

1 **Future fish distributions constrained by depth in warming seas**

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31

32

33 **European continental shelf seas have experienced intense warming over the last 30**
34 **years¹. In the North Sea, fishes have been comprehensively monitored throughout**
35 **this period and resulting data provide a unique record of changes in distribution and**
36 **abundance in response to climate change^{2,3}. We use these data to demonstrate the**
37 **remarkable power of Generalised Additive Models (GAMs), trained on data earlier in**
38 **the time-series, to reliably predict trends in distribution and abundance in later years.**
39 **Then, challenging process-based models that predict substantial and ongoing**
40 **poleward shifts of cold-water species^{4,5}, we find that GAMs coupled with climate**
41 **projections predict future distributions of demersal (bottom-dwelling) fish species**
42 **over the next 50 years will be strongly constrained by availability of habitat of suitable**
43 **depth. This will lead to pronounced changes in community structure, species**
44 **interactions and commercial fisheries, unless individual acclimation or population-**
45 **level evolutionary adaptations enable fish to tolerate warmer conditions or move to**
46 **previously uninhabitable locations.**

47

48 While the temperature of the world's oceans has gradually risen through the 20th Century,
49 the northeast Atlantic has experienced particularly intense warming, resulting in the North
50 Sea mean annual sea-surface temperature increasing by 1.3°C over the last 30 years¹, a
51 rate four times faster than the global average⁶. Predictions for the North Sea suggest a
52 further 1.8°C rise in sea-surface temperatures during the next five decades⁷ (Fig. 1). Impacts
53 of recent warming on northeast Atlantic marine ecosystems have been diverse, including
54 reorganisation of the plankton community⁸, modification to the phenology of fish
55 spawning^{9,10}, and alterations of ecosystem interactions^{11,12}. Due to its longstanding
56 economic importance to fisheries (reported landings in 2007 valued at \$1.2 billion¹) and
57 other industries, the ecology of the North Sea has been intensively monitored throughout this
58 period of recent warming.

59

60 Analyses of North Sea fish surveys have revealed northerly range expansions of warmer-
61 water species¹³, population redistributions to higher latitudes² and deeper water¹⁴, and
62 widespread changes in local abundance associated with warming, with impacts on
63 community structure³. This substantial modification to fish community composition in the
64 region has had an observable economic impact on fisheries, with landings of cold-adapted
65 species halved but landings of warm-adapted species increasing 2.5 times since the 1980s³;
66 a pattern also identified in other marine ecosystems¹⁵. With a uniquely rich fish abundance
67 time-series from the period of warming, it is possible to split these data to assess how
68 predictions made using data from earlier years match observations from later years; a
69 validation approach which has been promoted for terrestrial systems¹⁶. Existing studies have
70 used survey data to describe past changes^{2,3,13,14}, or adopted process-based climate
71 envelope models to predict future abundance without validation¹⁷. Thus there is a need to
72 compare the predictions of climate-envelope models with those from more structurally-
73 complete data-driven models that have been developed and tested using spatially and
74 temporally explicit abundance data.

75

76 The GAM approach makes no *a priori* assumptions about the nature of associations
77 between predictors and response variables¹⁸ and has been used to assess the importance
78 of different environmental drivers on patterns of distributions and relative abundance in
79 marine ecosystems¹⁹⁻²¹. Here we developed GAMs to predict changes in the distribution and
80 abundance of the 10 most abundant North Sea demersal (bottom-dwelling) fish species,
81 which accounted for 68% of commercial landings by the North Sea fishery between 1980
82 and 2010 ([www.ices.dk/marine-data/dataset-collections/Pages/Fish-catch-and-stock-](http://www.ices.dk/marine-data/dataset-collections/Pages/Fish-catch-and-stock-assessment.aspx)
83 [assessment.aspx](http://www.ices.dk/marine-data/dataset-collections/Pages/Fish-catch-and-stock-assessment.aspx)). We used a two-step approach. First, predictive models with different sets
84 of variables were compared using data earlier in the time-series to train the models and
85 predict known distributions and abundances later in the time-series. Second, models were
86 used to predict changes in species distributions over the next 50 years.

