

## **Geography of non-melanoma skin cancer and ecological associations with environmental risk factors in England**

Running title: Non-melanoma skin cancer geography

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## **Abstract**

**Background** This study investigates the geography of non-melanoma skin cancer (NMSC) in England, and ecological associations with three widespread environmental hazards: radon, arsenic and UV radiation from the sun.

**Methods** Age/sex-standardised registration rates of NMSC were mapped for local authority areas (n=326), along with geographical data on bright sunshine, household radon and arsenic. Associations between NMSC and environmental variables, adjusted for socio-economic confounders, were investigated.

**Results** There was substantial geographical variation in NMSC rates across English local authorities and between cancer registration regions. 40% of variance in rates was at registry region level, and 60% at local authority level. No association was observed between environmental arsenic and NMSC rates. Rates were associated with area mean bright sunshine hours. An association with area mean radon concentration was suggested, although the strength of statistical evidence was sensitive to model specification.

**Conclusion** The significant geographical variation across England in NMSC registration rate is likely to be partly, but not wholly, explained by registry differences. Findings tentatively support suggestions that environmental radon may be a risk factor for NMSC. Although NMSC is rarely fatal, it has significant implications for individuals and health services, and further research into NMSC geography and environmental risk factors is warranted.

**Keywords:** Non-melanoma skin cancer; radon; arsenic; ecological study; England

## Introduction

Non-melanoma skin cancer (NMSC) is one of the most common cancers, although it is a much less dangerous form of skin cancer than malignant melanoma (Diepgen & Mahler, 2002; Madan *et al*, 2010). Although rarely fatal, NMSC can be a precursor to more severe conditions (Grant & Garland, 2012). Whilst NMSC is very common, geographical variation in its incidence is not well studied, partly because registration is not mandatory and it is often under-enumerated. There is therefore considerable variation between the regional cancer registries in the completeness of skin cancer registration (Goodwin *et al*, 2004; ONS, 2010); an issue that is not restricted to the UK (Curado *et al*, 2007).

Exposure to ultraviolet radiation from sunlight is a well-established environmental cause of NMSC (de Gruijl, 1999; Leiter & Garbe, 2008), with others including radon, for which there is relatively limited evidence (Advisory Group on Ionising Radiation, 2009; Henshaw & Eatough, 1995; Wheeler *et al*, 2012) and arsenic, for which evidence is stronger (Applebaum *et al*, 2007; Centeno *et al*, 2002; Guo *et al*, 2001; IARC, 2004; Karagas *et al*, 2001; Leonardi *et al*, 2012).

Arsenic is classified as a Group 1 carcinogen by the International Agency for Research on Cancer (IARC, 2012), and the suggestion that skin cancer could be caused by long-term arsenic exposure was suggested as long ago as 1888 (Perschagen, 1981). The mechanisms by which arsenic exposure leads to the development of NMSC have been demonstrated through experimentation on rodents (Burns *et al*, 2004; Waalkes *et al*, 2008). Human exposure to environmental arsenic is primarily through drinking contaminated groundwater (Smith *et al*, 1998; Tapio & Grosche, 2006). Poisoning through inorganic arsenic can also occur through long-term ingestion of food (fish,

seafood, algae and cereals) and inhalation around emissions sources such as coal-fired power stations (Pesch *et al*, 2002). Environmental arsenic exposure is widespread; populations with long-term exposure to arsenic-contaminated drinking water include those in areas of Bangladesh (Chakraborti *et al*, 2010), Taiwan (Yu *et al*, 2000), the United States (Beane Freeman *et al*, 2004), Chile (Alonso *et al*, 2010) and Argentina (Hopenhayn-Rich *et al*, 1998). Globally, the population having consumed arsenic-contaminated groundwater is estimated to be 100 to 160 million people (IARC, 2004; Martinez *et al*, 2011; Melkonian *et al*, 2011).

Although the risk of arsenic contamination of UK mains water is negligible due to stringent water quality measures (Pritchard, 2007), other exposure routes may be important. For example, food grown for consumption is in contact with soils with arsenic concentrations that vary greatly (Webb *et al*, 1978). Furthermore, environmental exposure has been indicated in studies of biological samples from people living in areas of the UK with elevated environmental arsenic, particularly ex-mining areas of south west England (Button *et al*, 2009; Kavanagh *et al*, 1998).

Radon is another naturally occurring, IARC Group 1 carcinogen, rated as such for its known effects as a risk factor for lung cancer (El Ghissassi *et al*, 2009; IARC, 2012). It has a widespread, international geographical distribution (WHO, 2007). A radioactive gas, radon is produced as part of the decay chain of uranium-238. It seeps from uranium-bearing rocks and soils and emits an alpha particle when it decays, with further alpha and beta radiation emissions from subsequent short-lived progeny (Darby *et al*, 2001). The gas disperses rapidly in the outdoors, but can accumulate inside buildings and other enclosed areas, where it can be inhaled, and can also adhere to the skin

(Eatough, 1997). Radon gas exposure is responsible for a significant proportion of human exposure to natural radiation (UNSCEAR, 2000).

Evidence for population health effects from environmental exposure to radon is much more limited than that for arsenic, and strong causal evidence is currently limited to lung cancer, primarily from occupational studies of miners (Advisory Group on Ionising Radiation, 2009). Despite limited evidence for health outcomes other than lung cancer, radiation dosimetry models have indicated a hypothetical increase in NMSC risk at UK average household radon concentrations, around 20 Becquerels per cubic metre (Bq/m<sup>3</sup>) (Eatough & Henshaw, 1991). In addition, local studies in the south west of England, where very high radon concentrations can be found, have indicated an association between radon and NMSC (Etherington *et al*, 1996; Wheeler *et al*, 2012). A comprehensive review in 2007 of the biological effects of radon concluded that the balance of evidence was against a causal relationship between radon exposure and skin cancer initiation (Charles, 2007a). However, a companion study to that review estimated the attributable risk to be around 0.7% of skin cancers at average indoor radon levels in the UK, although this was theoretically derived and subject to considerable uncertainty (Charles, 2007b).

Bringing together these issues, we investigate the geography of NMSC in England, and address the question: are non-melanoma skin cancer rates ecologically associated with three common environmental carcinogens: arsenic, radon and ultraviolet radiation from sunlight?

