restored
444 community, planning restoration to encompass species succession or substitutions in order to
445 fully explore the potential of legacy effects and inocula in restoration. In this case for
446 example, *D. fastigiata* stimulated higher SRL, and potentially even more when inoculated
447 with native soil, stimulating water and nutrient uptake (Pérez-Harguindeguy et al. 2003),
448 these are likely to induce higher growth rates for coming species. It would be necessary
449 therefore, to define goals of a desired targeted restored ecosystem in terms of functional traits,
450 species composition and community response to foster ecosystem restoration using soil
451 legacies and inoculation. Important to note, as of yet, there are no evidences if species
452 succession occurs in the Cerrado grasslands because the vast majority of native species recruit
453 via vegetative growth (Pilon et al. 2019; Silveira et al. 2020). Thus, restoration efforts in the
454 Cerrado should not rely on the expectation that succession will occur in this ecosystem.
455 Studies regarding vegetation dynamics of Cerrado communities are deeply needed and
456 welcomed.

457 ***Future directions***

458 Our experimental design was limited to the use of two grass species only. Mixed
459 plant species assemblages should be used for both legacy and inoculation experiments,
460 especially focusing on slow-growing species, to assess how multiple interactions between
461 plants and microbes are shaped in complex settings. Future studies need to address the
462 questions if and why some grass species are benefitted by lower diversity microbial
463 communities. Finally, it is important that upcoming studies try to identify specifically which
464 groups of organisms are being affected by the plants studied, so that more precise assertions
465 can be made regarding soil legacies and plant-soil feed-backs. New studies targeting both
466 natural vegetation and restored ecosystems dynamics, encompassing legacy and microbial
467 effects are highly advised.

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645 **Tables and Figures**

646 Table 1. Survival of individuals (in percentage) used for statistical analysis according to each treatment, in the legacy phase. The species line,
 647 corresponds to which species grew in the legacy phase of the experiment, the legacy line indicates which species left a legacy in the pot,
 648 inoculum corresponds to the inoculum added to the soil in the conditioning phase: **I** corresponds to invaded-soil inoculum, **N** corresponds to
 649 native-soil inoculum and **S** corresponds to sterilized-soil inoculum. Each treatment consisted of 8 pots and two individuals per pot, to a maximum
 650 of 16 individuals per treatment.

Species	<i>Diectomis fastigiata</i>						<i>Urochloa eminii</i>					
Legacy	<i>D. fastigiata</i>			<i>U. eminii</i>			<i>U. eminii</i>			<i>D. fastigiata</i>		
Inoculum	I	N	S	I	N	S	I	N	S	I	N	S
Survival	81.25%	62.5%	18.75%	87.5%	93.75%	81.25%	100%	87.5%	87.5%	100%	68.75%	93.75%

