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**Subfossil peatland trees as proxies for Holocene palaeohydrology and palaeoclimate**

Johannes Edvardsson<sup>1</sup>, Markus Stoffel<sup>1,2,3\*</sup>, Christophe Corona<sup>1,4</sup>, Luca Bragazza<sup>5,6,7</sup>, Hanns Hubert Leuschner<sup>8</sup>, Dan J. Charman<sup>9</sup>, Samuli Helama<sup>10</sup>

<sup>1</sup> Dendrolab.ch, Institute of Geological Sciences, University of Bern, Baltzerstrasse 1+3, 3012, Bern, Switzerland

<sup>2</sup> Climatic Change and Climate Impacts, Institute for Environmental Sciences, University of Geneva, 66 Bvd Carl Vogt, 1205 Geneva, Switzerland

<sup>3</sup> Department of Earth Sciences, University of Geneva, rue des Maraîchers 13, CH-1205 Geneva, Switzerland

<sup>4</sup> GEOLAB, UMR6042 CNRS, 63057 Clermont-Ferrand, France

<sup>5</sup> Department of Life Science and Biotechnologies, University of Ferrara, Corso Ercole I d'Este 32, I-44121 Ferrara, Italy

<sup>6</sup> Swiss Federal Institute for Forest, Snow and Landscape Research, WSL Site Lausanne, Station 2, CH-1015 Lausanne, Switzerland

<sup>7</sup> Ecole Polytechnique Fédérale de Lausanne EPFL, School of Architecture, Civil and Environmental Engineering ENAC, Laboratory of Ecological Systems ECOS, Station 2, CH-1015 Lausanne, Switzerland

<sup>8</sup> University of Göttingen, Albrecht-von-Haller-Institute, Department of Palynology and Climate Dynamics, Untere Karspüle 2, 37073 Göttingen, Germany

<sup>9</sup> Geography, College of Life and Environmental Sciences, University of Exeter, EX4 4RJ, UK

<sup>10</sup> Natural Resources Institute Finland, P.O. Box 16, 96301 Rovaniemi, Finland

\*Corresponding author: markus.stoffel@dendrolab.ch

1 **Abstract:**

2 Due to the scarcity of reliable and highly resolved moisture proxies covering much of the  
3 Holocene, there has been increased interest in the study of living and subfossil peatland trees  
4 sensitive to gradual and extreme changes in hydrology, precipitation, and related environmental  
5 processes. Peatland development and the associated carbon accumulation, which are strongly  
6 influenced by hydrological fluctuations, are also of prime importance as peatlands represent long-  
7 term sinks of atmospheric carbon. Improved knowledge of peatland development and soil  
8 moisture variability during the Holocene is therefore essential to our understanding of long-term  
9 hydroclimate changes, the terrestrial carbon cycle, and to enable more robust predictions of  
10 peatland response to future climate changes.

11 Here, we review the existing mid- to late Holocene peatland tree-ring chronologies that  
12 have been used to study climate variability on (sub-)annual to centennial scales with a primary  
13 focus on northern Europe. Since the 1970s, absolutely dated tree-ring chronologies covering  
14 substantial parts of the Holocene have been developed from excavated remains of oak (*Quercus*  
15 spp.) and pine (*Pinus sylvestris* L.). The annual tree-ring patterns of these trees are often  
16 characterized by periods of depressed growth reflecting annual to decadal hydroclimatic changes.  
17 In addition, changes in the spatio-temporal distribution of trees throughout the Holocene are often  
18 found to reflect decadal to centennial climate and hydrological changes. Moreover, synchronicity  
19 between tree-ring chronologies and tree-population dynamics over larger geographical areas  
20 show periods of coherent regional climate forcing, especially during the mid-Holocene.

21 This review (i) provides an overview of pioneering and recent studies presenting tree-ring  
22 chronologies developed from subfossil peatland trees, and (ii) presents recent developments in the  
23 fields of dendroecology (i.e. the response of tree growth and changes in vitality as a result of  
24 changes in climatic variables) and dendroclimatology (i.e. the reconstruction of climate

25 fluctuations based on tree-ring analyses) in peatland regions. Moreover, we (iii) use long-term  
26 climate reconstructions based on alternative proxies for comparison, and (iv) present different  
27 ways to analyse tree-ring records to generate novel information on annual to centennial  
28 timescales. This analysis is based on an unprecedented network of tree-ring chronologies from  
29 Denmark, Finland, Germany, Great Britain, Ireland, Lithuania, the Netherlands, Poland, Sweden,  
30 and Canada, as well as a wealth of old and previously (un) published literature from Scandinavia  
31 and Germany, which has not been accessible to a wider audience in the past due to inaccessibility  
32 or linguistic barriers. Finally, a map of possible hotspots for the assessment of continuous  
33 peatland-tree studies is presented, along with suggestions for new research directions in the field.

34

35 **Keywords:** Peatland; Raised bog; Palaeohydrology; Palaeoclimatology; Dendrochronology;  
36 Subfossil trees; *Quercus robur*; *Pinus sylvestris*.

37

### 38 **Highlights**

- 39 • We review extensive peatland tree-ring chronologies from northern Europe covering  
40 major parts of the Holocene.
- 41 • Long distance cross-matches (correlograms) and significant positive correlation point to  
42 larger-scale climate signals.
- 43 • Subfossil peatland trees can be used for detailed palaeoclimatic and palaeohydrological  
44 reconstructions.
- 45 • Warm/dry conditions are favourable for peatland tree growth, but may turn peatlands from  
46 carbon sinks to carbon sources.

47

48

## 49        **1. Introduction**

50    A prerequisite for any effort to distinguish natural climate variability from anthropogenic effects  
51    is to reconstruct past climatic variations prior to industrialization (Jansen et al., 2007; Marcott et  
52    al., 2013). In recent decades, improved palaeoclimatic methodologies and chronological  
53    techniques have enabled the development of robust, accurate and detailed climate and ecosystem  
54    reconstructions, and thereby increased our knowledge of climate dynamics over the course of the  
55    Holocene (Wanner et al., 2008; Marcott et al., 2013). Several multi-proxy temperature  
56    reconstructions have been created both in the northern and southern hemispheres for the last two  
57    millennia (Moberg et al., 2005; D'Arrigo et al., 2006; Mann et al., 2008; Birks et al., 2014;  
58    Stoffel et al., 2015), many of them being derived from tree rings and with annual resolution. The  
59    vast majority of dendroclimatic studies have been used to infer past temperatures at high latitudes  
60    (Briffa et al., 1990; D'Arrigo et al., 2006; McCarroll et al., 2013; Linderholm et al., 2014). A  
61    much smaller proportion of time series can be used for the reconstruction of long-term, highly  
62    resolved precipitation series (Büntgen et al., 2014; Cook et al., 2015), thereby hampering our  
63    understanding of past moisture or hydrological variations.

64        Tree growth in natural peatlands strongly depends on water-table depth (Boggie, 1972;  
65    Linderholm et al., 2002), such that rising water tables cause reduced growth due to hypoxia or  
66    anoxia, whereas lowered water tables create relatively well-oxygenated peat-surface layers  
67    promoting tree growth (Boggie, 1972; Leuschner et al., 2002). As a consequence, trees growing  
68    on peatlands have been shown to respond to hydrological changes (Leuschner et al., 2002;  
69    Eckstein et al., 2009; Edvardsson et al., 2015a) and can therefore be used for the construction of  
70    time series of past moisture fluctuations.

71        Peatlands are not only valuable long-term climate archives (Aaby, 1976; Barber, 1981;  
72    Charman et al., 2009). With an approximate coverage of 4 million km<sup>2</sup> across Eurasia and North

73 America (Lappalainen, 1996; MacDonald et al., 2006), they also constitute an important sink for  
74 atmospheric carbon dioxide (CO<sub>2</sub>) and a considerable source of methane (CH<sub>4</sub>) (MacDonald et  
75 al., 2006; Yu et al., 2010; Limpens et al., 2011). In fact, peatlands store twice as much carbon (C)  
76 as forests (Joosten and Couwenberg, 2008). Spatio-temporal dynamics of CO<sub>2</sub> and CH<sub>4</sub> fluxes in  
77 peatlands are mainly driven by hydrological conditions, which can be expressed by water-table  
78 levels (Mitsch et al., 2013). In addition to driving the carbon flux, water-table fluctuations in  
79 peatland ecosystems also influence a wide range of environmental processes such as vegetation  
80 patterns, the fixation of nitrogen, and the chemical composition of water passing through  
81 peatlands (Limpens et al., 2011; Waddington et al., 2014). Consequently, reconstructions of past  
82 water-table variations based on tree-ring records are of prime importance not only to improve our  
83 understanding of climate variability, but also to the assessment of the functioning of peatlands in  
84 a context of changing climatic conditions and under increasing levels of greenhouse gases.

85

86 In this context, the specific aims of this paper are to:

- 87 (1) Synthesize current knowledge on relationships between tree growth, hydrology, and  
88 climate changes in peatlands from both an ecophysiological and dendroecological  
89 perspective, to evaluate and document the potential of peatland trees as a proxy for water  
90 table and/or regional hydroclimatic reconstructions;
- 91 (2) Review and map past work on peatland tree-ring chronologies across the Northern  
92 Hemisphere. We include studies from e.g. Germany and Scandinavia, written in languages  
93 other than English, to make them accessible to a wider audience;
- 94 (3) Compare the spatial and temporal distribution of peatland forests at regional to  
95 hemispheric scales, to remove site effects and to interpret results in a longer-term climatic  
96 context;

- 97 (4) Compare regional patterns of change with hydrological proxies covering the same period  
98 to test how the tree-ring records capture changes in (hydro)climatic variables;
- 99 (5) Discuss the potential of tree-ring records derived from peatland trees for palaeoclimatic  
100 and palaeohydrological reconstructions over the course of the Holocene, and to illustrate  
101 related possibilities, benefits, and limitations of subfossil peatland trees in achieving these  
102 goals;
- 103 (6) Conclude with a call for further research directions and for work in new geographic  
104 regions.

105

## 106 **2. Peatlands, peat, structure and development**

107 Peat can be defined as a natural substance that is formed by partially decomposed remains of  
108 plants and contains less than 20-30% of minerals (= ash content) (Warner, 2003). The chemical  
109 and physical properties of peat are highly variable in space and time, depending on the type of  
110 plant species contributing to its formation, the degree of decomposition, and the geochemistry of  
111 water supply. In an undisturbed peatland, peat is primarily composed of water (88-97%), dry  
112 matter (2-10%), and gas (1-7%). Peat changes in composition over time with very slow  
113 decomposition, compaction, and geochemical alteration, such that it eventually forms coal after  
114 several million years. Based on the dominant botanical composition and degree of decomposition,  
115 four main categories of peat can be described: moss, herbaceous, wood, and humified peat  
116 (Charman, 2002; Loisel et al., 2014). In humified peat, the plant remains are so decomposed that  
117 the bulk of plants are no longer identifiable.

118 A peatland can be defined as an area covered by at least 30 cm of peat (Joosten and Clark,  
119 2002; Charman, 2002). Peatlands, in a broader context, are a type of wetland characterized by  
120 organic soil. The term “mire” is sometimes used to indicate a peatland with active accumulation

121 of peat, but for our purposes the two terms are equivalent. On a global scale, three main regions  
122 of peatland development can be distinguished; northern peatlands, primarily distributed in boreal  
123 and subarctic climate in the Northern Hemisphere (Fig. 1), tropical peatlands (primarily  
124 distributed in Southeast Asia, but also in South America and Africa), and Southern peatlands  
125 (primarily distributed in Patagonia, but also in temperate Australasia) (Yu, 2011). Estimates of  
126 the global peatland area are highly variable due to the different wetland classification systems and  
127 insufficient survey data (Lappalainen, 1996; Rydin and Jeglum, 2006), but it is generally agreed  
128 that total peatland area is approximately 4 million km<sup>2</sup>, corresponding to c. 2.7% of the Earth's  
129 land surface (Lappalainen, 1996; MacDonald et al., 2006). Estimates for tropical regions are  
130 subject to much greater error than those for northern peatlands. Of the total peatland area, 1.5% is  
131 found in Africa, 28% in Asia, 0.4% in Australia and Oceania, 2.6% in Central and South  
132 America, 24% in Europe, and 44% in North America (Lappalainen, 1996).

133 Peatland initiation patterns differ for the three main regions because of different histories  
134 of ice cover recession and climate change over the glacial-interglacial transition (Fig. 2).  
135 Northern peatlands show an initiation peak around 9050-7050 BCE (11,000-9000 years before  
136 present, BP), South American peatlands around 15,050-12,550 BCE (17,000-14,500 years BP),  
137 and many tropical peatlands started to form more than 18,050 BCE (20,000 years BP; Yu et al.,  
138 2010). The average peat accumulation rate is c. 13 g C m<sup>-2</sup> yr<sup>-1</sup> in tropical peatlands and c. 19-22  
139 g C m<sup>-2</sup> yr<sup>-1</sup> in northern and southern peatlands (Yu et al., 2010; Loisel et al., 2014). A  
140 conservative estimate of overall peat C pool indicates that world peatlands store c. 610 GtC,  
141 primarily in northern peatlands (c. 89%) with a small amount in tropical (c. 8%) and southern (c.  
142 3%) peatlands (Yu et al., 2010). Overall, although covering only c. 2.7% of the Earth's land  
143 surface, peatlands store at least 25% of global soil organic C (Hiederer and Köchy, 2012).