## **Materials and methods**

### *Geography*

The spatial units employed for this analysis were local authority (LA) areas for England, constrained by the availability of estimates of NMSC incidence. Environmental data were available at higher resolution, and ideally analyses would have been conducted using smaller spatial units to allow for more localised variation. However, there is still substantial variation in the environmental measures between LA areas, and they have the advantage of providing robust skin cancer rate estimates due to large populations. There were 326 local (county district and unitary) authorities as at April 2009, with mean population at that time estimated at 159,000 (ONS, 2009).

#### *Non-melanoma skin cancer data*

The incidence of NMSC per LA area was estimated using the registration rate produced by the eight regional cancer registries of England. These registries collect and collate data on cancer incidence and survival using a variety of sources including health care providers, cancer screening programmes and death certificates (UKACR, 2012).

Registration data are provided to the Office for National Statistics, and in turn these are distributed by the NHS Information Centre for Health and Social Care (NHSIC, 2012). Age/sex-standardised registration rates of NMSC, pooled for 2006-2008 and which had been standardised using the European Standard population were analysed. At the time of analysis these were the most recent data available, and the use of a 3-year aggregate provides a more stable, reliable rate than annual data. Whilst comparable data are collected by cancer registries for other countries of the UK, these are not all collated into a coherent dataset using common time periods; for this reason we focus here on data for England.

#### *Environmental Data*

For each of the three environmental risk factors, we used secondary data sources giving long-term estimates at sufficient geographical resolution across the whole of England.

In the absence of readily available data, the spatial distribution of environmental arsenic at LA level was estimated using the 1978 Wolfson Geochemical Atlas of England and Wales (Webb *et al*, 1978), which was only available in hard copy. The atlas includes a map of the distribution of arsenic across a grid of 2.5x2.5 km square cells, modelled from approximately 50,000 stream sediment samples. The map classifies arsenic concentrations into 10 categories, demarcated at the 10<sup>th</sup>, 20<sup>th</sup>, 40<sup>th</sup>, 60<sup>th</sup>, 80<sup>th</sup>, 90<sup>th</sup>, 95<sup>th</sup>, 99<sup>th</sup> and 99.9<sup>th</sup> percentiles of grid cell values, with the minimum category 0-4 parts per million (ppm) and the maximum  $\geq 433$ ppm. The map was scanned, georeferenced and analysed using ArcGIS 10.0 (ESRI, Redlands, CA). The resulting digital grid of arsenic concentration estimates was overlaid with LA boundaries in the GIS, and an area-weighted average of cell values within each LA calculated to produce an estimated mean arsenic concentration for each local authority. These mean concentrations were then classified back to the original atlas categories to prevent production of inappropriately precise values.

The geographical distribution of radon was obtained using an atlas produced by the Health Protection Agency (HPA) of England and Wales (Rees *et al*, 2011). This describes average household radon concentrations based on around 465,000 measurements made across England between 1980 and 2009. Data were extracted for the 326 local and unitary authorities, with the exception of a small number of areas where there was a mismatch in LA boundaries due to changes over time to unitary authority (UA) status. In these four cases (Wiltshire UA, Cheshire West and Chester UA, Cheshire East UA, Central Bedfordshire UA), best fit estimates were used, for

example with data for the old Wiltshire county boundary applied to the new Wiltshire UA. Mean radon concentrations were then classified using categories defined in previous work, for comparability (Etherington *et al*, 1996; Wheeler *et al*, 2012).

Population ultraviolet radiation exposure in each LA was estimated using data on mean daily duration of bright sunshine, based on long term estimates from the UK Met Office (Met Office, 2010), which provides the baseline of the UK Climate Projections (UKCP09). These estimates were available aggregated for 1961 to 1990, giving average daily bright sunshine hours for each month over the thirty year period, for a 5km grid across the UK. The mean value for each cell across the twelve monthly grids was calculated, and the resulting grid overlaid with the LA boundaries. A similar procedure to that used for the arsenic grid was then applied to calculate a long term, area weighted, mean daily bright sunshine hours value for each LA.

There is evidence of an inverse socio-economic gradient for NMSC, with higher rates amongst those in higher socio-economic groups (Doherty *et al*, 2010). There is also evidence of higher rates amongst those who work outdoors compared to indoors (Melkonian *et al*, 2011). To allow for potential confounding by population socio-economic status, analyses were adjusted for three domains of the 2007 Indices of Deprivation for England, employment, income and education deprivation (DCLG, 2008). These deprivation indices are produced for the c.32,000 lower-layer super output areas (LSOAs) across England. For each local authority the population-weighted mean of each deprivation domain score for its constituent LSOAs was derived. In order to estimate the prevalence of outdoor occupations, data from the 2001 UK census (ONS, 2001) were used to calculate the proportion of each LA's working population employed in primarily outdoor industries (agriculture, hunting, forestry, fishing and construction).



### *Statistical analysis*

Linear regression models were used to assess associations between age/sex standardised rates of NMSC and arsenic, radon and bright sunshine hours, with adjustment for area deprivation and outdoor occupation prevalence. All analyses were conducted using Stata version 12 (StataCorp, College Station, TX). To account for variation in under-enumeration between cancer registries, the application of random and fixed effects regression (using Stata xtreg with 'fe' and 're' options) was tested, to model and allow for variance in NMSC rates within and between registries. Primary models included arsenic and radon as categorical variables, for reasons specified above, and sunshine hours as a continuous, linear predictor. Since arsenic is believed to exacerbate the carcinogenic effects of ultraviolet radiation through inhibiting DNA repair (Danaee *et al*, 2004), possible effect modification between arsenic and bright sunshine hours was investigated using a likelihood ratio test to compare models.