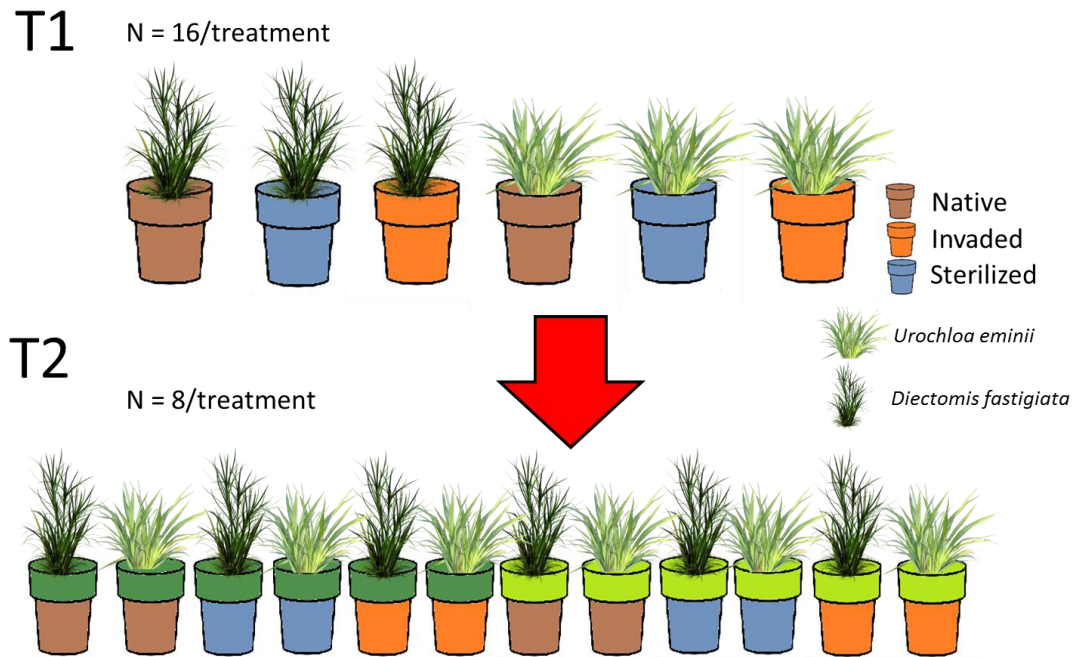


Figure 1. Schematic design of the experiment. T1 represents the first phase of the experiment and T2 the second phase. In T1 pot color represents soil inoculum origin in T1. In T2 the color in the base of the pot represents previous inoculum whereas the color in the top of the pot represents previous legacy, dark green: *D. fastigiata* legacies and light green: *U. eminii* legacies.

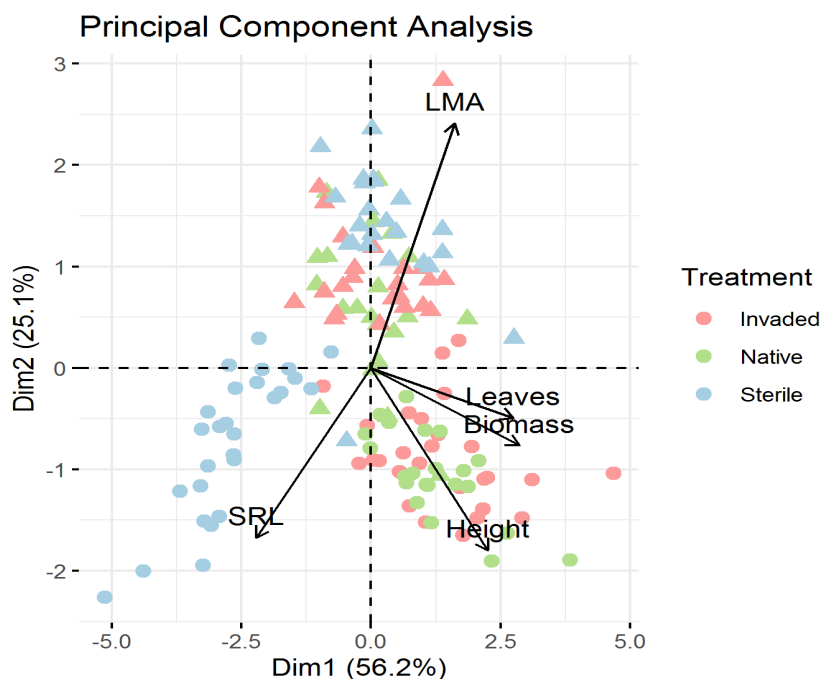


Figure 2. PCA with plant functional traits and biomass from the conditioning phase of the experiment. Triangles represent *Urochloa eminii* and circles represent *Diectomis fastigiata*. Colors indicate inoculum origin.