87

88 Predictors of species' abundance were identified from a wider array of potential variables
89 (annual sea-surface and near-bottom temperatures; seasonal sea-surface and near-bottom
90 temperatures; depth; salinity; fishing pressure: all of which are expected to influence fish
91 abundance and distribution ^{e.g. 2,3,14,22}). For each species we calculated from summer and
92 winter monitoring surveys the mean annual abundance per grid cell in a 10 year time-slice
93 (2000–2009, inclusive) and used these data to train GAM models based on different
94 combinations of variable sets to predict the same data. We then analysed associations
95 between the predictions and original observations. All model combinations performed well
96 with predictions against known data all exceeding correlation coefficients of 0.67 and only
97 marginal changes with the loss of each variable for each species (Supplementary Table 1).
98 Following an assessment of the performance of alternate GAMs (Supplementary Figure 1,
99 Supplementary Tables 1 and 2), a model that included temperature, depth and salinity
100 variables was applied to each species (Fig. 2a and Supplementary Table 1). The selected
101 models excluded the metric for fishing pressure since this was a relatively poor predictor
102 variable in the majority of cases (Fig. 2a and Supplementary Table 1).

103

104 To assess the most appropriate length of time-series to use for future projections, we
105 developed models to predict the abundance of species across the region in a decade using
106 annual and seasonal temperature, salinity and depth data from the periods 10, 20 and 30
107 years beforehand. There was no consistent improvement in model fit with increasing periods
108 of training data (Fig. 2b and Supplementary Fig. 2), thus we used 10-year training periods
109 for all subsequent projections. The final stage of the model development stage was to
110 assess the ability of GAMs, using an effective set of variables, to predict distributions for 10,
111 20 and 30 year periods into the future and compare with observations. Predictions closely
112 matched observations for 8 of the 10 species using both survey datasets (Fig. 2c and
113 Supplementary Fig. 2).

114

115 Following model development and testing, models trained on data from 2000–2009 were
116 used to predict future distributions, abundance and thermal occupancy of the eight species
117 for which the models were effective, based on environmental conditions forecasted with the
118 Hadley Centre *QUMP_ens_00* model (Fig. 3 and Supplementary Fig. 2). Predictions based
119 on independent winter and summer fish surveys showed congruent temperature occupancy
120 patterns, with species predicted to experience warmer conditions and maintain existing
121 distributions, rather than maintaining their preferred temperature ranges by redistributing to
122 other locations (Fig. 3).

123

124 We quantified latitudinal ranges, a commonly used estimator of distributions, which showed
125 considerable overlap between present and future conditions, with no consistent pattern
126 among species in predicted changes in distributions (Fig. 3). This indicates that poleward
127 advances of North Sea demersal fishes following preferences for colder waters are unlikely
128 to be commonplace, and highlights how process-based models that predict northward shifts
129 may underestimate dependence on non-thermal habitat. Importantly, predicted depth ranges
130 were also similar for present and future conditions (Fig. 3), implying that depth-associated
131 niches are the primary drivers and constraints of the distributions of demersal species. One
132 species predicted here to have the most marked reduction in abundance alongside a
133 proportionate increase in individuals in deeper water was dab (Fig. 3). As a shallow water
134 species predominantly found in the southern North Sea their current thermal experience is
135 expected to be exceeded through the projection period (Supplementary Fig. 3) suggesting
136 that expected climate change may force the species into less optimal habitats.