## **Results**

### *Preliminary analyses*

Due to missing data on arsenic for 15 local authority areas (all in London), complete data were available for 311 areas. The total cases of NMSC registered across all 326 LAs for the three years 2006-8 was 218,475. Excluding the 15 LAs without arsenic data resulted in a reduction in cases reported of only 0.9%, to 216,497 cases. The distribution of NMSC rates across the remaining 311 areas is illustrated in the histogram in Figure 1A. Most values are approximately normally distributed around a mean of about 120 registrations per 100,000 population, but a second distribution peaks at around 20 per 100,000. Inspection of maps of the rates revealed that this range of low values comes entirely from LAs in the Thames Regional Cancer Registry in south east England. This

region includes London, explaining the small loss of registrations through exclusion of the 15 London LAs with no arsenic data. Figure 1B shows the distribution of NMSC rates across the 255 LAs remaining once Thames region LAs are excluded, resulting in a further small reduction to a total 206,454 cases, 94.5% of the cases in the original dataset. These data suggest that it is highly likely that data collection policies/practices and/or access to data in this registry are significantly different to the others, and that the very low NMSC rates observed here are an artefact. To account for this, analyses were conducted both with and without data from the Thames registry, with an assumption that analyses excluding these data are the most reliable.

### *NMSC Geography*

Age/sex standardised NMSC rates, excluding those for Thames region, are mapped in Figure 2. The mean LA standardised rate across these 255 LAs for 2006-8 was 125.9 registrations per 100,000 population per year, ranging from 37.3 to 226.5 per 100,000 per year (plus one outlier at 313.8 per 100,000). The intra-class correlation for the NMSC rate is 0.40, indicating that 40% of the variance in rates is between regional registries, and 60% at the local authority level. The map indicates that high rates are found in much of the South West, and in the Trent and North West regions. Lower rates appear particularly in parts of Eastern and West Midlands cancer registry regions.

### *Environmental risk factors*

A Hausman specification test of the full regression model, comparing random and fixed effects specifications, indicated that assumptions for random effects regression were not met, and fixed effects models were therefore applied. Results from regression models for the dataset excluding the Thames registry are presented in Table 1 and are described

here. Results for the full dataset are available in supplementary online material, along with maps of the three environmental risk factors. The number of LAs in some categories of radon and arsenic concentrations were very small, and these categories were aggregated to permit comparisons across the range of values. Intra-class correlation (ICC) coefficients indicate that for the full model including all data, 82% of variance is between regions; excluding the Thames region data reduces this substantially to 40% (as above), again supporting the exclusion of Thames data from the analysis.

Table 1 indicates a very strong association between bright sunshine hours and NMSC rates, with an increase in daily mean bright sunshine of 1 hour associated with an increase in standardised rate of 32.1 registrations per 100,000 per year (95% CI 15.9, 48.3). This association did not change substantively following adjustment for other environmental measures and confounders. The results were also suggestive of an association between NMSC and mean household radon, particularly comparing the two highest categories at concentrations above 75 Bq/m<sup>3</sup> to the reference category, although this was attenuated following adjustment. Given that the two highest radon categories only include 13 and 12 LAs, limiting statistical power, we ran the full model specifying mean radon concentration as a linear, continuous predictor. This more parsimonious model resulted in an adjusted coefficient of 0.18 registrations per 100,000 per year, per 1 Bq/m<sup>3</sup> increase (95% CI 0.04, 0.32),  $p=0.011$ . There was no clear association between estimated environmental arsenic concentration and NMSC rates, either before or after adjustment. A likelihood test for interaction between arsenic and bright sunshine hours in the full model gave a p-value of 0.25, indicating no statistical evidence of effect modification. To investigate model robustness to specification of the deprivation and

outdoor occupation variables, we ran sensitivity analyses of the full model including these measures as categorical predictors and continuous score variables. These sensitivity analyses resulted in negligible differences to the main radon and sunshine effect estimates.

## **Discussion**

This analysis demonstrates substantial variation in NMSC rates across England, and that geographical variation is unlikely to be primarily explained by differential registration, given that only 40% of the variance in rates is at the regional registry level (with 60% at local authority level). The finding of unusually low rates in the Thames region is consistent with previous analyses of registration data (ONS, 2010; South West Public Health Observatory, 2010). It is also consistent with an earlier study indicating that Basal Cell Carcinomas (the most common form of NMSC) were not recorded at this registry (Goodwin *et al*, 2004).

In this cross-sectional ecological study, geographical variation in bright sunshine hours is strongly associated with NMSC registration. Mean household radon is also associated with NMSC rates in a manner consistent with previous research in south west England (Wheeler *et al*, 2012), although the strength of statistical evidence is dependent on model specification. There is no evidence of an association with environmental arsenic.

The findings are subject to the limitations of the study design and data available. These are aggregate data, and inferring individual risk from population-level associations invokes the ecological fallacy (Morgenstern, 1982). As Savitz (2012) suggests, it would have been preferable to have individual -level data on disease and covariates, even if exposure data are ecological/geographical.

A key assumption of the study is that population exposure to the environmental risks under consideration is accurately represented by the measures used. Actual individual exposure will be depend on a variety of factors, such as behaviour (bright sunshine), diet (arsenic), dwelling characteristics (radon) and so on. The study therefore assumes that, on average, the environmental measures reflect relative levels of population exposure. The degree to which this is the case may well be different for the three different exposures. On a related issue, the LA-level environmental variables may themselves be subject to error, since they are summary measures derived from finer resolution data. In the case of arsenic, grid data were modelled from around 50,000 stream sediment samples by the atlas authors (Webb *et al*, 1978); bright sunshine hours grid data were similarly modelled from surface measurements by the Met Office (Perry & Hollis, 2005), meaning that these source data are subject to assumptions made during the spatial modelling processes. We overlaid these grids with LA boundaries and calculated area-weighted averages, introducing further potential error, given that population exposure within the LA is assumed to be uniform across its area. Radon data are simple means of all radon measurements taken within households within each LA (Rees *et al*, 2011). The mean for each LA area is therefore assumed to be representative of typical household radon concentrations within that area, again presuming population exposure across the area to be uniform.

Exposure estimates were also determined by the time periods for which data were available. Whilst health outcome data were for 2006-8, radon data were averaged from household surveys carried out between 1980 and 2009; sunshine data were long-term averages for 1961-90; arsenic data were based on surveys carried out before 1978. An assumption is therefore made that the geography of these environmental conditions is

relatively stable over time, and indicative of population exposures in the period prior to diagnosis in 2006-8. As a cross-sectional study, we do not intend to infer any latency period here. Since area arsenic and radon levels are primarily geologically determined, whilst the data pre-date NMSC data, they will still be good indicators of current geographical variation, especially at the relatively coarse spatial scale of local authorities. Bright sunshine hours data were specifically constructed by the Met Office to indicate long-term averages, and therefore should represent area chronic exposure. Whilst this may have changed to some extent recently with climate change effects (e.g. on cloud cover), these averages should again still be representative of variation in bright sunshine hours at the spatial scale employed here.