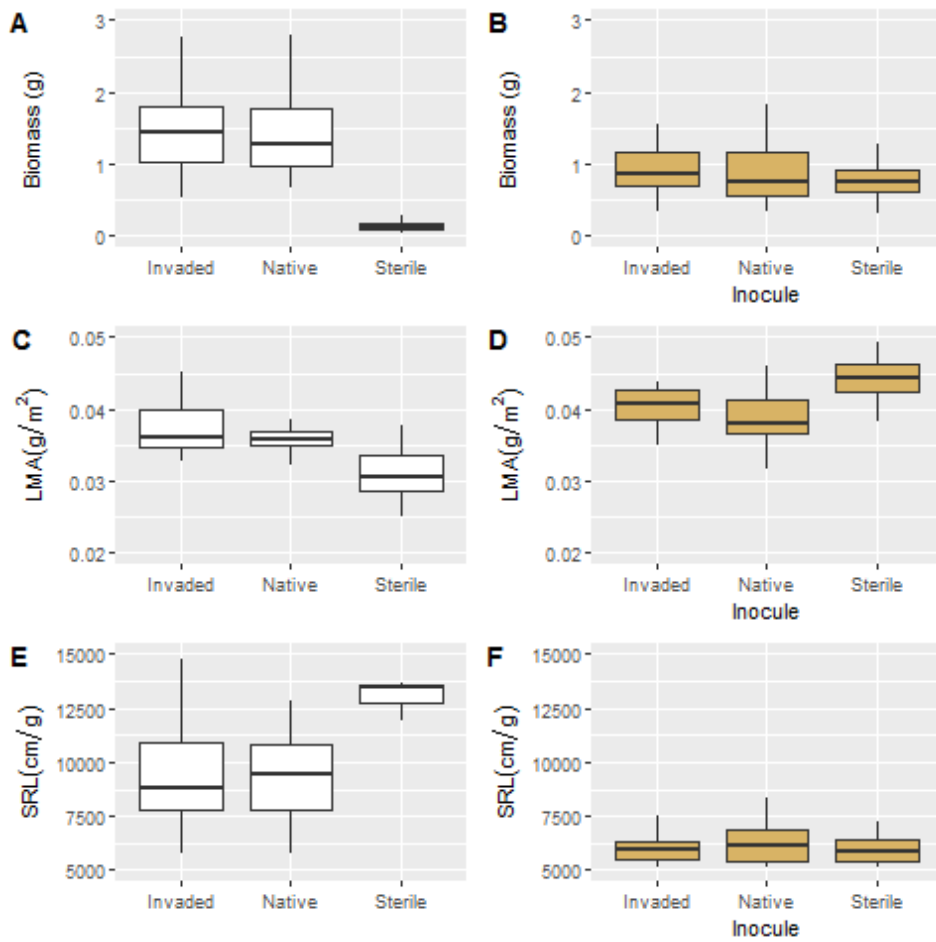


Figure 3. Aboveground biomass and functional traits of *D. fastigiata* (white; left) and *U. eminii* (orange; right) grown in the conditioning phase in sterilized soil with different soil inocula: Aboveground biomass (A, B); leaf mass per area (LMA) (C, D), specific root length (SRL) (E, F).

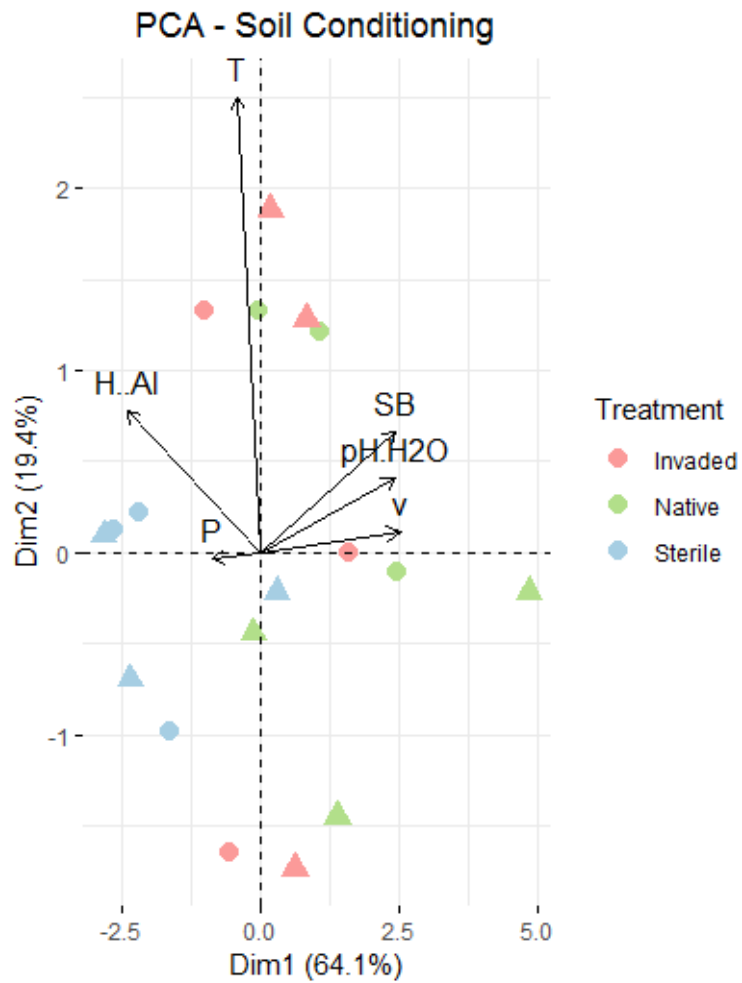


Figure 4. Principal component analysis from after the conditioning phase. Each point represents a sample, triangles represent *U. eminii* soils and circles represent *D. fastigiata* soils. Colors indicate which inoculum was used. T = cation exchange capacity, SB = sum of bases, pH.H2O = pH measured in water, v = base saturation, P = phosphorus, H.Al = potential acidity.