144 On the basis of vegetation composition, hydrology, peat, and water chemistry, northern  
145 peatlands (which are the focus of this review) can be distinguished into two major types:  
146 minerotrophic peatlands (= fens) and ombrotrophic peatlands (= bogs). Minerotrophic peatlands  
147 receive their water and nutrient input from both atmospheric deposition and surrounding  
148 groundwater, whereas ombrotrophic peatlands are fed entirely from atmospheric deposition  
149 (Rydin and Jeglum, 2006). The main consequence of such different hydrological pathways is that  
150 fens are richer in dissolved minerals and have a higher water pH than bogs. These geochemical  
151 differences are reflected in the floristic composition; fens are typically dominated by vascular  
152 plants (particularly sedges) and brown mosses (Amblystegiaceae family), whereas bogs are  
153 usually dominated by peat mosses (Sphagnaceae family) and a smaller vascular plant component.  
154 The role of vascular plants, such as dwarf shrubs, has however gained increased interest recently  
155 because they seem to have competitive advantages in response to climate warming (Breeuwer et  
156 al., 2010; Buttler et al., 2015). In both fens and bogs, tree cover is often limited, but there are  
157 densely tree-covered peatlands in the boreal region (Fig. 3). Under natural conditions, the most  
158 common tree species colonizing peatlands in Eurasia are Scots pine (*Pinus sylvestris* L.),  
159 Mountain pine (*Pinus uncinata* Ramond ex DC.), Dwarf mountain pine (*Pinus mugo* Turra), and  
160 Siberian pine (*Pinus sibirica* Du Tour), whereas Mountain birch (*Betula pubescens* Ehrh.),  
161 willow (*Salix* spp), and black alder (*Alnus glutinosa* L.) often colonize fens. Other tree species on  
162 northern peatlands include black spruce (*Picea mariana* (Mill.) Britton, Sterns and Poggenb.),  
163 whereas Tamarack (*Larix laricina* (Du Roi) K.Koch) sometimes colonizes fens. In this review,  
164 Scots pine and oak (*Quercus robur* L. and *Q. petraea* L.) will be in primary focus, as these  
165 species are commonly excavated in northern peatlands and are suitable for dendrochronological  
166 studies (Table 1). Excavated trunks and plant macrofossils show that oak trees once colonized  
167 significant peatland areas in northwest Europe (Pilcher et al., 1984; Leuschner et al., 2002).

168 However, analogous peatland ecosystems do not appear to exist any longer (Sass-Klaassen and  
169 Hanraets, 2006).

170 The presence of *Sphagnum* is particularly important for peat accumulation in  
171 ombrotrophic peatlands. *Sphagnum* plants have the capacity to acidify the surrounding water as a  
172 result of an extremely efficient cation exchange capacity (van Breemen, 1995; Rydin and Jeglum,  
173 2006). This enables *Sphagnum* to absorb nutrients from very low concentrations in the water  
174 source, and produces a litter that is resistant to decay. *Sphagnum* also increases soil anoxia by  
175 storing water in empty cells (hyalocysts). Because Sphagna, like all bryophytes, cannot regulate  
176 their water balance, their development is possible whenever there is a surplus of water. In  
177 addition, because they are most competitive under acidic conditions (pH < 4-5), most *Sphagnum*  
178 species benefit from the presence of low mineral content. This combination of hydrological and  
179 geochemical conditions is prevalent on ombrotrophic peatlands (van Breemen, 1995). The  
180 physiological requirements of *Sphagnum* are one of the key determinant of the global distribution  
181 of northern ombrotrophic peatlands, which are typically located north of 45° N, in particular at  
182 50°–70° N (Yu et al., 2011), under climatic conditions and landscape morphology that promote a  
183 surplus of water and low mineral availability. However, whilst *Sphagnum* is a key taxon for  
184 northern peatland formation, the presence of *Sphagnum* is not a pre-requisite for peat  
185 accumulation and some extraordinary examples exist where peat forms from plants such as e.g.,  
186 Poaceae (Smith and Clymo, 1984). Peat can be formed by a range of other vascular plants as  
187 well, particularly sedges in fens and wood in peatlands with shrub and tree cover. Peat derived  
188 from woody plants is especially important in the tropics. Specific taxa, such as the restiad rushes  
189 in Australasia, are also important peat formers in some regions (Clarkson, 2004). Peat  
190 accumulation can be equally fast in these peat types (Loisel et al., 2014) and prevailing climate  
191 appears to be a more important driver of peat accumulation rates over millennial timescales than

192 botanical composition (Charman et al., 2013, 2015). The presence of waterlogged conditions  
193 (persistent soil anoxia), low chemical quality of peat moss litter, and low pH, all hamper the  
194 microbial decomposition of plant remains, thereby favouring the long-term accumulation of peat  
195 (a sink of atmospheric C) and the preservation of plant macro- and micro-fossils creating an  
196 archive of past environmental conditions, especially in ombrotrophic northern peatlands  
197 (Chambers and Charman, 2004; Chambers et al., 2012).

198         A complex microtopography often develops on the surface of ombrotrophic peatlands,  
199 created by a process of differential growth and decay, sometimes enhanced by physical processes  
200 such as freezing. Microhabitats of pools, hollows, lawns, and hummocks are formed in relation  
201 to a water table gradient, which in turn drives vegetation composition and the spatial distribution  
202 of plants (Bragazza and Gerdol, 1999). Pools form in the wettest areas of peatlands (Fig. 3),  
203 characterized by standing water with only free-floating peat mosses and vascular plants. Hollows,  
204 lawns, and hummocks occur along a gradient of decreasing wetness and are often characterized  
205 by a continuous and dense layer of *Sphagnum* plants, and a sparse cover of vascular plants. Along  
206 the gradient from hollows to hummocks, *Sphagnum* species become more drought-adapted and  
207 vascular plants become more shrubby and abundant. In the driest locations, *Sphagnum* is absent  
208 and may be replaced by more drought tolerant mosses. In such a typical sequence of  
209 microhabitats on ombrotrophic peatlands, the presence of trees is determined by the moisture  
210 content of the peat due to the sensitivity of tree roots to anoxic conditions (Boggie, 1972; Vitas  
211 and Erlickytė, 2007).

212         Where hydrological conditions are stable, tree distribution is limited to the hummocks and  
213 individual trees tend to be small and slow growing. Where changes in hydrological conditions  
214 occur, the expansion of trees is favoured by any reduction in soil water content, such as decreased  
215 precipitation or increased evapotranspiration (Boggie, 1972; Edvardsson et al., 2015a). Peatlands

216 with consistently dry surface conditions can be heavily forested with relatively large trees,  
217 especially in boreal environments. The dominant hydrological control on tree growth means that  
218 the direct effect of temperature is less important, whereas lowering of the water table will  
219 promote expansion of tree cover and increased tree growth rates (Smiljanić et al., 2014;  
220 Edvardsson et al., 2015a, b). This theoretical concept has been confirmed by observations of tree  
221 growth following peatland drainage (Dang and Lieffers, 1989; MacDonald and Yin, 1999;  
222 Freléchoux et al., 2000). Lowering of the water table increases soil temperature, improves soil  
223 aeration, and increases nutrient availability through organic matter mineralization (Grootjans et  
224 al., 1985; Choi et al., 2007; Weedon et al., 2012), with positive effects on tree growth.

225         Although success of tree seed germination does not differ between lawns (wet  
226 microhabitats) and hummocks (drier microhabitats), the performance of tree seedlings and the  
227 probability of maturation is greater on hummocks (Holmgren et al., 2015). Positive feedbacks  
228 have also been shown between tree recruitment and dwarf shrubs on hummocks (Holmgren et al.,  
229 2015). Once they are established, trees can cause a further drying of the peat due to enhanced  
230 transpiration (Moir et al., 2010; Limpens et al., 2014a), promoting even greater tree  
231 encroachment.

232         Peatland development over time is driven primarily by successional processes and climate  
233 dynamics (Fig. 2; Hughes, 2003). Historically, the conventional successional sequence is via  
234 terrestrialisation of aquatic systems and fen peat formation to bog peat (Walker, 1970), but it is  
235 now commonly accepted that a wide variety of successional sequences can occur in different  
236 peatlands. Paludification (i.e. the gradual spread of peatland through waterlogging of previously  
237 dry ground) is the other main process leading to peat formation particularly in many boreal  
238 peatlands and blanket bogs. The occurrence and preservation of trees is influenced by  
239 successional stages, with tree growth occurring most likely in the later stages of succession, when

240 drier surfaces are more prevalent. In paludified peatlands, tree remains typically occur at the base  
241 of the sequence, as they were growing on the drier surfaces prior to peat formation. Moreover,  
242 trees can be preserved in peatlands even if they were not growing on the surface. For example,  
243 trees can fall into the margins of ponds or fens and may therefore be preserved in peat and  
244 sediment formed during the early aquatic stages of peatland growth.

245

### 246 **3. Historical background: Pioneering to recent studies**

247 Since the late nineteenth century, palaeobotanic and climatic research has been performed in  
248 peatlands, initially with a focus on Nordic countries (Dau 1829; Vaupell 1851; Blytt, 1876;  
249 Sernander, 1890, 1893; von Post 1924, 1930), and later in other regions of the northern  
250 hemisphere (Lewis, 1906, van Geel, 1978; Barber et al., 1994; Yu et al., 2014; Loisel et al.,  
251 2014). Most often, such studies have been based on plant macrofossils, pollen grains or peat  
252 decomposition, but the attention of many pioneer researchers was attracted by the presence of  
253 subfossil trees (Fig. 2 and 4) in peat stratigraphy (Dau 1829; Vaupell 1851; Blytt 1876; see also  
254 Birks and Seppä, 2010; Nielsen and Helama, 2012).

255 For example, Dau (1829) presented pioneering studies where he described the  
256 stratigraphic distribution of tree trunks in Danish peat bogs, and Vaupell (1851) studied buried  
257 trees and their relationship with soil nutrients, sunlight, and moisture conditions (Nielsen and  
258 Helama, 2012). Steenstrup (1842) classified tree horizons on the basis of species composition and  
259 interpreted tree population shifts as environmental changes associated with temperature and  
260 hydrological shifts during the postglacial period (Birks and Seppä, 2010; Nielsen and Helama,  
261 2012). For these observations, Steenstrup can be considered as one of the founders of Holocene  
262 palaeoecology and climate research (Birks and Seppä, 2010). In Norway, Blytt (1876)  
263 investigated tree horizons in relation to variations in peat decomposition, and interpreted

264 synchronous changes as shifts controlled by climate. The Swedish researcher Sernander thereafter  
265 improved the methods introduced by Blytt (1876) and associated dark peat layers containing  
266 wood material with periods of relatively dry continental climate, whereas less humified peat  
267 layers without fossil trees were interpreted as humid climatic periods (Sernander, 1890, 1893).  
268 Sernander also developed a schematic model of postglacial climate change, known as the Blytt-  
269 Sernander scheme, which has been used by many generations of Quaternary geologists and  
270 palaeobotanists (e.g., Smith and Pilcher, 1973; Blackford and Chambers, 1991; Birks and Seppä,  
271 2010). Outside Scandinavia, pine trees exposed in erosion scars in blanket peats across the  
272 Scottish Highlands (Lewis, 1906, 1908) were interpreted as evidence of constant changes in the  
273 character of the vegetation covering the Scottish peatlands.

274         During the early twentieth century, palaeoclimatological research was improved  
275 significantly by pollen analysis, introduced by von Post (1924, 1930). Moreover, stratigraphic  
276 studies by Granlund (1932) showed that abrupt peat humification shifts, referred to as “recurrence  
277 surfaces” (Fig. 2), were associated with rapid climatic shifts towards colder and wetter  
278 conditions. Following the introduction of radiocarbon dating, the studies by e.g., Aaby (1976) and  
279 Barber (1981) were able to link peat humification shifts to actual climate changes. Currently,  
280 climate reconstructions based on peatlands are common in the palaeoclimatic literature (e.g.,  
281 Langdon and Barber, 2005; Mauquoy et al., 2008) and generally based upon several parameters  
282 such as peat humification, pollen data, plant macrofossils, and testate amoebae (e.g., Hughes et  
283 al., 2006; Swindles et al., 2013).