Further, data are cross-sectional, and we infer chronic exposure to environmental conditions based on residence at the time of diagnosis. Since the analysis does not account for migration, exposure misclassification is likely. For example, an individual may have lived most of their life in a low radon area, then moved to a high radon area immediately prior to NMSC diagnosis, and vice-versa. There is no reason to expect this exposure misclassification to be non-random, in which case the most likely impact on results is a dilution of effect sizes (Armstrong, 1998). Whilst we have adjusted for measures of area socio-economic status and outdoor occupations, the ecological design leads to the potential for insufficient control of confounding (Morgenstern, 2008). In particular, it is possible that residual confounding by insufficiently specified UV exposure could explain the observed association between radon and NMSC, since we only have data on geographical bright sunshine hours variation, and not sun exposure behaviour. Residual confounding associated with other exposures is also possible, for example due to ambient temperature, which has been suggested to possibly amplify the

carcinogenic effects of UV radiation (van der Leun *et al*, 2008), or the prevalence of holidaying abroad in sunny locations (Rosso *et al*, 1998). If any of these exposures are independently associated with, for example, area radon levels, then it is possible that the observed effects may in fact be due to unmeasured confounding.

Finally, a previous study of radon and skin cancer in south west England by the authors found an association only with squamous cell carcinoma of the skin, and not with basal cell carcinoma. For the present study, data were only available for all NMSC combined, potentially diluting the observed effect if it is actually primarily - or only - on squamous cell carcinoma risk.

A significant advantage of the study, in common with many other secondary data analyses, is the comprehensive geographical extent of the data and the large population considered. Since environmental risk factors often have relatively weak effects, but affect large populations, large datasets are valuable in providing appropriate scale and sufficient statistical power. Whilst the regression models could not account for uncertainty in the standardised rates, confidence intervals and observed case counts published alongside the rates indicate that they are subject to relatively small standard errors. This could be expected given that they are 3-year aggregate rates of a relatively common disease for fairly large populations. All except two of the rates (outside of Thames region) are based on more than 200 cases of NMSC. There is substantial geographical variation between local authorities for all three environmental variables, providing the opportunity to explore differences between very low and very high estimated exposures. In contrast to the ecological fallacy mentioned above, study of environmental risks at ecological levels has been suggested to negate the 'atomistic

fallacy', attempting to infer area/group-level effects through micro-scale study of individuals (Willis *et al*, 2003).

If environmental arsenic and radon are both truly risk factors for NMSC, the difference in findings is plausible, given the different routes by which humans are exposed to either element. Household radon concentration is likely to be a valid predictor of everyday, chronic radon exposure, and this is likely to form the majority of an individual's total exposure. However, as described above, arsenic exposure routes are more complex, primarily through drinking water and food, and its presence in the local environment in the UK is therefore likely to be only one small component of exposure. If the primary route of exposure in the UK is via the food chain (Pritchard, 2007), the national/international distribution of the majority of UK food would indicate that local environmental concentrations are unlikely to dominate exposure patterns.

The fact that we observe such a strong relationship between long-term area estimates of bright sunshine, indicating exposure to a known risk factor (UV radiation), and cross-sectional NMSC rates lends credibility to the analysis. However, there are significant design and data limitations to the power of the study to infer causal relationships, especially regarding the possible association with radon. Therefore, this study by no means proves an effect of radon on NMSC risk, but it does add to the body of evidence indicating that this relationship may be worthy of further investigation. The most appropriate methods may be case-control or cohort study, as suggested by Charles (2007b), or other individual-level studies with area exposure estimates (Savitz, 2012).

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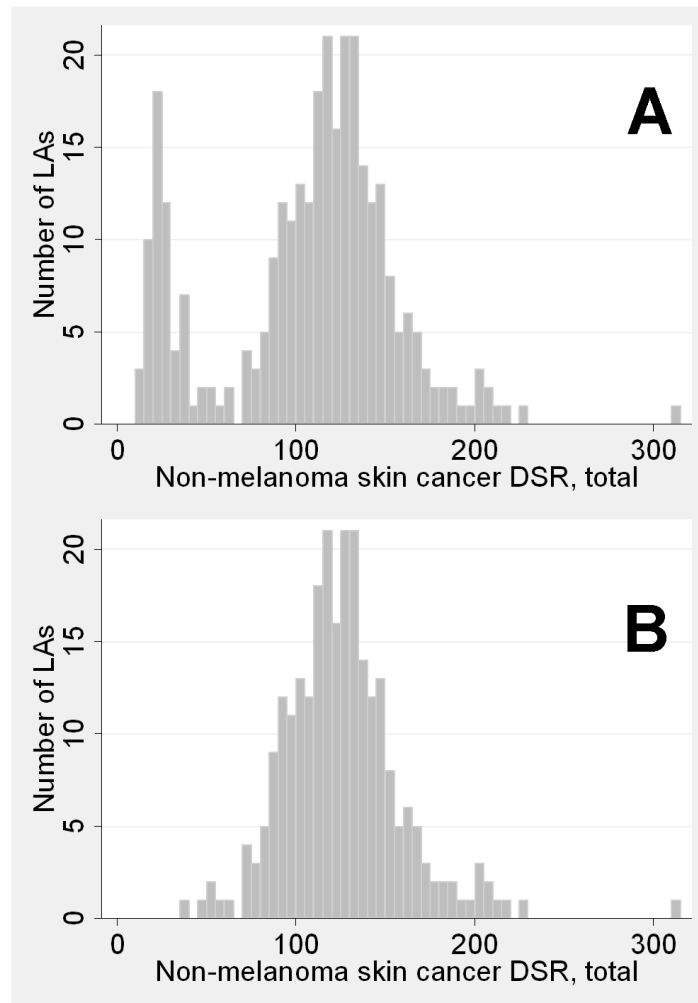
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## Figure Captions

**Figure 1.** Histograms of the distribution of non-melanoma skin cancer directly standardised registration rates (DSR) - registrations per 100,000 population per year, 2006-8 across A) All 311 study local authority areas (LAs) and B) 255 local authority areas excluding those within Thames regional cancer registry.