137

138 Seasonal temperatures, depth and salinity and likely co-varying habitat variables, appear to
139 be major determinants of current species distributions of commercially-important demersal
140 species in the North Sea, and were good predictors of past changes in distribution for many
141 species. Looking to the future, our results suggest that the strong associations of species
142 with specific habitats may ultimately prevent further poleward movement of species in

143 response to warming as previously predicted¹⁷. A recent study demonstrated that 1.6°C of
144 warming across the European continental shelf over the last 30 years locally favoured some
145 demersal species suited to warmer waters, but drove local declines in cold-adapted species,
146 despite long-term stability in spatial patterns of species presence-absence³. Dependence of
147 species on specific non-thermal habitat, together with spatially-contrasting local changes in
148 responses to warming³, may explain why mean latitudinal range shifts are only apparent in
149 some species², and are not detected in others despite sharing similar temperature
150 preferences. Dependence on specific non-thermal habitat has been observed in tagged
151 Atlantic cod (*Gadus morhua*), where fish occupied suboptimal thermal habitat for extended
152 periods with likely costs to metabolism and somatic growth²³. Indeed a dominant driver of
153 changes in the central distributions of cod in the North Sea appears to have been intense
154 fishing pressure over the last century rather than warming, which has depleted former
155 strongholds in the western North Sea, driving an eastward longitudinal shift in relative
156 population abundance but no apparent poleward shift²². These factors, together with
157 potential indirect effects of warming potentially not captured in our models, for example from
158 changes to prey abundance, may explain why models based on depth and temperature
159 were not effective for longer term projections for Atlantic cod and whiting (*Merlangius*
160 *merlangus*). It is necessary to evaluate the performance of alternate predictor variables for
161 data-driven models of these species.

162

163 Mean depth distributions of North Sea fishes that had preferences for cooler water increased
164 by approximately 5m during the warming of the 1980s but tended to slow or stabilise
165 thereafter¹⁴. Based on the GAM results we do not expect or predict substantial further
166 deepening for cooler water species because depth is such a strong predictor of distribution.
167 Collectively, the studies imply that capacity to remain in cooler water by changing their depth
168 distribution has been largely exhausted in the 1980s and that fish with preferences for cooler
169 water are being increasingly exposed to higher temperatures, with expected physiological,
170 life history and population consequences.

171

172 In the absence of substantial distributional shifts that would allow fish to occupy different
173 habitats and depths, North Sea populations are likely to experience 3.2°C of warming over
174 the coming century²⁴. Although such temperature increases are within observed thermal
175 limits for these species the ecological consequences are unknown, especially when warmer
176 conditions are closer to thermal preferences of other species using the same habitats.
177 Furthermore, physiological theory suggests that responses of species to projected warming
178 will eventually reach thermal thresholds. As species' optimal temperatures are reached,
179 increased metabolic costs will compromise growth with associated declines in population
180 productivity²⁵. Capacity to tolerate warming will thus depend on scope for thermal
181 acclimation²⁶ and adaptation²⁷, with the degree of connectivity between thermally-adapted
182 sub-populations across the geographic range of species influencing the rate of adaptation to
183 future warming. Unless adaptation or acclimation can track the rate of warming, it is likely
184 that stocks will be affected, both directly through individual physiological tolerances, and
185 indirectly through climate-related changes to the abundance of prey, predators, competitors
186 and pathogens.

187

188 Our study demonstrates the power of data-driven GAM models for predicting future fish
189 distributions. In contrast to process-based models that attempt to integrate discrete
190 ecological mechanisms such as dispersal and density dependence, GAMs are grounded by
191 past net responses of populations to all these processes, in addition to interspecific
192 interactions and habitat associations that are not typically considered in process-based
193 modelling, perhaps explaining the strong predictive power of our GAM approach for
194 predicting known future conditions. The results of this study suggest that we should be
195 cautious when interpreting process-based model projections of distributional shifts, and that
196 interpretations should be informed by data-driven modelling approaches, especially when
197 using predictions for policy and management planning. Our projections suggest that if
198 populations fail to adapt or acclimatise to a warmer environment, warming will change

199 fishing opportunities for currently-targeted species in the North Sea over the next century.
200 Historically, fishing pressure has substantially modified the North Sea²⁸ and ongoing
201 changes in management will play an important role in shaping future fisheries resources.
202 Species responses to temperature should be considered when planning future fisheries
203 management strategies to ensure that anticipated long-term benefits of management are
204 ecologically feasible in this period of intense warming.