284         Even though tree horizons in peat deposits were discovered over a century ago and their  
285 potential role as indicators of environmental changes was rapidly realized, subfossil trees in  
286 peatlands remained an underutilized source of palaeoclimatic information for a long time. Then,  
287 in the late 1960s, a number of projects were initiated with the aim of developing tree-ring

288 chronologies using subfossil trees excavated from peatlands in northwest Europe (Pilcher, 1973;  
289 Pilcher et al., 1977). This approach was made possible by the large number of well-preserved tree  
290 remains that have been exposed by peat extraction for fuel and horticulture (Fig. 4-5, Joosten and  
291 Clarke, 2002). During the 1970s and 1980s, oak samples from such peatlands in Ireland and  
292 Germany were used for the establishment of the first multi-millennial peatland tree-ring  
293 chronologies (Pilcher 1973; Pilcher et al., 1977, 1984; Leuschner and Delorme, 1984). At a very  
294 early stage, dendrochronological studies proved that subfossil wood material originating from  
295 different Irish peatlands could be cross-dated (Pilcher, 1973; Pilcher et al., 1977), which  
296 encouraged further work. Cross-dating is a standard procedure in dendrochronology in which  
297 individual tree-ring width (TRW) series are statistically and visually compared (Douglass, 1941;  
298 Fritts, 1976; Wigley et al., 1987). Several statistical parameters are normally tested, of which the  
299 t-test (Baillie and Pilcher, 1973) is the most commonly used (Table 1) and referred to when cross-  
300 dating statistics are presented (Fig. 6-7).

301         During the 1970s, floating tree-ring chronologies were developed from oaks found in  
302 multiple peatlands across Ireland (Pilcher, 1973; Pilcher et al., 1977). In parallel, an Irish oak  
303 chronology from living and archaeologically excavated trees was developed by Baillie (1973,  
304 1995), which would later enable absolute dating of some of the floating chronologies to obtain a  
305 continuous tree-ring chronology spanning 7272 years (Pilcher et al., 1984). The chronology  
306 proved to be a valuable calendric climate record and provided a standard for the radiocarbon  
307 calibration curve (Pilcher et al., 1984; Pearson et al., 1986; Reimer et al., 2013). As the Holocene  
308 record of subfossil peatland trees is of significance to both palaeoclimatologists and archeologists  
309 the literature contains a mixture of chronologies that quote ages as both BP and CE/BCE.  
310 Although, for the purposes of a review it may seem appropriate to adopt one scheme or the other,  
311 however, this has inherent problems for both of the relevant communities. Consequently, in this

312 study where we refer to an absolute age we quote the age in CE/BCE with the BP alternative in  
313 brackets, but where an event is known by a BP age, we quote the cal. BP age (calibrated years  
314 before AD 1950) and refer to an approximate BCE age in brackets.

315         Meanwhile, tree-ring chronologies for environment and climatic reconstructions were also  
316 constructed in Germany (Bauerochse et al., 2015). In this case, thousands of trees preserved in  
317 central and north German riverine sediments (Becker and Schirmer, 1977; Becker, 1983;  
318 Friedrich et al., 2004) and in north German peatlands (Leuschner and Delorme, 1984) were used.  
319 Due to stepwise and long-distance cross-dating it was possible to age determine previously  
320 undated material and to correct problematic and falsely dated sections (Fig. 6-7; Pilcher et al.,  
321 1984; Brown and Baillie 2012). Ever since then, the use of stepwise cross-dating has given an  
322 absolute age to tree-ring chronologies from England (Chambers et al., 1997; Lageard et al. 1999;  
323 Boswijk and Whitehouse, 2002), Scotland (Moir et al., 2010; Moir, 2012) the Netherlands  
324 (Jansma, 1996; Sass-Klaassen and Hanraets, 2006), Denmark (Christensen, 2007) and Sweden  
325 (Edvardsson et al., 2012a, 2014b).

326         In the late 1960s, Munaut (1967) and Smith et al. (1972) successfully developed the first  
327 tree-ring chronologies from subfossil peatland pines. However, pine trees were often considered  
328 as unreliable for chronology construction due to problems such as missing and wedging rings  
329 (Fig. 5). Attempts in Scotland during the 1980s, for example, failed to cross-date approximately  
330 200 samples (Bridge et al., 1990). Radiocarbon dates later showed that the samples had a large  
331 temporal range, which explained parts of the problems with cross-dating. By contrast, more  
332 recent studies focusing on Irish, Scottish, and British peatland pine chronologies have been  
333 successful in cross-dating tree-ring records (Pilcher et al., 1995; Lageard et al., 2000; Moir et al.,  
334 2010).



335           The interest in peatland pines increased, especially as they were found to have the  
336 potential to extend further back in time than the Irish oak chronology (then at 5452 BCE or 7401  
337 BP; Pilcher et al., 1995). In the Lower Saxony area, northwest Germany, extensive work was  
338 initiated on peatland pines (Leuschner et al., 2007; Eckstein et al., 2009, 2011; Achterberg et al.,  
339 2015, 2016). The German pine chronologies have been constructed from some 4000 peatland  
340 trees (Bauerochse et al., 2015) and show a temporal spread over the period 6703-1165 BCE  
341 (8652-3114 BP; Achterberg et al., 2016). Scots pine has been shown to represent by far the most  
342 common tree species in Scandinavian (Gunnarson, 1999; Edvardsson et al., 2012a) and Baltic  
343 (Pukienė, 1997; Edvardsson et al., 2016a) peatlands and in northern Sweden, pine remains in  
344 peaty environments have been used for treeline reconstructions during the Holocene (Kullman,  
345 1987). However, these trees have proved to be unsuitable for chronology development due to of  
346 the smaller sample number and wide temporal range. In central and southern Sweden, by contrast,  
347 peatland pines have been used successfully for the development of multi-millennial tree-ring  
348 chronologies (Gunnarson 1999; 2008; Edvardsson et al., 2012a, b). In northern Finland, subfossil  
349 pine remains have been recovered from peatlands and lacustrine sediments of lakes with thick  
350 layers of organogenic riparian substrate (Eronen et al., 1999). Dendrochronological analyses of  
351 this material yielded a continuous tree-ring chronology over the mid- and late-Holocene (from  
352 5634 BCE, Eronen et al., 2002; Helama et al., 2008), with indications that pine population  
353 density and hydrologic conditions varied over time (Helama et al., 2005). The suitability of  
354 subfossil peatland pines for long-term climate and palaeohydrology reconstructions has now been  
355 shown in a range of studies (Pilcher et al., 1995; Lageard et al., 2000; Boswijk and Whitehouse,  
356 2002; Leuschner et al., 2007; Eckstein et al., 2009; Moir et al., 2010; Edvardsson et al., 2012a;  
357 Achterberg et al., 2016).

358

359

#### 360 **4. General interpretations of peatland tree-growth dynamics**

361

##### 362 *4.1. Tree growth, climate and hydrology*

363

364 High water levels in peatlands have profound effects on seed germination and seedling  
365 development (Kozłowski, 1997; Holmgren et al., 2015). Indeed, the activation of the  
366 physiological processes required for seed germination depends on oxygen (O<sub>2</sub>) supply. A water  
367 table located close to the peat surface will restrict O<sub>2</sub> availability to the embryo and thereby  
368 prevent or postpone seed germination (Dang et al., 1991; Kozłowski and Pallardy, 1997;  
369 Fragnière et al., 2015; Kozłowski et al., 2015). Only once the seeds germinate will fluctuations in  
370 water level determine seedling survival and modify the capacity of seedlings to elongate rapidly  
371 and protrude above the water level (Sacchi and Price, 1992). As a consequence, lowered water  
372 tables are typically associated with the colonization of trees due to improved growth conditions  
373 (Freléchoux et al., 2000; Eckstein et al., 2009; Edvardsson et al., 2015b).

374 Most often, the annual growth of peatland trees is highly dependent on the depth and  
375 variability of the water table beneath the tree root system (Boggie, 1972; Eckstein et al., 2009).  
376 Wetter conditions are usually characterized by growth patterns showing recurrent, narrow annual  
377 growth rings, multi-annual growth depressions, and sometimes missing rings (Leuschner et al.,  
378 2002; Fig. 5). Such moist and unfavourable growth conditions are associated with physical,  
379 chemical, and biological restrictions on tree growth, of which reduced availability of nutrients in  
380 the water unsaturated zone probably represents the most important limiting factor (Boggie, 1972;  
381 Mannerkoski, 1991; Linderholm, 2001). Moreover, nitrate is rapidly depleted by denitrification  
382 under conditions of soil hypoxia (Kozłowski and Pallardy, 1984). By contrast, the uptake of iron

383 (Fe) and manganese (Mn) is increased because of an increase in soluble ferric and manganic  
384 forms. Yet, abundant ferrous and manganous ions in wet soils may be toxic to some tree species  
385 (Crawford, 1989). Moist conditions reduce root growth of most woody plants by inhibiting root  
386 formation and branching, repressing growth of existing roots and mycorrhizae, and by promoting  
387 root decay (DeBell et al., 1984; Kozłowski, 1984; Kozłowski and Pallardy, 1997; Eckstein et al.,  
388 2011). Tree growth is inhibited by the lack of O<sub>2</sub> (Dang et al., 1991; Kozłowski and Pallardy,  
389 1997) and in particular, reduced macronutrients uptake (especially nitrogen, phosphorus, and  
390 potassium) driven by a reduction in O<sub>2</sub> supply to the roots (e.g., Osundina and Osonubi, 1989).  
391 Finally, a variety of toxic compounds such as aldehydes, organic acids, and ethanol accumulated  
392 in wet soils have been demonstrated to variously contribute to injury, growth reduction, and  
393 mortality of woody plants (Kozłowski, 1997; Vitas and Erlickytė, 2007).

394         Consequently, a change toward wetter climatic conditions would lead to rising water  
395 tables, causing shallower unsaturated zones (Schouwenaars, 1988; Hunt et al., 1999), and thereby  
396 reduce recruitment of new trees and growth reductions in trees already growing at the peat  
397 surface (Boggie, 1972; Linderholm et al., 2002; Edvardsson et al., 2015a). Soil moisture is  
398 therefore a key driver of both radial tree growth and tree population dynamics in peatlands.  
399 Growth depressions and dying-off phases in subfossil chronologies are therefore an expression of  
400 unfavourable growth conditions, linked to changes in site hydrology towards moister conditions  
401 (Leuschner et al., 2002; Edvardsson et al., 2012b; Scharnweber et al., 2015). Lowered water  
402 tables, on the other hand, are typically associated with enhanced radial growth or the colonization  
403 of peatland trees (Freléhoux et al., 2000; Eckstein et al., 2009; Edvardsson et al., 2015b). This  
404 hypothesis is supported by a comparison of climate-growth relationships obtained for pine trees  
405 growing on peat and adjacent mineral soils (Fig. 8). This comparison also clearly demonstrates  
406 that tree growth on mineral soils is significantly correlated with winter and early summer

407 temperatures, whereas a more complex, weaker and clearly site-dependent response to monthly  
408 temperature and precipitation is detected in peatland trees. The latter presumably reflects the  
409 multiannual synthesis of moisture variability and changing hydrology (Fig. 9), which could be  
410 further complicated by a hydrological lag and/or feedback response in peatlands (Linderholm et  
411 al., 2002; Edvardsson et al., 2015a; Edvardsson and Hansson, 2015).

412 Studies exploring climate-growth relationships have not been realized for subfossil  
413 peatland oaks so far. Indeed, during extended periods of the Holocene, deciduous oaks have  
414 established in peatlands across northwestern Europe (Pilcher et al., 1984; Leuschner et al., 2002),  
415 but any detailed interpretation of the subfossil oak data has long been problematic due to the  
416 absence of modern peatlands with oak tree cover (Sass-Klaassen, 2004). In a study by  
417 Scharnweber et al. (2015), however, a catastrophic rewetting of a drained, oak covered peatland  
418 in northeast Germany has provided the opportunity to study growth responses in oaks following  
419 well-documented hydrological changes. Similar to the work on pine, moisture is also shown to be  
420 the main growth-limiting factor in peatland oak trees.

421

#### 422 *4.2. Hydrological lag effects*

423 Several studies show that a hydrological lag response to changing precipitation in peatlands  
424 exists, meaning that some delay can be observed between actual climate changes and related  
425 hydrological shifts in the peatlands (Kilian et al., 1995; Waddington et al., 2014). Such delayed  
426 responses have in the past been attributed to (i) water transport in soils towards the peatlands  
427 (e.g., Cowan, 1965; Hillel, 1971), (ii) water interception or accumulation in plants (Shukla and  
428 Mintz, 1982; Seneviratne et al., 2010), (iii) delays between snowfall and associated snowmelt  
429 (Price and Schlotzhauer, 1999), and (iv) the process by which water is replacing air in the  
430 relatively large pore spaces of the unsaturated zone of peat (Ingram, 1983; Almendinger et al.,

431 1986). In addition, it is known that (v) peat soils are compressible such that changes in water  
432 content may result in volumetric fluctuations and thereby variations in surface elevation, which in  
433 turn can vary depending on the amount of water stored in the peat (Almendinger et al., 1986;  
434 Price and Schlotzhauer, 1999). These slow hydrological changes may therefore result in monthly  
435 to multiannual lags between climate and tree growth responses, as observed and discussed in  
436 studies by Linderholm et al. (2002), Edvardsson et al. (2014a, 2015a), or Edvardsson and  
437 Hansson (2015). Knowledge about such delays is important for our understanding of relations  
438 between hydrological variability and tree growth, and for all studies aiming at; (i) reconstructing  
439 past and future hydroclimate changes, (ii) long-term carbon storage capacity of northern  
440 peatlands, (iii) the prediction of future peatland vegetation dynamics, and (iv) an improved  
441 understanding of peatland vulnerability to changing climate (Waddington et al., 2014).