**Figure 2.** Non-melanoma skin cancer directly age/sex standardised registration rates, registrations per 100,000 population per year, 2006-8, English Local Authorities (LAs). Rates for Thames region excluded as described in text. Cancer registry region names are labelled.

**Figure 1.** Histograms of the distribution of non-melanoma skin cancer directly standardised registration rates (DSR) - registrations per 100,000 population per year, 2006-8 across A) All 311 study local authority areas (LAs) and B) 255 local authority areas excluding those within Thames regional cancer registry.



**Figure 2.** Non-melanoma skin cancer directly age/sex standardised registration rates, registrations per 100,000 population per year, 2006-8, English Local Authorities (LAs). Rates for Thames region excluded as described in text. Cancer registry region names are labelled.

## Non-melanoma skin cancer, England, 2006-8

Directly standardised registration rate per 100,000 per year

37.3 - 91.4

91.5 - 118.5

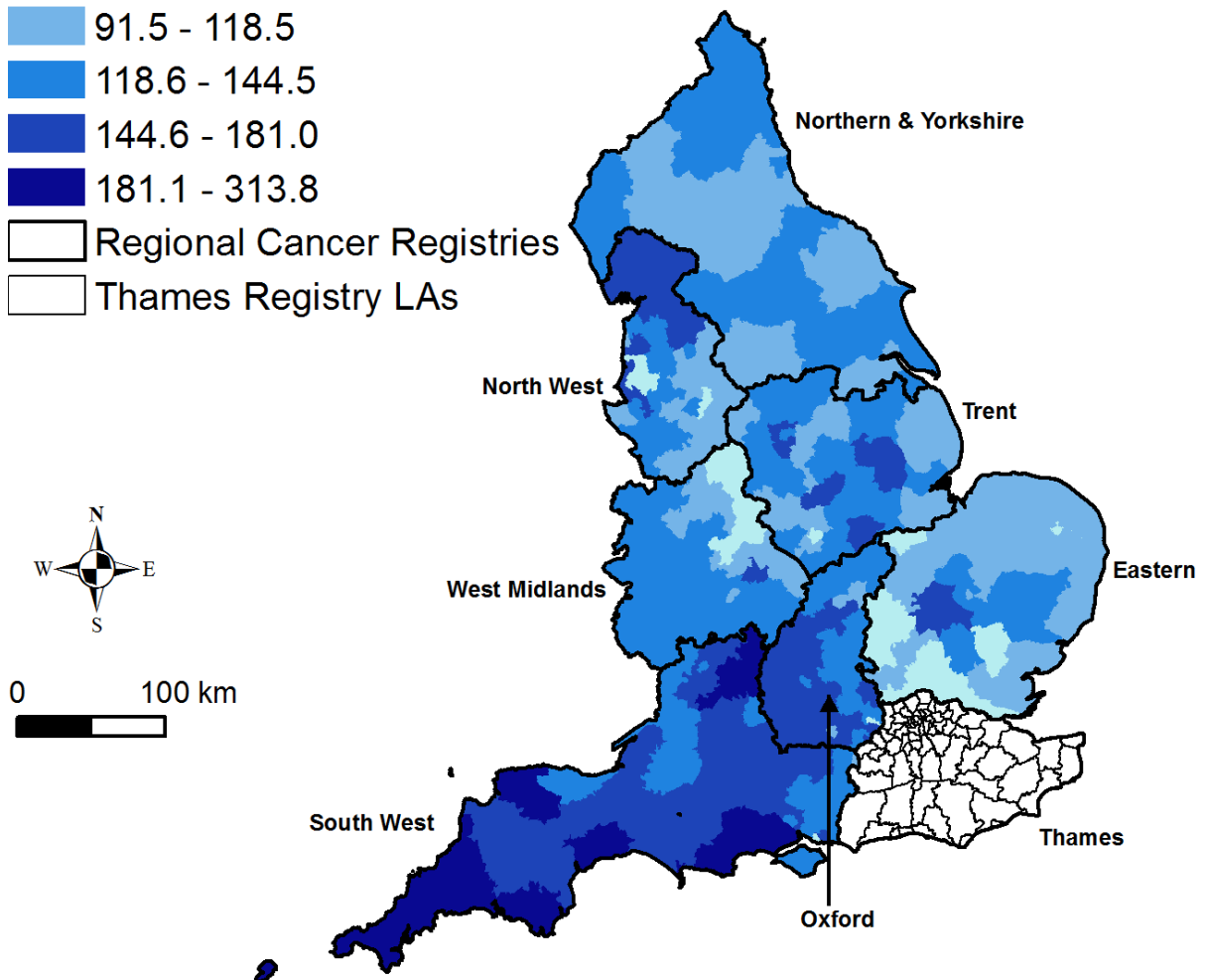
118.6 - 144.5

144.6 - 181.0

181.1 - 313.8

Regional Cancer Registries

Thames Registry LAs



**Table 1.** Random effects regression results, local authorities excluding those in Thames region.  $\beta$  are unstandardised coefficients predicting non-melanoma skin cancer directly standardised registration rate per 100,000 population per year.