205

206 **METHODS**

207 ***Fish surveys.*** We used two long-term monitoring surveys that give detailed descriptions of
208 the distribution and abundance of demersal (bottom-dwelling) fishes in the North Sea. The
209 Centre for Environment, Fisheries and Aquaculture Science UK (Cefas) time-series is a
210 summer survey (August–September) conducted since 1980 encompassing 69 1x1° latitude-
211 longitude cells. The International Council for the Exploration of the Sea (ICES) International
212 Bottom Trawl Survey (IBTS) time-series is a winter survey (January–March) conducted since
213 1980 encompassing 84 1x1° cells. (We only considered cells that had been surveyed at
214 least three times each decade.) Both surveys are conducted using otter trawling gear
215 (Granton trawl for pre-1992 Cefas surveys, otherwise Grande Ouverture Verticale (GOV)
216 trawls). Raw catch data were 4th-root transformed to reduce skew that is inherent in
217 ecological abundance data.

218

219 Our study focused on the 10 most abundant demersal species targeted by commercial
220 fisheries or taken as bycatch (Fig. 2c), which together accounted for 68% of commercial
221 landings (by weight) in the North Sea fishery from 1980–2010 ([www.ices.dk/marine-
222 data/dataset-collections/Pages/Fish-catch-and-stock-assessment.aspx](http://www.ices.dk/marine-data/dataset-collections/Pages/Fish-catch-and-stock-assessment.aspx)). For both surveys,
223 we grouped data into three 10-year time slices and one three-year time slice for the
224 analyses: 1980–1989, 1990–1999, 2000–2009 and 2010–2012. The limited 2010–2012 time
225 slice was only used for testing predictions from the GAMs. To ensure a balanced design,

226 mean values for each for each decadal time period were used. This method controls for the
227 variable numbers of survey hauls taken in each cell and ensures that longer-term responses
228 to climate change are identified rather than year on year variability. All data were 4th root
229 transformed before being subject to GAM modelling, and predictions were back transformed
230 before calculation of correlation coefficients.

231

232 **Depth.** We used mean 1x1° cell *in situ* measures of depth taken during the hauls for each
233 survey (Supplementary Fig. 4), which closely matched data from the 1x1° resolution GEBCO
234 Digital Atlas (summer survey, $r = 0.91$; winter survey, $r = 0.90$;
235 www.gebco.net/data_and_products/gebco_digital_atlas/)³.

236

237 **Temperature and salinity.** We calculated Sea-Surface Temperature (SST), Near-Bottom
238 Temperature (NBT) and salinity (Supplementary Fig. 4) for the period 1980–2012 using the
239 UK Meteorological Office Hadley Centre *QUMP_ens_00* standard model for the northwest
240 European shelf seas. Modelled temperatures closely matched data from the Hadley Centre
241 global ocean surface temperature database (*HadISST1.1*; 92 cells, Pearson's $r = 0.84$;
242 www.metoffice.gov.uk/hadobs/hadisst/). Data from the *QUMP_ens_00* model were provided
243 as monthly means for 1x1° cells, enabling mean winter (January–March), summer (July–
244 September) and mean annual values to be calculated (Fig. 1).

245

246 **Fishing pressure.** We calculated a spatially-explicit metric of fishing pressure for each 10-
247 year time-slice by combining annual multispecies fishing mortality (F) estimates for North
248 Sea demersal species (mean estimates of regional F for cod, dab, haddock, hake, lemon
249 sole, ling, long rough dab, plaice, saithe and whiting, weighted by spawning-stock biomass,
250 from ICES stock assessments; www.ices.dk/datacentre/StdGraphDB.asp)³ with mean otter
251 and beam trawling effort for each 1x1° cell based on hours of fishing²⁹ (Supplementary Fig.
252 4). This integrated metric combining temporal trends in fishing mortality and spatial

253 distribution of fishing effort enabled us to test the importance of fishing pressure as a
254 predictor of abundance.