442

#### 443 *4.3. The influence of trees on peatland hydrology*

444 As discussed, surface moisture in peatlands is the most important factor governing tree growth  
445 and colonization (Boggie, 1972; Edvardsson et al., 2015b). Tree density may in turn influence  
446 surface moisture and ultimately carbon sequestration in the peatlands, as a result of feedback  
447 effects (Limpens et al., 2014a; Holmgren et al., 2015). Trees can affect the moisture status of  
448 peatland environments through transpiration and interception (Limpens et al., 2014b).  
449 Colonisation by trees therefore has the potential to cause a positive drying feedback, whereby  
450 increased root uptake for transpiration demand causes a lowering of the water table, in turn  
451 promoting improved tree growth and seedling recruitment and/or shrubification as root  
452 oxygenation and root zone thickness increase (Limpens et al., 2014b). This positive feedback is  
453 amplified further by; (i) increased soil nutrient availability by litter fall, (ii) increased levels of  
454 interception with increased canopy (Dubé et al., 1995), and (iii) reduced aerodynamic resistance

455 (by e.g., wind) favoured by higher stand density which increases evaporation (Kellner, 2001).  
456 The change in hydrology may also lead to increased decay rates in the upper peat profile due to a  
457 deepening of the aerobic zone. Changes in tree cover can therefore be a fundamental shift in the  
458 peatland carbon cycle and it is possible that tree-covered peatlands could become net sources of  
459 carbon if increased net photosynthetic rates and carbon fixation are lower than increased peat  
460 decomposition rates (Juutinen et al., 2010). The effects that tree density has on water loss have  
461 important implications for the structure and functioning of peatlands but the relation between  
462 evapotranspiration and tree density is non-linear (Limpens et al., 2014b). This is because  
463 increasing density of tree stands can result in shading of the peat surface, thereby reducing the  
464 energy available for evaporation (Ohlson et al., 2001; Strakova et al., 2012).

465

466

## 467 **5. Climate – peatland relations: drawing a continental-scale picture**

468

469 The present interglacial, the Holocene, started about 11,700 cal BP (calibrated years before AD  
470 1950) with a rapid transition from the cold Younger Dryas to subsequent conditions characterized  
471 by significantly warmer and more stable conditions (Bond et al., 2001; Wanner et al., 2008;  
472 Marcott et al., 2013). At the millennial timescale, the climate of the Holocene has been influenced  
473 strongly by hemispheric trends in solar insolation, but was also interrupted by shorter centennial  
474 events related to other drivers such as solar and volcanic forcing (Dansgaard et al., 1989; Bond et  
475 al., 2001; Wanner et al., 2011). High spatial and temporal variability temperature and moisture  
476 records covering the Holocene have been reconstructed from archives such as ice cores, ocean,  
477 lake, and peat sediment cores, speleothems and glacier fluctuations (Wanner et al., 2011).  
478 Climatic and environmental mismatches or delays between different geographical regions,

479 however, underline the importance of multiple site studies and call for high-resolution  
480 reconstructions using proxies such as tree rings, which remain the most widespread source of  
481 annual resolution, precisely dated palaeoclimatic data. In the following discussion, the Holocene  
482 is subdivided into the early (c. 9750-6250 BCE or 11,700–8200 cal. BP), mid (c. 6250-2250 BCE  
483 or 8200–4200 cal. BP), and late (c. 2250 BCE or 4200 cal. BP to present) Holocene as suggested  
484 by Walker et al. (2012).

485

### 486 *5.1. The early Holocene*

487 Subfossil peatland trees from the early Holocene are scarce for obvious reasons, primarily due to  
488 the fact that most northern peatlands were not ombrotrophic until later periods of the Holocene  
489 due to slow peat formation and accumulation (Barber et al., 2003; Yu et al., 2010). Despite this  
490 limitation, radiocarbon dated pine trees as old as c. 9050 BCE (11,000 cal. BP) have been  
491 discovered in northwestern German peatlands (Achterberg et al., 2016). The formation of raised  
492 bogs in this region, however, only began about 2000 years later (Petzelberger et al., 1999;  
493 Eckstein et al., 2011), which suggests that the material consists of mineral soil pines overgrown  
494 by and preserved in expanding peatlands. Tree colonization directly on peat soils of northwestern  
495 Germany has, however, been recorded about c. 6750 BCE (8700 cal. BP), evident from peatland  
496 pines dated to 6703 BCE (8652 BP; Achterberg et al., 2015) and oaks dated to 6627 BCE (8576  
497 BP; Leuschner et al., 2002; Achterberg et al., 2015). Oak trunks from this period, but dated by  
498 radiocarbon, have also been discovered in Denmark (Christensen, 2007). Some centuries later,  
499 about 6350 BCE (8300 cal. BP), minor periods of pine establishment have been noted in the Irish  
500 and Scottish pine material (Bridge et al., 1990; Torbenson et al., 2015). Otherwise, few  
501 establishment phases have been recorded prior to the 8200 cal. BP cold event (referred to as the  
502 8.2 ka event, c. 6250 BCE; Alley et al., 1997), which may have generated moist and unfavourable

503 conditions for tree colonization at the peatlands. The 8.2 ka event is recorded on a hemispheric  
504 scale in a broad range of proxy records (Alley et al., 1997; Veski et al., 2004; Nicolussi and  
505 Scüchter, 2012). The number of German pines recorded (replication in Fig. 10) drops during (or  
506 slightly after) the 8.2 ka event and a 108-year gap has been documented in the German bog oak  
507 record (Achterberg et al., 2016), indicating a shift towards unfavourable conditions for peatland  
508 tree growth (Fig. 10). Although from a different context, the lowest tree replication in the Alpine  
509 tree-ring chronology by Nicolussi et al. (2009) is observed around 6100 BCE, very close to the  
510 8.2 ka event. In a similar way, Helama et al. (2008) also suspect that the lack of subfossil trees in  
511 Finnish Lapland was caused by the same climatic effect and that it would likely have decelerated  
512 pine colonization in the region.

### 513 *5.2. The mid-Holocene*

514 About 8100 years ago, following the cold 8.2 ka event (Alley et al., 1997), tree establishment has  
515 been recorded in various peatlands in northwestern Europe, namely in Ireland (Pilcher et al.,  
516 1995; Torbenson et al., 2015), Germany (Leuschner et al., 2002; Eckstein et al., 2009; Achterberg  
517 et al., 2015), Denmark (Christensen, 2007), and Sweden (Edvardsson et al., 2012b). These initial  
518 pulses of tree establishment were followed by continuous and fluctuating mid-Holocene tree  
519 colonization across northwest European peatlands (Fig. 8).

520 During the mid-Holocene, cross-dating between series from different regions yields  
521 statistically significant results and points to striking similarities in annual growth and tree  
522 population dynamics over substantial parts of northwestern Europe, especially for the period 5300  
523 to 2500 BCE (7250-4550 BP; Fig. 6-7). The successful cross-dating between various sites,  
524 sometimes separated by hundreds of kilometres, also indicates that similar year-by-year climate  
525 variability has influenced environmental conditions at these peatlands, and thereby radial growth



526 of trees (Pilcher et al., 1977, 1984; Leuschner et al., 2002; Eckstein et al., 2009; Edvardsson et  
527 al., 2012a). Similarities in tree replication records on the other hand, show coinciding  
528 regeneration and later life stages, whereas mean-age records indicate changes in the age structure  
529 of tree populations. The German pines, for instance, were not randomly distributed over time.  
530 Instead, woodland establishment and dying-off phases were periodic, as detected in pine  
531 populations from 36 different German peatlands (Eckstein et al., 2009). Together, the significant  
532 cross-dating statistics and similarities in tree population dynamics imply similar high- (i.e. annual  
533 to decadal scale) and low-frequency variability (i.e. decadal to centennial scale) governed  
534 hydroclimatic forcing mechanisms influencing peatland hydrology and associated tree growth  
535 over a large geographic region (Fig. 6 and 10). Moreover, widespread colonization also points to  
536 relatively dry and therefore favourable growth conditions for peatland trees, which is underlined  
537 by the main woodland phases at German and Swedish peatlands between 5300 and 3700 BCE  
538 (7250-5650 BP; Eckstein et al., 2009, 2011; Edvardsson et al., 2012a, b).

539 Pine records from the Alps show colonization at unusually high elevations (Tessier et al.,  
540 1993) and strong replication (Nicolussi et al., 2009) during the same period. A combination of  
541 high-altitude occurrence and an increased replication of subfossil pines is also evident in northern  
542 Finland from 4200 to 2200 BCE (6150-4150 BP; Helama et al., 2004, 2010). The pine  
543 colonisation of peatlands in northern Scotland occurred around 5000 years ago, and has been  
544 interpreted as a regional phenomenon (Gear and Huntley, 1991; Charman, 1994; Daniell, 1997;  
545 Huntley et al., 1997). By contrast, both oak and pine trees colonized Baltic peatlands during the  
546 mid-Holocene (Pukienė, 2003; Vitas, 2010; Edvardsson et al., 2016a), and therefore add further  
547 evidence of widespread impacts of favourable climatic conditions.

548 These favourable growth conditions for peatland trees during the mid-Holocene were  
549 probably closely linked to a relatively warm and dry climate regime, which commonly is referred

550 to as the Holocene Thermal Maximum (HTM; Seppä et al., 2005; Renssen et al., 2009). The onset  
551 and termination of the HTM varies clearly between different geographical regions and studies,  
552 but is generally dated to the period c. 6850-2450 BCE (8000–4400 cal. BP) across northern  
553 Europe (Jessen et al., 2005; Seppä et al., 2005; Renssen et al., 2009; De Jong et al., 2009). The  
554 comparably high air temperatures and dry conditions in northwestern Europe were probably  
555 caused by (i) an orbitally-forced increase in summer insolation, and (ii) a stronger meridional  
556 overturning circulation and intensified northward ocean heat transport after the Laurentide Ice  
557 Sheet had vanished (Snowball et al., 2004; Renssen et al., 2009; Wanner et al., 2011).  
558 Temperature reconstructions based on stable isotopes from ice-cores as well as pollen data from  
559 lake sediments and peat cores all point to summer temperatures which would have been  
560 approximately 2-3°C above present values across most parts of northwestern Europe and  
561 Scandinavia (Huntley and Prentice, 1988; Seppä et al., 2005; Brown et al., 2011; Helama et al.,  
562 2012). At the same time, moisture records indicate a dry phase at about 5250 to 3750 BCE (7200-  
563 5700 BP; Wanner et al., 2011), which is confirmed indirectly by lake level proxies pointing to  
564 generally low water levels (Digerfeldt, 1988; Hammarlund et al., 2003; Magny, 2004, 2013;  
565 Sohar and Kalm, 2008).

566         The annual resolution and accurate dating of the tree-ring chronologies allow detailed  
567 studies of rapid climatic events during the HTM, which may be missed in lower resolution  
568 proxies. Several peaks in the Irish replication records have, for instance, been recorded about  
569 6050, 5350, 4250, and 1250 BCE (8000, 7300, 6200 and 3200 BP; Turney et al., 2006), and are  
570 consistent with peaks and establishment phases detected in replication records from Germany, the  
571 Netherlands, and Sweden (Fig. 10). Moreover, many dying-off phases described in the German  
572 records (Leuschner et al., 2002; Eckstein et al., 2009; Bauerochse et al., 2015; Achterberg et al.,  
573 2015) can be found in other records from northwestern Europe. For instance, the dying-off phase

574 between 4600-4550 BCE (6550-6500 BP) described by Achterberg et al. (2015) is also visible in  
575 the oak records from Germany and the Netherlands (Leuschner et al., 2002), as well as in the  
576 Scottish (Bridge et al., 1990) and Swedish (Edvardsson et al., 2012a) pine records.

577         Towards the end of the HTM, shifts towards wetter conditions about 3250 BCE and 2250  
578 BCE (5200 and 4200 cal. BP) have been described in the literature (referred to as the 5.2 and 4.2  
579 ka events; Roland et al., 2014, 2015). The 5.2 ka event (c. 3250 BCE) has been identified as a  
580 coherent shift towards wetter conditions caused by prolonged positive North Atlantic Oscillation  
581 conditions, and was found in stable isotope, testate amoebae, plant macrofossil, and humification  
582 records from peatlands in Northern Ireland (Roland et al., 2015). Evidence for such a wet shift is  
583 lacking in the Irish oak replication records, whereas a decrease in German peatland pines and a  
584 dying-off phase in Lithuanian pines can be observed (Fig. 10). Moreover, a coincident phase of  
585 lowered replication of subfossil pines is evident in northern Finland (Helama et al., 2004, 2010).  
586 Several synchronous and massive dying-off phases can also be detected about 2500 BCE (4450  
587 BP) in Scotland (Gear and Huntley, 1991; Charman, 1994; Daniell, 1997), Ireland (Pilcher et al.,  
588 1995), Great Britain (Lageard et al., 1999; Boswijk and Whitehouse, 2002), Germany (Eckstein  
589 et al., 2011; Bauerochse et al., 2015), and Sweden (Edvardsson et al., 2014b), and thus point to a  
590 spatially structured, regional environmental change. By way of example, the British peatland pine  
591 chronologies from Garry Bog (2569 BCE; Pilcher et al., 1995), White Moss (2559 BCE; Lageard  
592 et al., 1999), Thorne (2475 BCE; Boswijk, 1998), and Hatfield (2445 BCE; Boswijk and  
593 Whitehouse, 2002) all terminate at  $2507 \pm 62$  BCE ( $4456 \pm 62$  BP).