| Explanatory variable                             | Value              | n   | Unadjusted univariate models <sup>a</sup> |        |                      | Fully adjusted model                 |        |         |
|--|--------------------|-----|---|--------|----------------------|--------------------------------------|--------|---------|
|  |                    |     | $\beta$                                   | 95% CI | p-value              | $\beta$                              | 95% CI | p-value |
| Mean stream sediment arsenic concentration (ppm) | 0-4                | 8   | 18.97 (0.87,37.06)                        | 0.040  | 10.73 (-5.47,26.93)  | 0.193                                |        |         |
|  | 5-7                | 42  | 0.65 (-8.79,10.09)                        | 0.893  | -4.49 (-13.21,4.22)  | 0.311                                |        |         |
|  | 8-10               | 68  | -9.67 (-17.69,-1.66)                      | 0.018  | -8.34 (-15.60,-1.08) | 0.024                                |        |         |
|  | 11-14 <sup>b</sup> | 88  | 0   |        | 0                    |                                      |        |         |
|  | 15-19              | 29  | 3.83 (-6.66,14.32)                        | 0.473  | 0.32 (-8.99,9.64)    | 0.946                                |        |         |
|  | >=20               | 20  | 1.48 (-10.67,13.63)                       | 0.811  | -5.85 (-17.90,6.19)  | 0.339                                |        |         |
| Mean household radon concentration (Bq/m3)       | 0-39 <sup>b</sup>  | 158 | 0   |        | 0                    |                                      |        |         |
|  | 40-44              | 20  | 4.93 (-7.02,16.88)                        | 0.417  | 2.08 (-8.56,12.71)   | 0.701                                |        |         |
|  | 45-49              | 16  | 6.67 (-6.33,19.67)                        | 0.312  | -1.94 (-14.02,10.13) | 0.752                                |        |         |
|  | 50-59              | 19  | -1.98 (-14.32,10.36)                      | 0.752  | 3.08 (-8.23,14.39)   | 0.592                                |        |         |
|  | 60-74              | 17  | 7.17 (-6.03,20.37)                        | 0.285  | 1.37 (-11.50,14.23)  | 0.834                                |        |         |
|  | 75-99              | 13  | 17.44 (2.93,31.94)                        | 0.018  | 9.54 (-4.08,23.16)   | 0.169                                |        |         |
|  | >=100              | 12  | 20.81 (5.73,35.89)                        | 0.007  | 12.47 (-3.90,28.84)  | 0.135                                |        |         |
| Mean daily hours bright sunshine                 | per hour           | 255 | 32.09 (15.82,48.35)                       | <0.001 | 35.46 (19.08,51.84)  | <0.001                               |        |         |
| Income deprivation score                         | per quintile       | 255 |   |        | -6.31 (-12.30,-0.31) | 0.039                                |        |         |
| Employment deprivation score                     | per quintile       | 255 |   |        | 9.85 (4.04,15.65)    | 0.001                                |        |         |
| Education deprivation score                      | per quintile       | 255 |   |        | -7.99 (-11.99,-4.00) | <0.001                               |        |         |
| % working in primarily outdoor occupations       | per quintile       | 255 |   |        | 2.37 (-0.19,4.93)    | 0.069                                |        |         |
|  |                    |     |   |        |                      | 'Within' <sup>c</sup> R <sup>2</sup> | 0.31   |         |

a. Number of local authority (LA) areas

b. Ref=Reference categories; due to low number of areas at extremes, 11-15ppm used as reference category for Arsenic

c. R<sup>2</sup> for the variance explained 'within' regions, i.e. excluding variance between regions

## ONLINE SUPPLEMENTARY MATERIAL

**Table OS1.** Fixed effects regression results, including local authorities from all regions.  $\beta$  are unstandardised coefficients predicting non-melanoma skin cancer directly standardised registration rate (registrations per 100,000 population per year).

| Explanatory variable                             | Value             | n   | Unadjusted, univariate models <sup>a</sup> |                |         | Fully adjusted model |                                      |         |
|--|-------------------|-----|--|----------------|---------|----------------------|--------------------------------------|---------|
|  |                   |     | $\beta$                                    | 95% CI         | p-value | $\beta$              | 95% CI                               | p-value |
| Mean stream sediment arsenic concentration (ppm) | 0-4               | 8   | 19.69                                      | (3.17,36.20)   | 0.020   | 12.82                | (-2.23,27.87)                        | 0.095   |
|  | 5-7               | 42  | 1.36                                       | (-7.14,9.87)   | 0.753   | -3.55                | (-11.50,4.39)                        | 0.379   |
|  | 8-10              | 75  | -8.02                                      | (-14.81,-1.23) | 0.021   | -8.05                | (-14.30,-1.79)                       | 0.012   |
|  | 11-14*            | 123 | ref  |                |         | ref                  |                                      |         |
|  | 15-19             | 38  | 3.22                                       | (-5.10,11.55)  | 0.447   | 1.93                 | (-5.57,9.42)                         | 0.613   |
|  | >=20              | 25  | 3.46                                       | (-6.40,13.32)  | 0.491   | 0.22                 | (-9.28,9.72)                         | 0.964   |
| Mean household radon concentration (Bq/m3)       | 0-39 <sup>b</sup> | 203 | ref  |                |         | ref                  |                                      |         |
|  | 40-44             | 24  | 3.72                                       | (-6.18,13.62)  | 0.461   | 0.71                 | (-8.23,9.64)                         | 0.876   |
|  | 45-49             | 18  | 5.24                                       | (-5.92,16.40)  | 0.356   | -2.62                | (-12.98,7.74)                        | 0.619   |
|  | 50-59             | 22  | -0.76                                      | (-11.15,9.64)  | 0.886   | 2.07                 | (-7.36,11.49)                        | 0.666   |
|  | 60-74             | 17  | 7.00                                       | (-5.04,19.03)  | 0.253   | 1.16                 | (-10.44,12.76)                       | 0.844   |
|  | 75-99             | 14  | 16.80                                      | (4.08,29.51)   | 0.010   | 8.82                 | (-3.00,20.64)                        | 0.143   |
|  | >=100             | 13  | 19.61                                      | (6.43,32.79)   | 0.004   | 8.62                 | (-5.00,22.25)                        | 0.214   |
| Mean daily hours bright sunshine                 | per hour          | 311 | 27.05                                      | (13.86,40.23)  | <0.001  | 27.02                | (14.27,39.76)                        | <0.001  |
| Income deprivation score                         | per quintile      | 311 |  |                |         | -6.48                | (-11.26,-1.70)                       | 0.008   |
| Employment deprivation score                     | per quintile      | 311 |  |                |         | 8.31                 | (3.38,13.23)                         | 0.001   |
| Education deprivation score                      | per quintile      | 311 |  |                |         | -6.37                | (-9.41,-3.33)                        | <0.001  |
| % working in primarily outdoor occupations       | per quintile      | 311 |  |                |         | 1.78                 | (-0.26,3.82)                         | 0.088   |
|  |                   |     |  |                |         |                      | 'Within' <sup>c</sup> R <sup>2</sup> | 0.27    |

a. Number of local authority (LA) areas

b. Ref=Reference categories; due to low number of areas at extremes, 11-15ppm used as reference category for Arsenic

c. R<sup>2</sup> for the variance explained 'within' regions, i.e. excluding variance between regions

**Figure OS1.** Map of mean household radon concentrations for English local authority areas, derived from Health Protection Agency data. Source: Rees DM, Bradley, E.J., Green, B.M.R.. HPA-CRCE-015 - Radon in Homes in England and Wales: 2010 Data Review: Health Protection Agency; 2011.