255

256 **Identifying key predictors.** We used GAM models, coded using the *gam* package in *R*
257 (www.r-project.org), to test the performance of GAMs for predicting changes in fish species'
258 distribution and identify the importance of different variables to these predictions. The *s*
259 smooth was used with $k = 7$ for all variables to limit the degrees freedom in-line with the
260 number of data points. The Gaussian model was used. Assessment of the plots for each
261 variable using the *gam.plot* function showed that increasing the k value did not improve
262 model fit to each variable. The *gam.check* function was used to check the k index was above
263 or close to 1 with non-significant p values. Analysis of the residuals showed no obvious
264 deviations from normal distributions, while the response to fitted values relationship was
265 close to linear.

266

267 Data from 2000–2009 were used to test sets of variables as this period had the greatest
268 survey intensity. To identify variables that most strongly influenced prediction we first
269 developed a model with all variables (annual temperatures, seasonal temperatures, depth,
270 salinity and fishing), and a subsequent five models each excluding one set of variables
271 (Supplementary Table 1). Sea surface and near bottom temperatures from both the summer
272 and winter were grouped together to characterise seasonal fluctuations. This suite of
273 potentially correlated variables captured the extremes of temperatures that all species may
274 experience at different life stages, and ensured that thermal conditions with and without the
275 seasonal thermocline, annually varying ocean currents and land mass effects are all
276 included. We compared the performance of models based on i) the strength of correlation r
277 between observed and predicted data, ii) weighted AIC³⁰ using data from the AIC function in
278 *R*, and iii) using generalised cross validation (GCV, through *summary.gam* in *R*). Inclusion of
279 interaction terms between depth and seasonal temperature extremes either reduced or had
280 little influence on model performance (Supplementary Table 2 and summaries based on

281 Akaike weights in Supplementary Fig. 1).

282

283 ***Model development***

284 We developed predictive GAMs with a set of variables that were effective across all species.

285 The correlation coefficient r , AIC values and GCV values of modelled and observed data

286 were compared. Across-species inclusion of depth, seasonal temperature, annual

287 temperature, salinity and fishing effort all improved the predictions (Fig. 2a). A key finding

288 from this model development stage is that variables that are readily measured and projected

289 in climate models effectively predict species distributions. On average models that excluded

290 fishing effort were most similar to the all-variable models (Fig. 2a, Supplementary Table 1).

291 Since this metric had little predictive value, and we have no robust models of future fishing

292 effort, we excluded it when making future predictions.

293

294 ***Training period and predictive performance.*** To assess the influence of the duration of

295 training data on predictive power, GAMs trained on sets of one, two and three decades of

296 data for each species were used to predict 10 years into the future (Supplementary Fig. 2),

297 and the associations between predicted and known data compared. We also assessed the

298 performance of the model to predict further into the future within the historic records

299 available (Supplementary Fig. 2). We compared predicted with known abundance data for

300 each species for each forecasting period (0–30 years).

301

302 ***Forecasting future distributions.*** We used surface and near bottom annual and seasonal

303 temperature projections from the *QUMP_ens_00* model, surface and near bottom salinity,

304 and average depths from surveys between 1980–2012 as the environmental variables for

305 our predictions. We predicted fish abundances for sequential decades from 2000–2009 to

306 2050–2059 (Supplementary Figs. 5 & 6) using environmental variables (Supplementary

307 Figs. 4 & 7), and observed fish abundances from 2000–2009. Throughout the projection

308 period many cells do not experience temperatures outside of the range used to train the

309 model (Supplementary Fig. 3). For the widespread species in this study it is therefore likely
310 that at least parts of the population have experienced future conditions. However we
311 recognise that in future projected conditions the climate in some areas of the North Sea will
312 depart from existing variability in the model training period. Since it is not possible to test the
313 model beyond current thermal conditions using know data, some caution should be taken
314 interpreting projections for cells as they begin to experience temperatures beyond those
315 currently in the region (Supplementary Fig. 3).