594

### 595 *5.3. The late Holocene*

596 The unfavourable conditions recorded towards the very end of the HTM continued in a similar  
597 manner between 2200 and 2160 BCE (4150-4110 BP; Fig. 10), and resulted in abrupt dying-off

598 phases at various sites in northwestern Germany (Eckstein et al., 2010; Bauerochse et al., 2015),  
599 the Netherlands (Leuschner et al., 2002), and Finland (Helama et al., 2004). The similarities  
600 between continental and Irish mean-age records, which were present during the entire HTM, also  
601 start to disappear at about 2000 BCE (4000 BP; Leuschner et al., 2002). The effects of a large-  
602 scale shift in climatic conditions are also obvious in northern Finland where the replication of the  
603 subfossil pine chronology collapses around 2200-2000 BCE. Moreover, indications of pine  
604 germination are clearly missing at the high elevation sites in Finland as of 2400 BCE (4350 BP),  
605 and are followed by a gradual decline in pine populations until 600 BCE (2550 BP; Helama et al.,  
606 2004). Since about 1500 BCE (3450 BP), following yet another dying-off phase recorded across  
607 northwestern European peatlands, fewer dated peatland pines have been found in Germany  
608 (Achterberg et al., 2015) and Ireland (Pilcher et al., 1973). Most of these observations can be  
609 linked to the onset of a widespread climatic transition, sometimes referred to as the Neoglacial  
610 phase (Nesje et al., 1991; Wanner et al., 2008).

611         The climatic change about 2250 BCE (4200 cal. BP and often referred to as the 4.2 ka  
612 event; Roland et al., 2014) is evident at the hemispheric scale in numerous proxy records and was  
613 most likely driven by changes in orbital insolation patterns (Bradley, 2003; Booth et al., 2005;  
614 Morley et al., 2014), but changes in tree populations do not appear abruptly or synchronously  
615 between different sites (Fig. 10). Instead, a multi-proxy study from Sweden points to a stepwise  
616 transition phase dated to between c. 2650 and 1450 BCE (4600 and 3400 cal. BP; Jessen et al.,  
617 2005). In a similar way, several dying-off events have been recorded among Swedish peatland  
618 trees (Edvardsson, 2016), whereas a two-stage temperature decrease with an onset c. 3050-2550  
619 BCE (5000-4500 cal. BP) and a termination c. 2350-1350 BCE (4300-3300 cal. BP) has been  
620 described around the Baltic Sea (Borzenkova et al., 2015). Moreover, testate amoebae in Irish  
621 peatlands do not provide any compelling evidence for a rapid transition about 2250 BCE (4200

622 cal. BP; Roland et al., 2014), whereas records from Great Britain and Denmark suggest a system  
623 shift at about 2150 BCE (4100 cal. BP; Mauquoy et al., 2008). These results show that the origin  
624 and impact of the 4.2 ka event is spatially complex and was probably less marked than in low  
625 latitudes, and may also explain why the drop in tree replication in Ireland and Sweden started 200  
626 years later than in Germany and the Netherlands (Fig. 10). These results also confirm the need for  
627 climate reconstructions with high spatio-temporal resolution to improve our understanding of  
628 complex climate transitions. Pollen-based temperature reconstructions from Scandinavia show  
629 that the annual mean temperature was reduced by almost 2 °C during this transition phase (Seppä  
630 et al., 2005; Heikkilä and Seppä, 2010), whereas lake-level reconstructions suggest wetter  
631 conditions with rising lake levels between 3000 and 500 BCE (5000-2500 cal. BP; Digerfeldt,  
632 1988; Magny, 2004, 2013; Sohar and Kalm, 2008). Indications of rising water levels after the  
633 termination of the HTM also exist for northern Finland in the form of submerged pines with their  
634 root systems anchored in the palaeo-shoreline, two meters below the current water level (Eronen  
635 et al., 1999). The rising water tables and wetter peatland surface conditions are therefore also  
636 believed to have caused widespread dying-off phases, leading to the termination of most Swedish  
637 and German pine records at about 1100 BCE (3050 BP; Eckstein et al., 2011; Edvardsson, 2016).  
638 Together, these observations indicate that the widespread transition towards wetter and colder  
639 conditions continued for at least a millennium after the end of the HTM. These findings are also  
640 confirmed by radiocarbon-dated recurrence surfaces detected in several Swedish peat sequences,  
641 for which the wet shift is synchronous with the final dying-off phases in the German and Swedish  
642 pine chronologies (Borgmark and Wastegård, 2008; Rundgren, 2008) and peat records from  
643 northern Britain (Charman et al., 2006).

644         Although the period 2250-1100 BCE (4200-3050 BP) is in general associated with  
645 increasingly wetter and colder conditions, some short and regional colonization events were still

646 taking place. The German pine colonization phase at around 1900 BCE (3850 BP), described by  
647 Bauerochse et al. (2015), is also visible in records from Sweden, the Netherlands, and Scotland  
648 (Fig. 10). Over the last 3000 years, however, relatively few tree establishment phases have been  
649 recorded in German and Irish peatlands, whereas trees appear to be close to absent at all known  
650 Swedish sites (Fig. 10). Despite this fact, tree colonization has been recorded at peatlands in the  
651 Netherlands (Sass-Klaassen and Hanraets, 2006), Poland (Barniak et al., 2014), Lithuania  
652 (Pukienė, 1997; Edvardsson et al., 2016a), and in one of the few regions documented in North  
653 America (Québec; Arseneault and Payette, 1997). With the exception of the Canadian records,  
654 these phases are relatively short, which may also explain why only the Dutch chronologies have  
655 been cross-dated successfully over long distances (Fig. 6-7). It is therefore difficult to judge to  
656 what extent these tree colonization phases are related to site-specific variations or are linked to  
657 regional climate dynamics.

658 About 850 BCE (2800 cal. BP), yet another shift towards cooler, wetter, and windier  
659 conditions has been recorded in European peatlands (referred to as the 2.8 ka event; van Geel et  
660 al., 1996, 2014; Mauquoy et al., 2008; Mellström et al., 2015). Observations made in peatlands  
661 are confirmed by other proxy records, which suggest rising lake levels (Magny, 2004, 2013) and  
662 more drift ice in the North Atlantic (Bond et al., 1997). The nature and causes of this cold and  
663 wet event have been subject to debate, but the shift broadly coincides with an increase of the  
664 atmospheric  $^{14}\text{C}$  concentration due to reduced solar activity (Kilian et al., 1995; Swindles et al.,  
665 2013; Mellström et al., 2015). The Irish replication and mean-age records indicate dying-off  
666 phases or generation shifts, whereas the German and Dutch series point to a decline in replication  
667 records (Fig. 10). A dying-off phase about 850 BCE has also been detected in the Polish tree-ring  
668 chronologies (Barniak et al., 2014). Despite the general absence of dated peatland pine  
669 chronologies from Sweden, macrofossil analysis from a southwest Swedish peat bog shows that

670 pine must have been present locally at the site prior to 910 BCE (2860 cal. BP), but was absent  
671 after 740 BCE (2690 cal. BP; Mellström et al., 2015). Gaps also exist in Lithuanian (Edvardsson  
672 et al., 2016a) and Canadian (Arseneault and Payette, 1997) tree records during the same period,  
673 which may strengthen the hypothesis that the 2.8 ka event did indeed significantly influence large  
674 parts of the northern Hemisphere.

675         Between c. 500 BCE and CE 300 (2500 and 1600 BP), tree replication records from  
676 Ireland, Germany, and the Netherlands again show some tree colonization phases, but of minor  
677 importance (Fig. 10). These colonizations coincide with somewhat warmer and drier climatic  
678 conditions during the period often referred to as the Roman Warm Period (RWP; Seneviratne et  
679 al., 2010; Büntgen et al., 2011). Partial agreements between the Irish and continental mean-age  
680 records presumably indicate changes between relatively similar and diverse forcing mechanisms  
681 between the two regions over the period (Leuschner et al., 2002). In the Netherlands, a shift in  
682 tree population dynamics with establishment of both oak and ash trees has been recorded as late  
683 as AD 300 (1600 cal. BP) followed by striking changes in site hydrology and a dying-off phase at  
684 about AD 530 (Sass-Klaassen and Hanraets, 2006).

685         Dying-off phases recorded at several peatlands in Lithuania (Pukienė, 1997; Edvardsson  
686 et al., 2016a), by contrast, coincide with the well-known cooling at about 300-600 AD (1650-  
687 1350 BP), which is commonly referred to as the Dark Age or Migration Period (Wanner et al.,  
688 2011). At about AD 800 (1200 cal. BP), massive tree establishment has been recorded in  
689 Lithuanian peatlands (Pukienė, 1997; Edvardsson et al., 2016a) and at the same time, peat records  
690 from both Great Britain and Denmark point to a dry phase (Mauquoy et al., 2008). These changes  
691 coincide with the onset of the Medieval Warm Period (MWP) and indicate yet another change  
692 towards drier conditions in peatlands. Furthermore, Arseneault and Payette (1997) have reported

693 strong radial tree growth over the corresponding period, which indicates relatively favourable  
694 conditions for tree growth at Canadian peatlands as well.

695 By contrast, an almost complete absence of subfossil peatland trees has been recorded  
696 during the Little Ice Age (LIA; AD 1350-1850), which clearly indicates colder and moister  
697 conditions compared to the preceding periods (Wanner et al., 2008) and a widespread lack of tree  
698 establishment at peatland sites (Fig. 10). Widespread wet shifts have also been reported in  
699 European peat stratigraphic records at about CE 1350 and 1600 (600 and 350 cal. BP; Mauquoy  
700 et al., 2008). Apparently these conditions generated growth conditions that were too harsh for  
701 peatland trees as the final dying-off phases, recorded at about AD 1400 (550 cal. BP), coincide  
702 with a shift towards colder (Wanner et al., 2011) and wetter (Charman et al., 2006) climatic  
703 conditions. These observations are also supported by the fact that finding recent peatland pines  
704 older than 200 years has proven very difficult (Linderholm et al., 2002; Cedro and Lamentowicz,  
705 2011; Edvardsson et al., 2015a), confirming the likely absence of peatland trees during the LIA.  
706 As the LIA came to an end, widespread pine colonization occurred in peatlands across Estonia  
707 (Smiljanić et al., 2014), Lithuania (Edvardsson et al., 2015a), Poland (Cedro and Lamentowicz,  
708 2011), and Sweden (Linderholm and Leine, 2004; Edvardsson and Hansson, 2015). Observed tree  
709 establishment in recent periods is likely to result from a combination of climatic and land-use  
710 changes. Nevertheless, changing climatic conditions over the twentieth century have, at least in  
711 some cases, been regarded to be the most important driver of increasing tree establishment rates  
712 (Edvardsson et al., 2015b).

713

714

## 715 **6. Target areas for future research**

716



717 There is still a need for improved understanding of the relations between peatland tree growth,  
718 moisture variability, and climate dynamics, to enable more robust hydrology and climate  
719 interpretation from peatland tree-ring, replication, and mean-age records. For example, there is a  
720 clear need for new methods to isolate the temperature and precipitation signals in tree-growth  
721 patterns. Here, we address several suggestions for future studies, which will help to advance our  
722 understanding of peatland tree-growth variability, as well as of the connections between tree  
723 growth, peatland hydrology, and climate dynamics. We also discuss potential new research fields  
724 using subfossil peatland trees.

725

## 726 **6.1. Isolating and amplifying the climate signal**

727

### 728 *6.1.1. Long-term reconstructions using multiple proxy records*

729 To understand the basic functions of peatland tree growth and colonization, tree growth and  
730 population dynamics data need to be compared and coupled with climate-moisture proxies.  
731 Studies using peat stratigraphic records have been carried out to detect large-scale hydrological  
732 shifts to put the subfossil tree record into a long-term environmental context (Gunnarson et al.,  
733 2003; Eckstein et al., 2010, 2011; Edvardsson et al., 2012b, 2014b). To advance these  
734 interpretations further, comparisons using pollen (Gear and Huntley, 1991; Eckstein et al., 2010,  
735 2011; Edvardsson et al., 2014b), plant macrofossils (Edvardsson et al., 2014b), and peat  
736 humification (Edvardsson et al., 2012b) have been undertaken. The disadvantage of such proxy  
737 records is their multi-annual temporal resolution and lower temporal precision that radiocarbon  
738 ages offer compared to trees. To improve comparative long-term studies further, proxy records of  
739 higher, preferably annual resolution are required. Such records can sometimes be obtained from  
740 laminated lake sediments (Veski et al., 2004; Tylmann et al., 2013), ice cores (Vinther et al.,

741 2006), speleothems (Labuhn et al., 2015), corals (Zinke et al., 2014), and highly resolved peat  
742 stratigraphic records (Gear and Huntley, 1991; Amesbury et al., 2012). The disadvantage of most  
743 annually resolved data sources is that they do not record peat hydrological changes directly,  
744 although they do provide an independent record of hydroclimate variability against which to test  
745 the dendroecological time series.