**Local authority areas (326 unitary and district councils at 2009)**

**Mean household radon, Bq/m<sup>3</sup>**

0 - 39

40 - 49

50 - 74

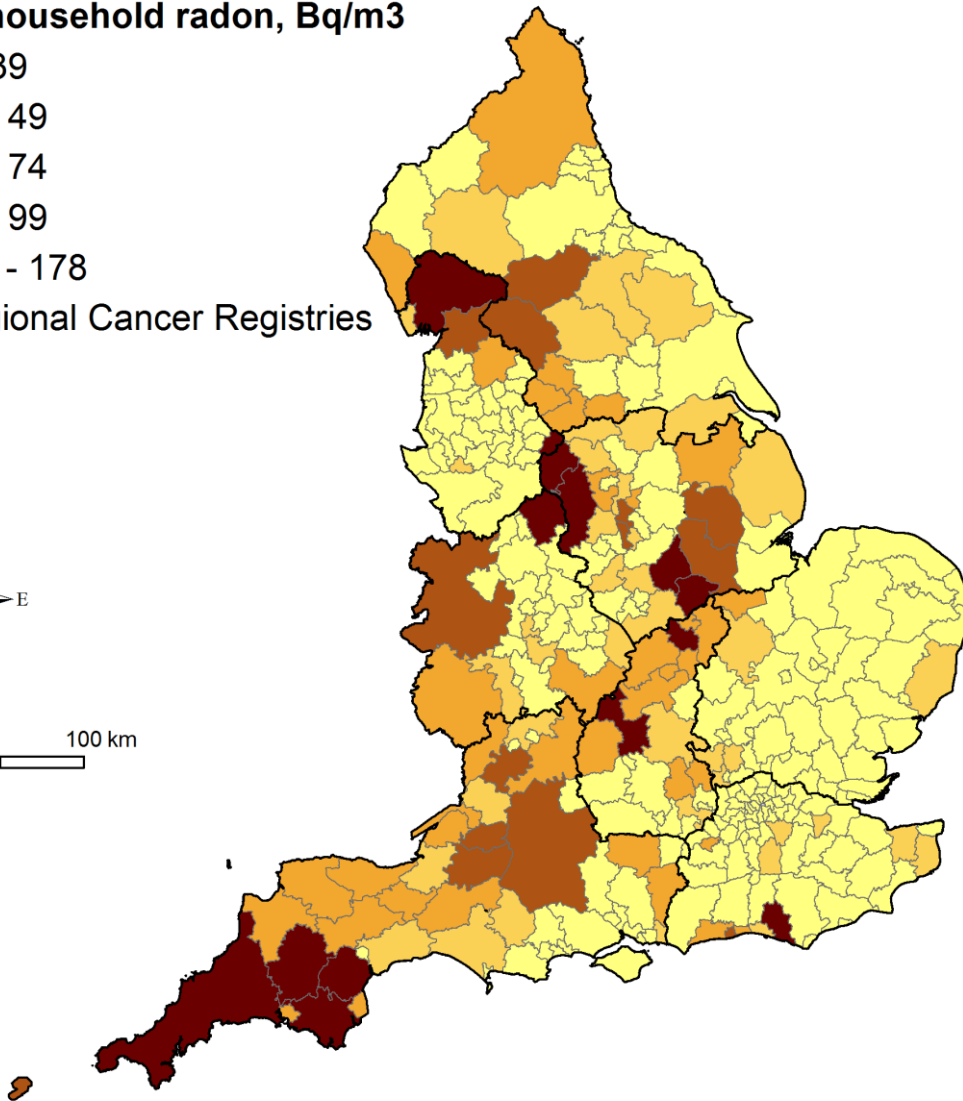
75 - 99

100 - 178

□ Regional Cancer Registries



0 100 km



Boundaries based on data provided through EDINA UKBORDERS with the support of the ESRC and JISC and use boundary material which is copyright of the Crown. Radon data sourced from Health Protection Agency.

**Figure OS2.** Map of estimated environmental arsenic concentrations (based on stream sediment samples) for English local authority areas.