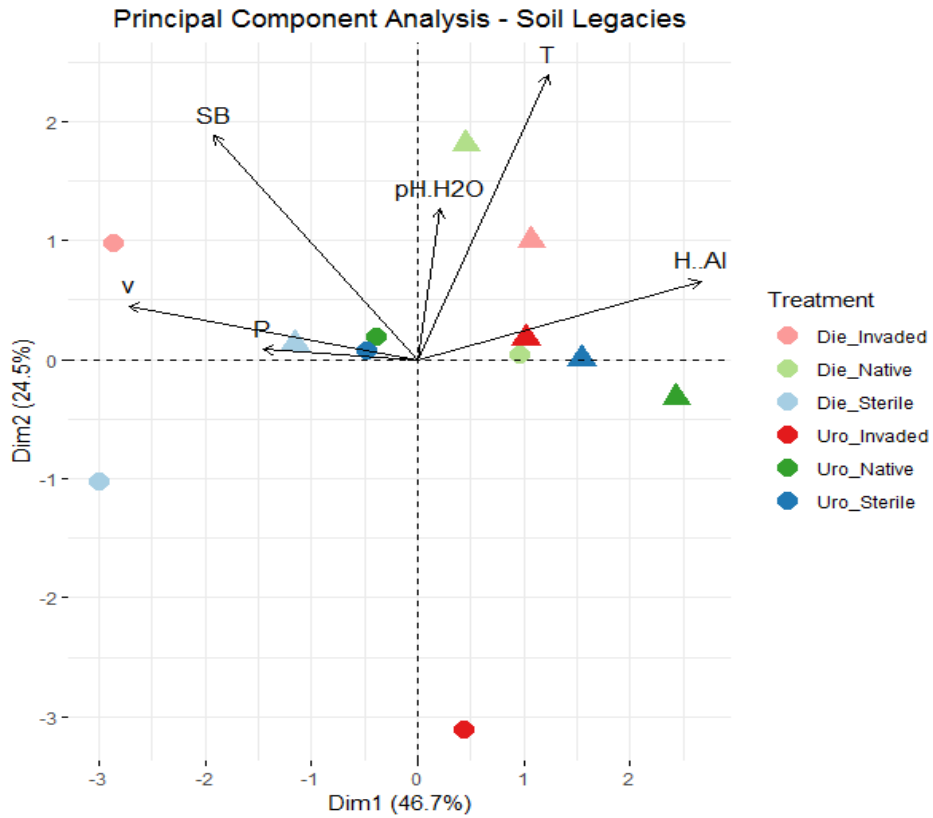


Figure 5. Principal Component Analysis of soils from after the legacy phase. Samples were compound containing soils from three pots. Point shapes represent the species grown in the legacy phase, circles represent *D. fastigiata* and triangles represent *U. eminii*, colors indicate which plant grew in the conditioning phase (soil legacy) combined with the soil inoculum, lighter colors with the prefix 'Die_' indicate *D. fastigiata* legacies whereas darker colors with the prefix 'Uro' indicate *U. eminii* legacies. pH.H2O = pH measured in water, T = cation exchange capacity, H..Al = potential acidity, P = phosphorus, v = base saturation, SB = sum of bases.