746 Further in-depth comparison between multi-millennial tree-ring chronologies representing  
747 different environments and soil types is another approach that is generating valuable  
748 paleoenvironmental information. Such comparative studies have been performed using living  
749 pine trees (Linderholm et al., 2002; Cedro and Lamentowicz, 2011; Edvardsson et al., 2015a), but  
750 Bauerochse et al. (2015) recently started to compare subfossil Hohenheim oaks conserved in river  
751 sediments and gravel pits (Spurk et al., 2002; Friedrich et al., 2004) to peatland oaks from Lower  
752 Saxony (Leuschner et al., 2002). Although comparisons showed promising potential, it would be  
753 premature to draw any conclusions as the focus was on a relatively short period of the past (2500-  
754 2000 BCE or 4450-3950 BP). We call for further comparisons of continuous material such as the  
755 German oak chronologies or attempts to correlate Scandinavian peatland tree-ring reconstructions  
756 with the Torneträsk (Grudd et al., 2002) or Jämtland (Gunnarson et al., 2003) chronologies, as  
757 moisture and temperature sensitive records would then be compared over multi-millennial time  
758 scales.

759

### 760 *6.1.2. Assessment of hydroclimate and tree growth relationships*

761 Dendroclimatic studies using living peatland trees and meteorological data represent yet another  
762 field, which has remained largely underexplored (Cedro and Lamentowicz, 2011; Smiljanić et al.,  
763 2014), but which has potential to increase our understanding of relations between peatland tree  
764 growth and climate variables. In addition, previous work may not have been conclusive due to

765 hydrological lag effects and poorly understood feedback mechanisms in peatlands (Linderholm et  
766 al., 2002; Edvardsson et al., 2015a; Edvardsson and Hansson, 2015). Most studies have  
767 consistently reported a negative correlation between annual tree growth and precipitation records,  
768 yet correlation statistics have mostly been insignificant (Fig. 8). The use of continuous growth  
769 monitoring devices (such as manual or automated dendrometers), micro-meteorological stations  
770 and local hydrological stations (i.e. piezometers) could be a valuable way to address relations  
771 between meteorological events, water-table fluctuations, and tree growth in more detail and with  
772 fewer uncertainties. The combination of such monitoring records may also provide key insights  
773 into the hydrological functioning of peatland ecosystems, and especially into lag effects and their  
774 impact on tree growth. This hypothesis of a lag effect is supported by unpublished data from  
775 Čepkeliai peatland (southeastern Lithuania) where TRW records (AD 1848-2014) have been  
776 correlated with hydrological and meteorological time series from the nearby Varėna station (AD  
777 1951-2013). At Čepkeliai, bootstrap correlation functions between the TRW chronology and  
778 monthly temperature and precipitation only show significant correlation ( $r=0.25$ ,  $p<0.05$ ) with  
779 previous October precipitation (Fig. 8). At the same time, comparison between the TRW  
780 chronology and a much shorter, but local peatland water-table record (AD 2001-2014) shows  
781 significantly stronger negative correlation ( $r=-0.67$ ,  $p<0.05$ ; Fig. 11). Although the water-table  
782 record only covers the last 13 years, these first correlation tests point to the potential of a  
783 combined approach. At the same time, they suggest that peatland tree growth might be a valuable  
784 direct measurement of moisture changes in peatlands, but a poorer proxy for direct precipitation  
785 reconstructions. As long instrumental hydrological records do not exist for peatlands, we suggest  
786 progress is made by: (i) testing for correlations between TRW series and hydrometeorological  
787 variables, such as the ratio of actual to potential evapotranspiration (AET/PET) computed from  
788 temperature and precipitation values; (ii) using tree-growth process models driven by climate and

789 other environmental factors (such as nutrient availability and soil moisture) as they might allow  
790 better understanding of tree functioning and the effect which the interaction of different  
791 environmental forcings has on tree growth (see Guiot et al., 2014 for a review); and (iii) deriving  
792 longer-term water-table fluctuations series from locally-calibrated, process-based hydrological  
793 models (see Gong et al., 2012 for a review on the topic) driven by local climatic parameters  
794 (temperature, precipitation).

795

### 796 *6.1.3. Stationarity in tree growth–climate relationships?*

797 Dendroclimatology commonly assumes stationarity in the relationship between tree growth and  
798 climate, and that trees are reacting to environmental changes in the same manner at present as  
799 they have in the past (uniformitarian principle; Fritts, 1976). This principle has, however, been  
800 challenged due to the divergence (Briffa et al., 1998a, b; Wilmking et al., 2005; D'Arrigo et al.,  
801 2008) of tree growth at some Nordic sites which used to be temperature-limited in the past, but  
802 show a weakening in their mean temperature response in recent decades as a result of global  
803 warming. One should thus also assume changes in tree growth–climate relationships at longer  
804 timescales as both climate and peatlands have undergone substantial changes over the course of  
805 the Holocene. As a consequence, results and conclusions drawn from living material may not  
806 therefore be directly transferable to subfossil trees. Trees in Swedish peatlands were shown to be  
807 smaller and less long-lived under current conditions (average diameter 16 cm, age 105 yrs.) as  
808 compared to the mid-Holocene when the same species produced significantly larger and older  
809 trees (average diameter 25-30 cm and age 175 yrs.; Lindbladh et al., 2013). These differences can  
810 be explained by more favourable climatic conditions and shallower peatlands in earlier periods,  
811 but that may not be the only explanation. Further studies should specifically look into age, size,  
812 and radial growth relations over various periods of the Holocene. Moreover, the development of

813 dedicated standardisation techniques, designed to remove internal, non-climatic growth trends in  
814 ring-width series, with an aim to account for tree productivity-related biases needs further  
815 evaluation. For example, the Signal-Free Regional Chronology Standardisation (Melvin and  
816 Briffa, 2014) approach should be preferred for reconstructions if they contain both living and  
817 subfossil peatland trees.

818

#### 819 *6.1.4. What does tree replication mean?*

820 Tree replication data are often used as a climate proxy, but this has recently caused debate about  
821 its value for climate inferences and on ways to interpret such records (Charman, 2010; Swindles  
822 and Plunkett, 2010; Swindles et al., 2013; McGeever and Mitchell, 2015). Looking at the 2500  
823 BCE–CE 1600 period, Swindles and Plunkett (2010) reported a lack of consistent relationship  
824 between the Irish oak replication (Turney et al., 2005) and palaeohydrological records (Mauquoy  
825 et al., 2008; Magny, 2004, 2013), changing solar activity (Hu et al., 2003; Muscheler et al.,  
826 2005), or stacked drift ice records (Bond et al., 2001). However, Charman (2010) successfully  
827 reconciled the Irish bog oak record with a large-scale compilation of peatland water table  
828 reconstruction and other evidence from northern Britain, and showed the commencement of  
829 declines in Irish oak populations were coincident with increased surface wetness in northern  
830 British peatlands. This suggests that it is the direction of change in the population that is  
831 sensitive to surface wetness, rather than the absolute numbers of trees in the record at any one  
832 time. Disagreement between records in Swindles and Plunkett (2010) could be a consequence of  
833 the mid-Holocene transition towards moister conditions about 2500 BCE (4450 BP; Wanner et  
834 al., 2008), which caused divergent reactions between Irish and continental oak records  
835 (Leuschner et al., 2002). It may also be a function of comparison with atypical surface wetness  
836 records and the relatively small number of records after c. 2200 years ago (Pilcher et al., 1996).

837           McGeever and Mitchell (2015) looked into the temporal distribution of Irish peatland  
838 pines dated by radiocarbon, and suggested that factors other than climate, such as the availability  
839 of pine seeds and peatland surface area, may have influenced the presence of trees at Irish  
840 peatlands and that this limitation might have disrupted the climate signal of the record during  
841 certain periods. Despite this possible shortcoming, their pine record nevertheless shows a  
842 significant peak during the late part of the HTM, a decline during the late-Holocene transition, an  
843 absence of material during the 8.2 and 2.8 ka events, as well as during the LIA (Fig. 10).

844           Whilst there is a sound theoretical underpinning and increasing evidence for interpreting  
845 tree records in terms of palaeoclimate, we acknowledge that tree replication will be influenced by  
846 factors other than climatic wetness changes (Swindles and Plunkett, 2010; Swindles et al., 2013;  
847 McGeever and Mitchell, 2015), such as natural succession and the transformation of a peatland  
848 from a wetland to a raised bog (Fig. 2). Moreover, tree replication in different geographical  
849 settings may not reflect the same environmental changes. Competition between tree species  
850 colonizing peatlands in parallel is another possible factor influencing tree replication. In-depth  
851 comparative studies between overlapping records from *in-situ* pine and oak trees from common  
852 study sites may generate valuable information on competition between species, habitat  
853 preferences and tree population dynamics. Furthermore, research on processes governing  
854 peatland tree colonization in different geographical regions might help us to understand if  
855 establishment and dying-off phases reflect similar environmental processes or if local controls are  
856 more important.

857

## 858 **6.2 New study areas and research fields**

859

### 860 *6.2.1. New study areas*

861 This review provides a literature-based overview of geographical areas for which tree-ring  
862 analyses have been performed with subfossil peatland trees (Fig. 1 and Table 1). The distribution  
863 of peatland regions in the Northern Hemisphere (Fig. 1) also enables us to identify new potential  
864 study areas, at least for living peatland trees. In that sense, we hypothesize that Québec, Ontario,  
865 and the Northern regions of Canada in more general terms, Southern Siberia (Russia), as well as  
866 the Baltic countries clearly represent areas in which the potential of peatland dendrochronology  
867 has been underutilized in the past, as shown by a series of very recent studies on Lithuanian  
868 (Edvardsson et al., 2016b), Polish (Krapiec and Szychowska-Krapiec, 2016), and Romainian  
869 (Árvai et al., 2016) peatland trees. Peat mining areas where there are likely to be exposed  
870 subfossil trees can be detected easily on aerial photographs and satellite imagery. In some cases,  
871 large wood deposits can also be observed on aerial photos of peat mining areas, but contacts with  
872 peat mining companies are probably the best way of identifying which sites also have subfossil  
873 wood remains.

874

#### 875 *6.2.2. Potential new indicators for detection of moisture changes*

876 Stable isotopes in tree rings have proven to provide palaeoclimatic information with annual  
877 resolution (see Loader and Switsur, 1996; McCarroll and Loader, 2004). Despite this, only two  
878 studies exploring the potential of stable carbon ( $\delta^{13}\text{C}$ ) and oxygen ( $\delta^{18}\text{O}$ ) isotopes have been  
879 performed so far with subfossil peatland tree material (Sass-Klaassen et al., 2005; Edvardsson et  
880 al., 2014a). When considered together, these isotopes have been shown to be sensitive to the  
881 variability in growing season moisture regimes experienced by trees. Based on the correlation  
882 analysis between TRW and isotope series, Edvardsson et al (2014a) demonstrated a three-year lag  
883 between a decline in  $\delta^{13}\text{C}$  and  $\delta^{18}\text{O}$  values in trees and the corresponding tree-ring response,  
884 possibly reflecting a relatively slow rise in local water table in response to wetter climate. Despite

885 the fact that isotopic records generate much additional information compared to TRW records  
886 alone, no further attempts have been performed to date, possibly as these methods are still  
887 expensive and time consuming.

888         Despite the fact that roots will be the first to experience any changes in water-table  
889 fluctuations, root morphology has been used only rarely to detect possible water table changes in  
890 peatland ecosystems (Lageard et al., 1995; Eckstein et al., 2010). For example, a substantial water  
891 table lowering may cause the formation of adventitious roots (Stoffel and Bollschweiler, 2008).  
892 Dendrochronological dating of roots (Stoffel et al., 2013), in combination with field  
893 measurements of their depth and morphological features, may generate detailed information  
894 about the timing and magnitude of hydrological shifts (Eckstein et al., 2010). Such information  
895 cannot be obtained without a detailed assessment of root morphology in the field and the  
896 sampling of second-generation, adventitious roots. We therefore recommend systematic field  
897 measurements and sampling of adventitious roots to test their response to a decline in water table,  
898 and to evaluate whether this can lead to instability of moist surfaces and subsequent formation of  
899 reaction wood (Stoffel and Bollschweiler, 2008) in trees (Fig. 5).

900         Another important aspect of peatland ecosystems and environmental changes is their  
901 response to fire, which can be studied through absolutely dated fire scars in subfossil or living  
902 trees (Arseneault and Payette, 1997; Lageard et al., 2000; Eckstein et al., 2009; Edvardsson et al.,  
903 2012a). The age and frequency of fires may not be sufficient to estimate the loss of burned carbon  
904 from peatlands, but periods showing an increase in fire activity may be indicative of phases with  
905 relatively dry surface conditions. The fire scar records cannot be compared yet, as they do not  
906 originate from continuous records and do not overlap sufficiently in time. Further studies are  
907 therefore required to evaluate if increased fire frequency in tree-ring records does indeed  
908 correspond to periods of relatively dry peatland surface conditions. In addition, fire-scar records



909 can be compared to existing charcoal records, which often represent continuous records and thus  
910 can provide clear evidence of fire frequency and intensity at local to regional scale (Patterson and  
911 Backman, 1988). If combined, the two approaches clearly have the potential to generate long-  
912 term data from charcoal and more detailed (i.e. annual to intra-annual) records for those periods  
913 for which tree-ring data are available.