**Local authority areas (326 unitary and district councils at 2009)  
Estimated environmental arsenic (ppm)**

0 - 5

6 - 8

9 - 11

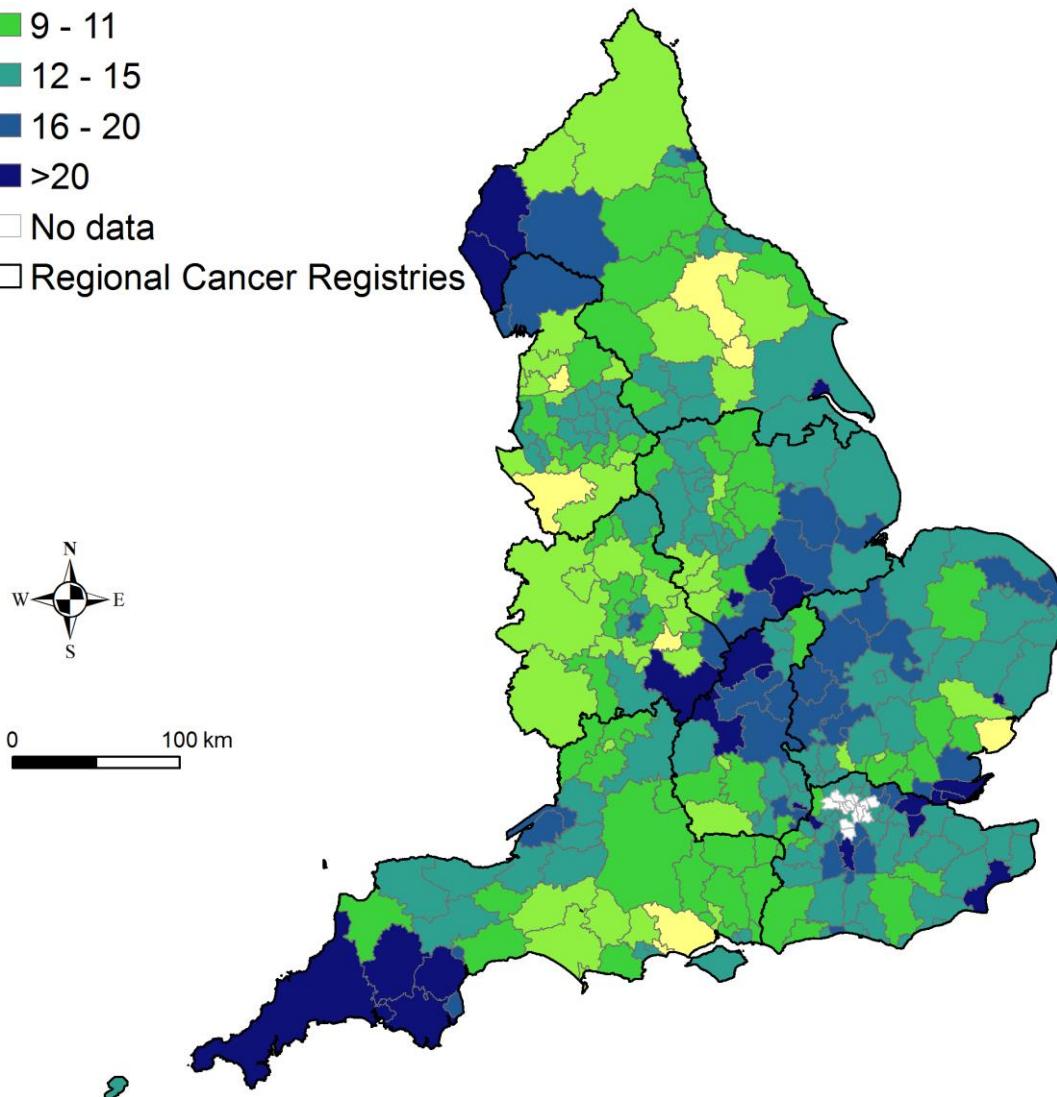
12 - 15

16 - 20

>20

No data

Regional Cancer Registries



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**Figure OS3.** Map of long-term mean daily bright sunshine hours for English local authority areas.

**Local authority areas (326 unitary and district councils at 2009)**

**Long-term mean daily bright sunshine hours (quintiles)**

3.25 - 3.66

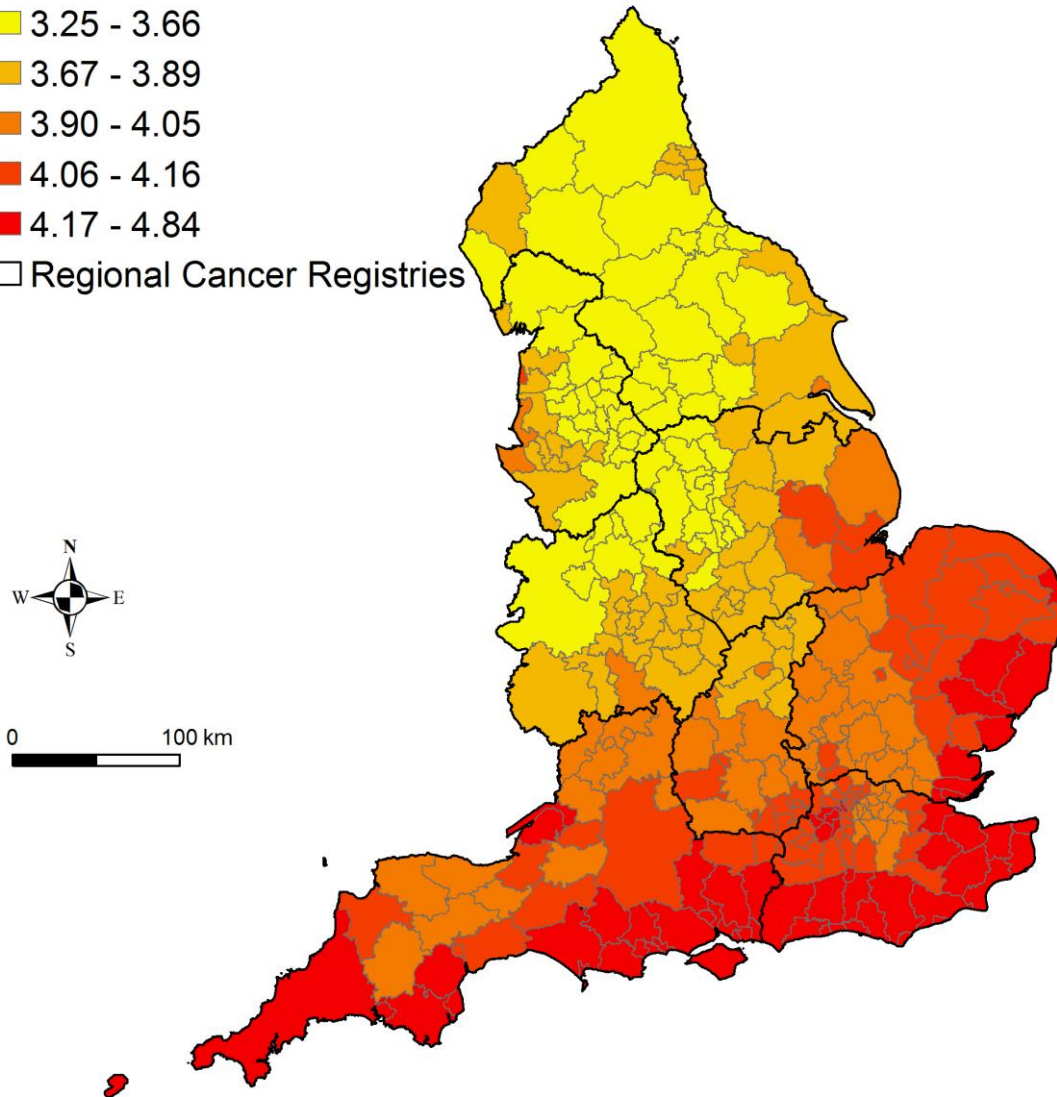
3.67 - 3.89

3.90 - 4.05

4.06 - 4.16

4.17 - 4.84

□ Regional Cancer Registries



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