316

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391

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404

405 **AUTHOR CONTRIBUTIONS**

406 M.J.G. and M.P.J. conceived the research; S.J., J.L.B. and D.W.S. contributed to project
407 development; S.D.S. and S.J. pre-processed fisheries agency data; L.A.R. and J.T. pre-
408 processed climate data; S.D.S., M.J.G., L.A.R., M.P.J. and S.J. designed the analysis; L.A.R
409 and S.D.S. conducted the analysis; S.D.S., L.A.R and M.J.G. prepared the initial manuscript
410 and all authors contributed to revisions.

411

412 **AUTHOR INFORMATION**

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416

417 **FIGURE LEGENDS**

418

419 **Figure 1.** Physical environment of the North Sea. (a) Bathymetry with an overlay showing
420 locations of the 84 1x1° latitude-longitude cells in which fish abundance, distribution and sea
421 temperature were reported and predicted; (b) mean Sea-Surface Temperature (SST, red)
422 and Near-Bottom Temperature (NBT, black) in the study cells from 1980–2060 in summer
423 (July–September, solid line) and winter (January–March, dashed line) from the
424 *QUMP_ens_00* northwest European shelf seas climate model. Mean decadal values (as
425 used in the model) are overlaid in the corresponding colours for SST and NBT for each
426 season.