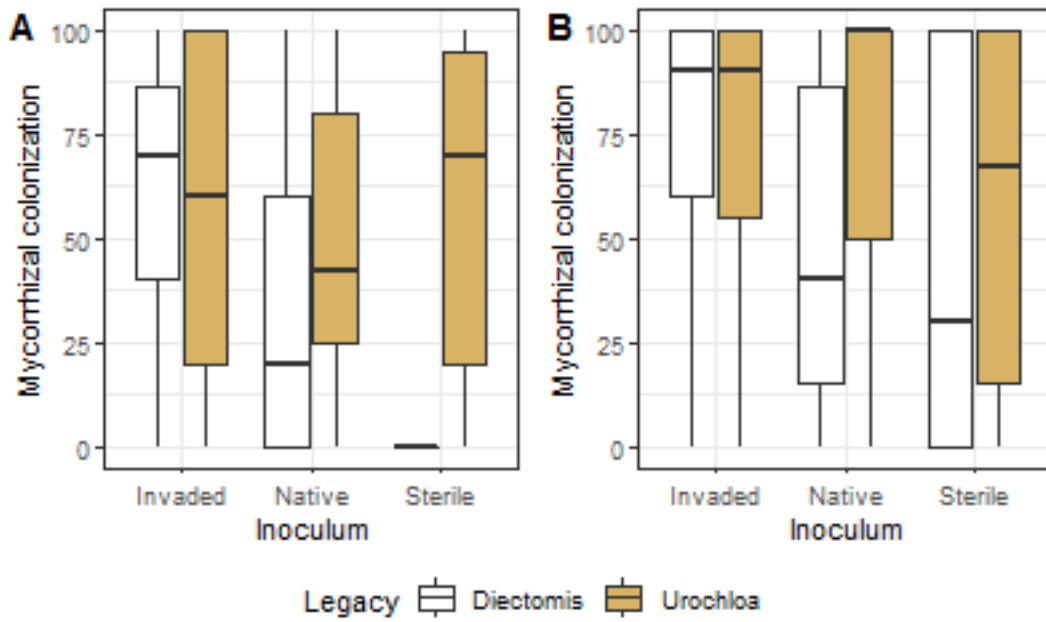


Figure 6. Mycorrhizal colonization for the legacy phase of the experiment. On the left side are presented results for *Diectomis fastigiata* and on the right for *Urochloa eminii*. White boxes represent *D. fastigiata* legacies and orange boxes *U. eminii*.

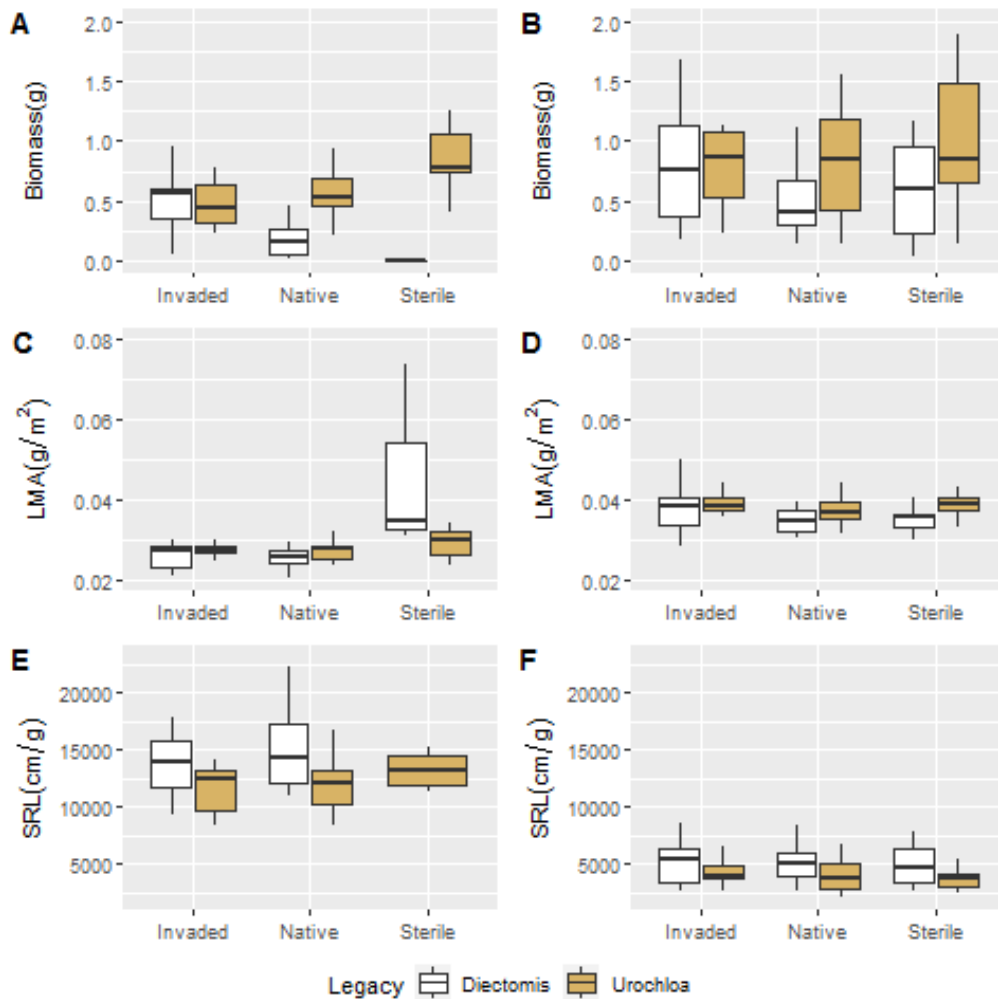


Figure 7. Aboveground biomass and functional traits for *D. Fastigiata* and *U. eminii* during the legacy

phase of the experiment. On the left side results are presented for *D. fastigiata* and on the right side for *U. eminii*. Legacies are represented by box color, white corresponding to *D. fastigiata* legacies and orange to *U. eminii*. On the x axis labels we can see the inoculum received during the conditioning phase. A and B correspond to aboveground biomass, C and D to leaf mass per area (LMA) and E and F to specific root length. *SRL in graph E does not show for *D. fastigiata* legacies because values were way out of range compared to others, starting well above 100.000 cm/g.

651