914 The anaerobic conditions in saturated peat soils generate exceptional preservation  
915 conditions for organic material (Fig. 4). Rapid wet shifts usually cause increased vertical rates of  
916 peat growth, which can be identified in the form of recurrence surfaces in stratigraphic records  
917 (Granlund, 1932; Rundgren, 2008). The same environmental changes will also result in quick  
918 burial of dead peatland trees, and thereby improved preservation conditions for woody material  
919 (Kokkonen, 1923; Leuschner et al., 2007). The degree of post-mortem degradation and stem  
920 height can therefore be used as indicators of the moisture status following tree death. Excellent  
921 preservation and conserved stem bark over distances greater than 30 cm above root level would  
922 indicate fast burial due to moist conditions and rapidly growing peat, whereas conical and eroded  
923 stumps lacking bark have been shown to be characteristic of slow burial and slow vertical peat  
924 growth (Eckstein et al., 2009; Edvardsson et al., 2012a). A guide or strategy for a degradation  
925 index for *in-situ* stumps would therefore be useful and would facilitate evaluation of peat  
926 moisture conditions following dying-off phases.

927 By capitalizing on variable phenologies and tree growth responses to water-table  
928 fluctuations, multispecies reconstructions may provide significantly more information about past  
929 hydroclimatic variability than what we have today. In this context, *Picea*, *Fraxinus*, *Alnus*, *Taxus*  
930 and *Betula* have repeatedly been observed in peat mining areas (Godwin, 1975; Arseneault and  
931 Payette, 1997; Vitas, 2010; Edvardsson et al., 2014b), but to date, they have only rarely been used  
932 as an additional source of information. Of course, the potential of developing multi-millennial

933 chronologies will be limited for many of these species, but they may provide additional  
934 contextual evidence for the longer chronologies (Sass-Klaassen and Hanraets, 2006; Edvardsson  
935 et al., 2014b).

936

### 937 *6.2.3. Carbon storage capacity*

938 The interaction between tree colonization, peatland surface moisture, and net carbon  
939 sequestration is complex (Limpens et al., 2014a; Holmgren et al., 2015). Some authors have  
940 suggested that peatlands may turn from carbon sinks into carbon sources during dry periods  
941 (MacDonald et al., 2006; Yu et al., 2010; Limpens et al., 2011), and that dry surface conditions  
942 will favour tree growth, tree colonization and increased standing biomass in peatland  
943 environments (Freléchoux et al., 2000; Limpens et al., 2014a; Edvardsson et al., 2015b). Phases  
944 with widespread peatland tree colonization may thus also represent periods of less important  
945 peatland carbon accumulation, or even point to phases during which peatlands were indeed acting  
946 as carbon sources. Although tree colonization is likely to increase productivity and litter input,  
947 which may in turn increase peat accumulation rates, total peat accumulation and carbon  
948 sequestration is believed to be reduced as soon as drier conditions start to turn open moss-  
949 dominated peatlands into tree-covered conditions (Limpens et al., 2014a; Holmgren et al., 2015).

950 Further studies are, however, needed on the effects of tree colonization on the drying of  
951 peatland surfaces due to increased water loss through tree evapotranspiration, and to determine  
952 whether such scenarios have the potential to transform peatlands from net carbon sinks into  
953 sources. The effects that tree stand density have on water loss are important and will have  
954 implications for the structure and functioning of peatlands, especially as the relation between  
955 evapotranspiration and tree stand density is non-linear (Limpens et al., 2014a). Systematic  
956 measurements of surface peat moisture, water-table depths, and carbon fluxes to the atmosphere

957 between tree-covered and open peatland areas may give a first indication of the drying effect of  
958 trees. Such studies would need to be performed on natural and managed peatlands representing  
959 different geographical regions and to look into feedbacks and lag effects between climate,  
960 peatland hydrology, and tree growth. To solve some of these remaining issues, we call for  
961 continuous monitoring – over at least some years to decades – of; (i) tree growth in peatlands  
962 with dendrometers, (ii) water table fluctuations, and (iii) carbon fluxes. To assess the overall  
963 carbon balance in peatlands fully, more sophisticated multi-scale approaches will however, be  
964 needed, such as those developed by e.g., Fox et al. (2008) and Hartley et al. (2015), as these will  
965 also take account of the gaseous carbon components (i.e. CO<sub>2</sub> and CH<sub>4</sub> fluxes) as well as of the  
966 dissolved carbon components (i.e. DOC, DIC, and POC). Moreover, such a monitoring setup  
967 would enable calibration and validation of dynamic vegetation and ecosystem models such as the  
968 modified version of LPJ-GUESS (Lund-Potsdam-Jena General Ecosystem Simulator; Smith et  
969 al., 2014) named LPJ-GUESS Why-Me (Wetland Hydrology and Methane, Wania et al., 2009,  
970 2010) and other models that are now including peatland carbon accumulation and methane  
971 emission processes. Given the likelihood of significant shifts in treeline distribution across the  
972 boreal and arctic, improved parameterisation of the interactions between peatland tree growth,  
973 hydrology and carbon cycling is essential. A better understanding of the impact climate change  
974 may have on tree cover, moisture variability and associated changes in the carbon balance in  
975 peatland ecosystems, will advance the possibilities for reliable predictions for future climate  
976 change.

977

978

## 979 **7. Conclusions**

980

981 Contemporary ecosystems can be used to study present-day conditions and to estimate historical  
982 environmental changes, typically over timescales of decades. Palaeoecological techniques  
983 provide much greater temporal depth (Jeffers et al., 2015), and allow assessments over periods  
984 that have experienced significantly different climatic conditions compared to the present day.  
985 Despite significant potential, subfossil wood has been identified as one of the most underutilized  
986 sources of information in palaeoecology and forest conservation research (Lindbladh et al., 2013).  
987 In this review, we have illustrated the origins and recent developments in the field of tree-ring  
988 research in peatland environments, compiled the major datasets from Europe and North America,  
989 but also identified thematic and geographic areas for which knowledge is still fragmentary. We  
990 call for further studies using subfossil peatland trees for detailed reconstructions of long-term  
991 hydroclimate dynamics and palaeoecological changes. Beyond palaeoclimate reconstruction and  
992 palaeoenvironmental understanding, we expect such data to be important in conservation  
993 planning and to determine restoration baselines for peatland management.

994 For more than a century, subfossil peatland trees have been used as a proxy record for  
995 environmental changes (Dau 1829; Vaupell 1851; Nielsen and Helama, 2012). Over the last  
996 decades, however, numerous studies have started to use the information contained in the annually  
997 resolved tree-ring records in more detail to assess climatic changes during the Holocene in the  
998 British Isles, Scandinavia, and throughout Central Europe (Table 1, Fig. 1). Through the  
999 comparison between tree replication records, tree-ring width chronologies, climate and moisture  
1000 proxies (i.e. ice cores, sediment records, pollen), as well as meteorological data, detection of  
1001 climatic signals in peatland trees has been improved substantially. We conclude that:

1002 (1) Our current knowledge of the effects of (hydro-)climatic changes and anthropogenic  
1003 activities on tree growth in peatland ecosystems is mainly based on ecophysiological,  
1004 dendroclimatic, and dendroecological perspectives;

- 1005 (2) A majority of existing studies shows that excess moisture is a growth-limiting factor for  
1006 peatland trees. Proxy records developed from subfossil peatland trees therefore have the  
1007 potential to help in the reconstruction of (i) local water-table fluctuations, and (ii) regional  
1008 hydroclimate dynamics. Annual tree-growth patterns can be used to assess changes at  
1009 annual to decadal scales, whereas tree population dynamics reflect changes at decadal to  
1010 centennial time scales;
- 1011 (3) Substantial work has been performed using subfossil trees from Germany, Great Britain,  
1012 Ireland, the Netherlands, and parts of Scandinavia, whereas the vast majority of Canada,  
1013 Russia, and the Baltic countries have significant under-exploited potential for the  
1014 development of geographically widespread records;
- 1015 (4) Our data compilation showing temporal distribution of subfossil peatland forests confirms  
1016 that relatively warm and dry periods – such as the HTM – offered conditions which were  
1017 favourable for tree growth in peatlands and thereby allowed widespread tree colonization,  
1018 whereas transitions towards moister and colder conditions (e.g., the mid-late Holocene  
1019 transition and the LIA) are associated with widespread dying-off phases and/or depressed  
1020 annual tree growth;
- 1021 (5) Further studies using e.g., stable isotopes, dendroclimatology, and field monitoring are  
1022 critical to increase our knowledge about tree growth responses associated with  
1023 hydroclimate changes, including understanding of hydrological lag and feedback effects  
1024 detectable from lagging tree growth responses to climate on peatlands.

1025

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1031

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1797 **Figure Caption**

1798

1799 Fig. 1. (a) Location of the main peatlands referred to in Tables 1 and 2 (Id. No. 1-81). Sites from  
1800 which subfossil trees have been sampled are shown by black squares while sites from which  
1801 living trees have been sampled are shown by red triangles. (b) Map over peatlands outside  
1802 northwest Europe. Peat coverage is based on Montanarella et al. (2006).

1803

1804 Fig. 2. (a) During the early Holocene, many topographic depressions were occupied by lakes in  
1805 which organic layers consisting of e.g. gyttja (1) was deposited. In the shallow areas along the  
1806 shores, reed or fen peat was developed (2). (b) During the mid-Holocene, relatively warm and dry  
1807 conditions allowed trees to colonize many peatlands. (c) Accelerating peat growth following wet-  
1808 shifts can often be observed as recurrence surfaces (RS) in stratigraphic records. Such  
1809 hydrological changes may cause tree dying-off phases. Stumps and trunks from these trees can be  
1810 preserved by the growing peat (3). (d) At present peatlands are valuable climate archives. Peat  
1811 stratigraphic records (4) generate long-term information showing environmental changes,  
1812 whereas tree-ring chronologies (5) can generate annually resolved information. The tree-ring  
1813 chronologies (example from Edvardsson, 2016) can be developed from both living and subfossil  
1814 trees (6). Tree growth is usually influenced by hydrological variations (7) related to climate (8).  
1815 Moreover, climate, hydrology, and tree growth ultimately influence the carbon budget (9) of the  
1816 peatlands.

1817

1818 Fig. 3. (a) The natural raised bog Fjällmossen (S. Sweden) is characterized by scattered  
1819 populations of birch and pine trees. (b) Open water ponds at the raised bog Aukštumala  
1820 (Lithuania) show how close the water table is to bog surface and the root systems of the pine

1821 trees. (c) Peatland pines growing at Saxnäs Mosse, Sweden. (d) A till hummock in  
1822 Hästhultsmossen (S. Sweden) generates stable and dry conditions favoring colonization of trees.  
1823 The area surrounding the mineral soil hummock, however, is relatively wet, unstable and  
1824 supports significantly smaller pine trees even though the trees are of same age.

1825  
1826 Fig. 4. (a) Several meters of peat have been harvested at the Rėkyva peat mining area (Lithuania)  
1827 with the result that subfossil trees preserved in peat have been exposed. (b) Tree horizon at the  
1828 Viss Mosse peatland (Sweden; Fig. 1, site 23) containing *in situ* pine stumps. These trees were  
1829 growing at the bog surface between 7200 and 6500 years BP. (c) Field sampling of a c. 6000-  
1830 years old pine tree at Hällarydsmossen (Sweden; Fig. 1, site 22b). (d) To enable continued peat  
1831 mining, tree trunks and stumps are often stored in piles surrounding the peat mining area  
1832 (Hällarydsmossen).

1833  
1834 Fig. 5. Cross-sections and annual growth rings from subfossil trees exhibit different growth  
1835 patterns and characteristics. (a) Two cross-sections from Scots pine (*Pinus sylvestris* L.)  
1836 excavated at Viss Mosse (Fig. 1, site 23). The left hand sample contains a complete ring sequence  
1837 from pith to bark corresponding to 5078–4970 BCE (7027-6919 BP). The right hand sample has  
1838 degraded after death. The growth rings correspond to the period 4872–4716 BCE (6821-6665  
1839 BP), but approximately 50–70 rings towards the center and another 10–20 of the outermost rings  
1840 are expected but have eroded away. (b) Fire scar dated to 4420 BCE (6369 BP). (c) Abrupt  
1841 growth depression visible in a pine tree from the peatland Åbuamossen (Fig. 1, site 24). The  
1842 growth depression most likely resulted from a period of extremely wet conditions between 1288  
1843 and 1270 BCE (3237-3219 BP) and is visible in all dated trees growing at the peatland. (d) Tree-  
1844 ring sequence from a subfossil oak (*Quercus robur* L.). (e) Eccentric pine-stem cross-section



1845 characteristic for trees growing on unstable soils. The innermost 70 annual rings are circular,  
1846 whereas about 4070 BCE (6020 BP) unstable conditions are evident having forced the tree to  
1847 form reaction wood. (f) About 70 narrow annual rings (inside the circle) around the pith,  
1848 indicating unfavourable conditions during establishment.