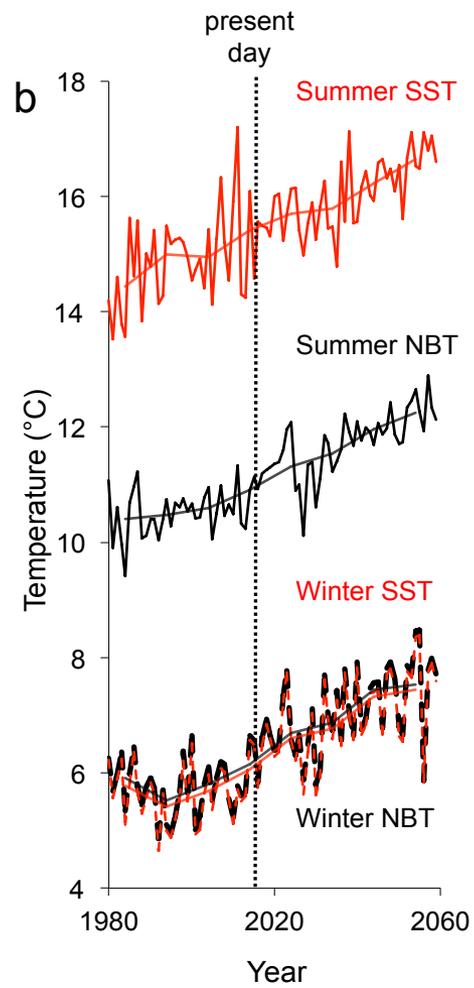
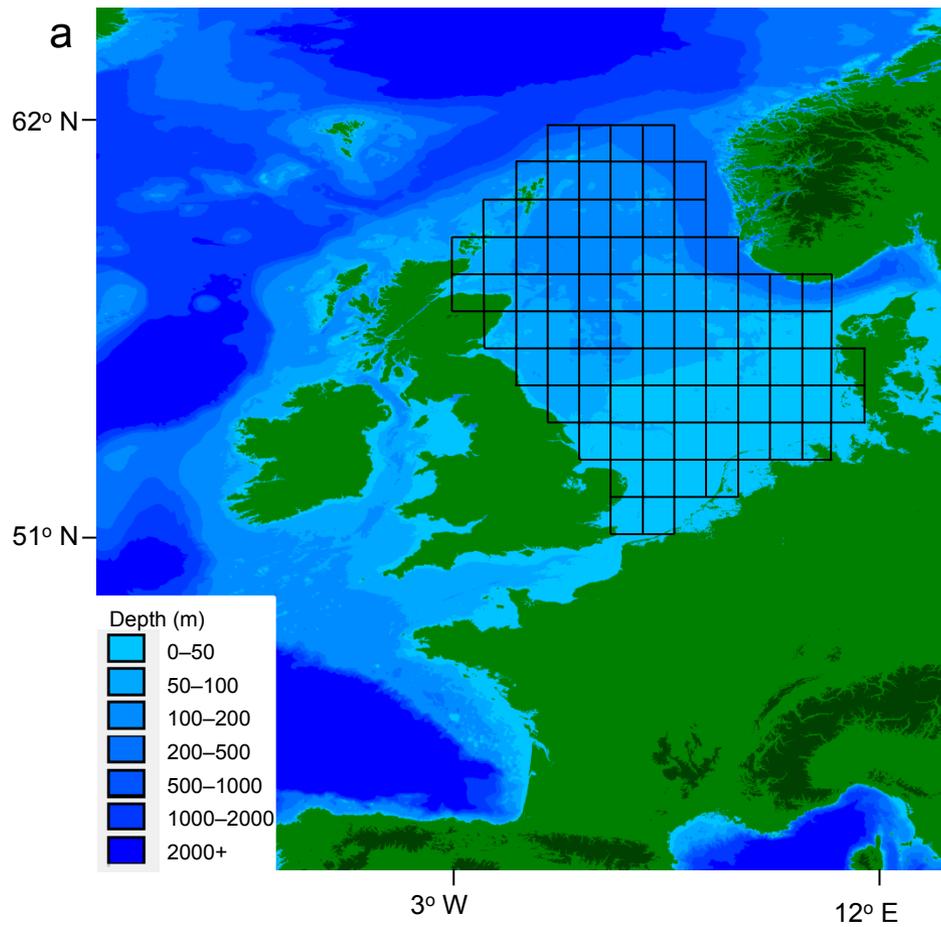
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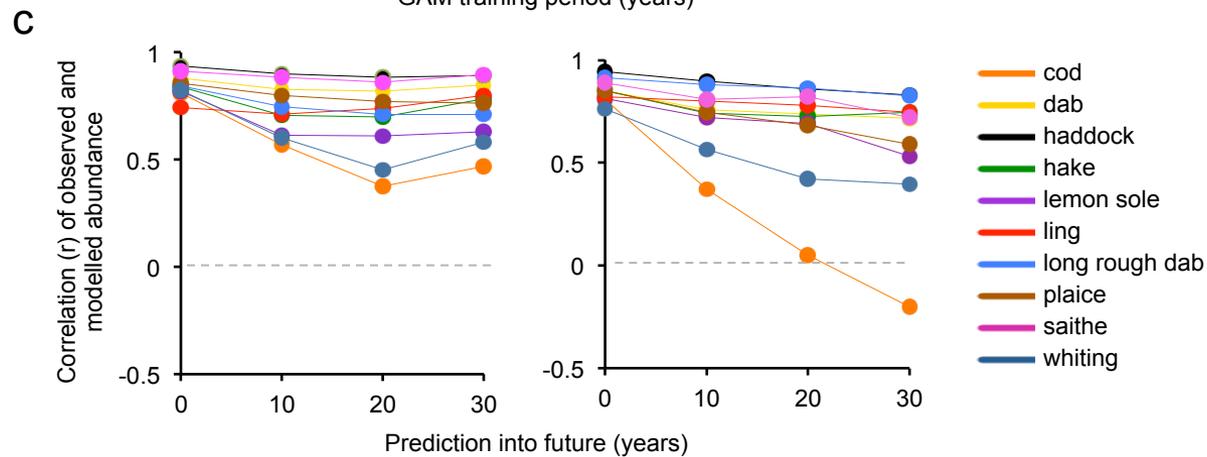
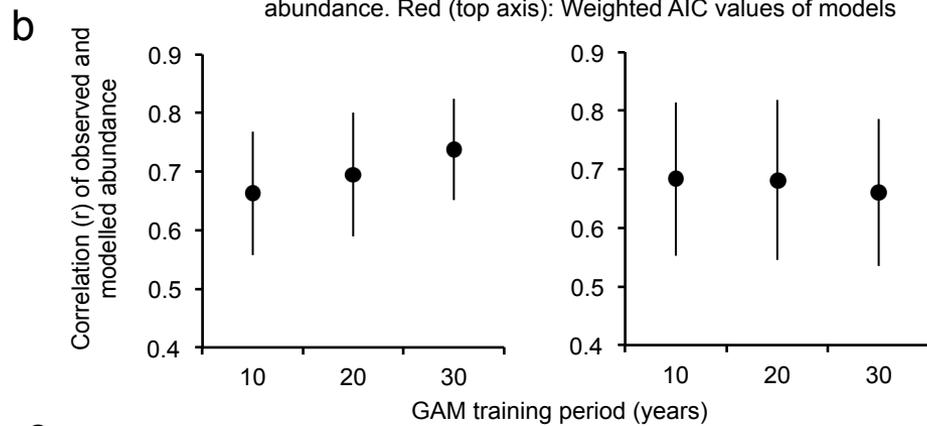
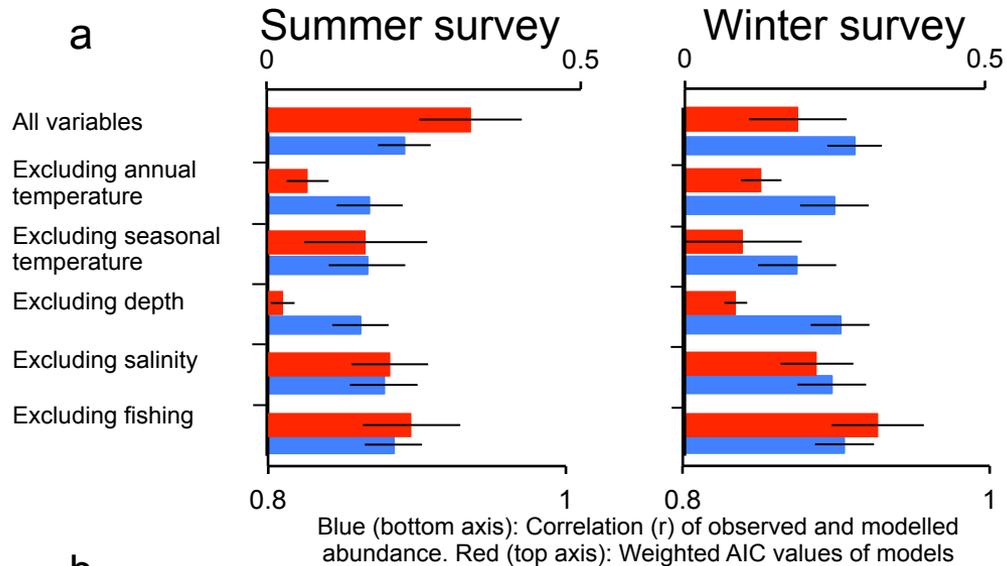
428 **Figure 2.** Predictive ability of Generalised Additive Models (GAMs). (a) Fits of predicted to
429 observed species abundance using 2000–2009 data. Variables were sequentially removed.
430 Model fits were evaluated using correlation (mean \pm SE Pearson's *r* coefficient across
431 species) and weighted Akaike Information Criterion (AIC: mean \pm SE across species). (b)
432 Duration of training data and predictive performance of GAMs using depth and seasonal
433 temperatures. Correlations (mean \pm SE Pearson's *r* coefficient across species) indicate no
434 improvement in performance with longer time-series. (c) Relationship between known data
435 and GAM predictions using depth, salinity and seasonal and annual temperature, for
436 decades beyond GAM training period.

437

438 **Figure 3.** Observed and predicted abundances of eight focal species along depth, latitude
439 and mean annual Near-Bottom Temperature (NBT) and Sea-Surface Temperature (SST)
440 gradients. Analyses were based on both the summer and winter survey datasets. Weighted
441 means are shown for each time period using arrows of corresponding colours along the x
442 axis.

443





Summer survey

Winter survey