1849  
1850 Fig. 6. Significant cross-dating statistics between continental and Irish peatland and oak  
1851 chronologies have been described by Pilcher et al. (1984) and Leuschner et al. (2002; arrow a).  
1852 German peatland-oaks also cross-date oaks from River Main (Friedrich et al., 2004, arrow b), the  
1853 Netherlands (Sass-Klaassen and Hanraets, 2006; arrow c), Denmark (unpublished, see Fig. 7;  
1854 arrow d), and Swedish pine (Edvardsson, 2013; arrow e), whereas the Irish chronologies (oak and  
1855 pine) cross-date with the material from Great Britain and Scotland (e.g. Boswik and Woodhouse,  
1856 2002; arrow f). Some of the Danish and Swedish oak chronologies also cross-date (Edvardsson et  
1857 al., 2014b; arrow g). Furthermore, the Swedish and German pine chronologies generated strong  
1858 cross-dating statistics (Edvardsson et al., 2012a; arrow e). Attempts to cross-date Baltic pine  
1859 chronologies using material from Germany and Sweden have, however, failed (Edvardsson et al.,  
1860 2016a; dashed arrows h). Tests to cross-date south Swedish pine chronologies with peatland pines  
1861 described by Gunnarson (1999; dashed arrow i) and subfossil mineral soil pines from Torneträsk  
1862 (Grudd et al., 2002; dashed arrow j) failed as well. By contrast, the north Swedish and Finnish  
1863 chronologies (Helama et al., 2005; arrow k) indeed cross-date well. Peat coverage is based on  
1864 Montanarella et al. (2006).

1865  
1866 Fig. 7. Stepwise cross-dating of tree-ring chronologies from northwest Europe. The t-values  
1867 between different chronologies are presented next to the arrows. Oak chronologies are shown in  
1868 black, pine in grey, and dashed lines represent historical material anchored in present. The figures

1869 to the right correspond to the site IDs in Fig. 1 and Table 1. The River Main oak chronology (\*,  
1870 Friedrich et al., 2004) is also included in the figure.

1871  
1872 Fig. 8. Comparisons between tree-ring width (TRW) and meteorological records (monthly  
1873 temperature and precipitation data). Sites from which both tree growth at mineral soils (MS) and  
1874 peat soils (PS) could be compared are presented. The identification numbers (Id. No.)  
1875 corresponds to those presented in Fig. 1 and Table 2.

1876  
1877 Fig. 9. (a–b) Examples of TRW chronologies from living trees growing at Aukštumala and  
1878 Rėkyva (Edvardsson et al., 2015a). (c) Precipitation data from the meteorological station in  
1879 Vilnius. (d) TRW chronologies from Buxabygds Mosse, Saxnäs Mosse and Store Mosse, Sweden  
1880 (Edvardsson and Hansson, 2015). (e) Local precipitation data and (f) river discharge data. The  
1881 blue bars emphasize moist periods associated with positive precipitation anomalies, grey lines  
1882 represent average values. Growth depressions or trends towards narrower growth rings can often  
1883 be observed during phases associated with positive precipitation/wet anomalies.

1884  
1885 Fig. 10. Main data series compared in the discussion, oak in black, pine in green, and numbers in  
1886 brackets correspond to those in Table 1. (a) Irish data series, in black Irish oak replication, in  
1887 green pine establishments, and cumulative radiocarbon dated pine samples (green/black squares).  
1888 (b) Cumulative weighted Scottish peatland trees dated by radiocarbon and coverage of peatland  
1889 pines chronologies. (c) German peatland oaks. (d) Continental and Irish oak mean-age records.  
1890 (e) German pine replication and woodland phases. (f) Replication and temporal spread of oak  
1891 from the Netherlands. (g) Periods covered by Danish peatland oaks. (h) Swedish pine replication

1892 and periods covered by oak chronologies. (i) Polish pine chronologies. (j) Lithuanian pine  
1893 replication as well as pine and oak coverage. (k) Pine replication from Finland. (l) Spruce  
1894 coverage from Canada. (m) Dry and wet shifts highlighted in yellow and blue respectively  
1895 (Wanner et al., 2011). (n) Wet shifts detected in peat bogs from northern Britain (Charman,  
1896 2006). (o) Lake levels in the Baltic region (Sohar and Kalm, 2008). (p) High lake-level phases in  
1897 mid-European lakes (Magny, 2004, 2013). (q) Wet shifts detected as recurrence surfaces in  
1898 Swedish peat bogs (Borgmark and Wastegård, 2008). (r) In blue, temperature anomalies in  
1899 relation to the average (Marcott et al., 2013) and in red, a pollen based temperature reconstruction  
1900 by Heikkilä and Seppä (2008).

1901 Fig. 11. Correlation tests between dimensionless TRW indices (grey) from the peatland  
1902 Čepkeliai, Lithuania and (a) water-table fluctuations (cm) in relation to the average level, (b) total  
1903 annual precipitation (mm), and (c) annual mean temperature (°C).

**Table 1.** Representation of the main studies used in this review for comparison. Study areas include Austria (AU), Canada (CA), Denmark (DK), Finland (FI), France (FR), Germany (DE), Great Britain (GB), Ireland (IE), Italy (IT), Lithuania (LT), the Netherlands (NL), Poland (PL), Sweden (SE), and Switzerland (CH). Tree species presented are Alder (A), Ash (F), Larch (La), Oak (Q), Pine (P), and Spruce (Pi), and what context the material originates from, bogs/peatlands (b), mainly bogs/peatlands (m) and partly bogs/peatlands (p).

<b>Id. No. and References</b>	<b>Country / Latitude, Longitude</b>	<b>Species / (b, m, p)</b>	<b>Main time span</b>
1. Achterberg et al. (2015)	DE / 52°50' N, 9°50' E *	P (b)	4629 BCE - CE 502
2. Achterberg et al. (2016)	DE / 52°50' N, 9°50' E *	P, Q (b)	6703 BCE - CE 931
3. Arseneaut & Payette (1997)	CA / 57°44' N, 76°10' W	Pi (b)	500 BCE - present
4. Baillie (1973)	IE / 54°50' N, 6°50' W *	Q (b)	5000-1380 CE
5. Baillie (1977)	IE / 54°50' N, 6°50' W *	Q (b)	5000-1001 CE
6. Baillie et al. (1983)	IE / 54°50' N, 6°50' W *	Q (b)	5000 – 1000 BCE
7. Baillie & Brown (1988)	IE / 54°50' N, 6°50' W *	Q (b)	5300 – 200 BCE
8. Barniak et al. (2014)	PL / 54°15' N, 19°44' E	P (b)	940-460 BCE **
9. Bauerochse et al. (2015)	DE / 52°50' N, 9°50' E *	P, Q (b)	2600-1500 BCE
10. Boswick & Whitehouse (2002)	GB / 53°32' N, 0°57' W	P (b)	2921–2445 BCE
11. Bridge et al. (1990)	GB / 56°52' N, 4°32' W *	P (b)	5900-1200 BCE ***
12. Brown et al. (1986)	IE / 54°50' N, 6°50' W *	Q (b)	5300-200 BCE
13. Brown & Baillie (1992)	IE / 54°50' N, 6°50' W *	P, Q (b)	3600-2600 BCE
14. Brown & Baillie (2012)	IE / 54°50' N, 6°50' W *	Q (m)	5200 BCE – present
15. Chambers et al. (1997)	GB / 53°40' N, 0°53' W	P (b)	2900-2400 BCE
16. Charman (2010)	IE / 54°60' N, 6°40' W *	Q (b)	5000 BCE - present
17. Christensen (2007)	DK / 56°30' N, 9°50' E *	Q (m)	6000 – 200 BCE
18. Delorme et al. (1981)	DE / 52°50' N, 9°50' E *	Q (b)	350-150 BCE **
19. Eckstein et al. (2009)	DE / 52°50' N, 9°50' E *	P (b)	5600-2200 BCE
20. Eckstein et al. (2010)	DE / 52°44' N, 8°17' E	P, Q (b)	2450-2050 BCE
21. Eckstein et al. (2011)	DE / 52°70' N, 9°00' E *	P (b)	6500-1200 BCE
22. Edvardsson et al. (2012a)	SE / 55°51' N, 13°49' E	P (b)	5219–4574 BCE
	SE / 57°20' N, 14°35' E	P (b)	4839–3728 BCE
23. Edvardsson et al. (2012b)	SE / 55°51' N, 13°49' E	P (b)	6100-4600 BCE
24. Edvardsson et al (2014a)	SE / 56°00' N, 14°00' E *	P (b)	4540-4460, 1320-1240 BCE
25. Edvardsson et al. (2014b)	SE / 55°51' N, 13°49' E	P, Q, A, F (b)	4500-1000 BCE **
26. Edvardsson (2016)	SE / 56°19' N, 13°55' E	P (b)	2667-1108 BCE
27. Edvardsson et al. (2016a)	LT / 55°30' N, 23°50' E *	P (b)	4000 BCE – CE 900 *
28. Gear & Huntley (1991)	GB / 58°24' N, 4°3' W	P (b)	2450-1900 BCE **
29. Grudd et al. (2000)	SE / 59°08' N, 17°55' E	P (b)	1700-1500 BCE **
30. Gunnarson (1999)	SE / 59°08' N, 17°55' E	P (b)	1686-1483 ± 23 BCE **
31. Gunnarson (2008)	SE / 63°16' N, 12°31' E	P (p)	5000-2000 BCE **
32. Hall et al. (1994)	IE / 55°05' N, 6°34' W	P, Q (b)	6000-0 BCE **
33. Helama et al. (2004)	FI / 67°50' N, 26°25' E *	P (p)	5600 BCE - present
34. Jansma (1996)	NL / 50°83' N, 5°57' E	Q (p)	2258-1141 BCE
35. Lageard et al. (1999)	GB / 53°40' N, 2°20' W *	P, Q (b)	3520-2462 BCE **, 3228-1891 BCE
36. Lageard et al. (2000)	GB / 53°40' N, 2°20' W *	P (b)	2881-2639 BCE
37. Leuschner et al. (1987)	DE / 52°40' N, 8°20' E *	Q (b)	6000 BCE-500 CE
38. Leuschner & Delorme (1988)	DE / 52°40' N, 8°20' E *	Q (m)	6225 BCE - present
39. Leuschner (1992)	DE / 52°40' N, 8°20' E *	Q (b)	6000 BCE-500 CE
40. Leuschner et al. (2000)	DE, NL *	Q (b)	6000 BCE-500 CE (GE), 6000-4500 BCE, 3000 BCE-500 CE
41. Leuschner et al. (2002)	DE, IE, NL *	Q (b)	6069 BCE-1596 CE
42. Leuschner et al. (2007)	DE / 52°27' N, 8°10' E	P (b)	3050-2700 BCE
43. McGeever & Mitchell (2015)	IE / 53°00' N, 7°00' W *	P (b)	6000 BCE-1000CE ***
44. Moir et al. (2010)	GB / 58°90' N, 4°30' W *	P (b)	3200-3000 BCE
45. Moir (2005, 2012)	GB / 58°90' N, 4°30' W *	P (b)	3200-2800 BCE
46. Munaut (1967)	NL / 52°60' N, 6°30' E	P (b)	5000, 4000, 3000 BCE **
47. Munaut & Casparie (1971)	NL / 52°60' N, 6°30' E	P (b)	5000, 4000, 3000 BCE **

48. Nicolussi et al. (2005)	AU / 46°90' N, 10°70' E *	P (p)	7100-2100BCE
49. Nicolussi et al. (2009)	AU, IT, CH *	P, Pi, La (p)	7100BCE - present
50. Pilcher et al. (1973)	IE / 54°50'N, 6°50' W *	Q, P (b)	6000 BCE - present
51. Pilcher et al. (1977)	IE / 54°50'N, 6°50' W *	Q (b)	4000-1000 BCE
52. Pilcher et al. (1984)	IE, GB, DE *	Q (m)	5218-158 BCE (present)
53. Pilcher et al. (1995)	IE / 54°60' N, 5°90' W *	P (b)	3451-2569 BCE
54. Pilcher et al. (1996)	IE / 54°60' N, 6°50' W *	Q (m)	5218-158 BCE
55. Pukienė (1997, 2001)	LT / 56°05' N, 21°50' E	P (b)	200 BCE-present**
56. Pukienė (2003)	LT / 56°12' N, 24°44' E	Q (b)	4600-4400 BCE**
57. Sass-Klassen et al. (2005)	NL / 52°31' N, 6°02' E	Q (b)	200 BCE-150 CE
58. Sass-Klassen & Hanraets (2006)	NL / 52°31' N, 6°02' E	Q, F (b)	600 BCE-200 CE
59. Simard et al. (2011)	CA / 48° N, 71° W *	Pi (b)	3200 BCE **
60. Tessier et al. (1993)	FR / 45°05' N, 5°92' E	P (p)	6000 BCE - present **
61. Torbenson et al. (2015)	IR / 54°49' N, 6°30' W *	P (b)	6300-5600 BCE
62. Turney et al. (2005)	IE / 54°60' N, 6°40' W *	Q, P (m)	5000 BCE - present
63. Vitas (2009)	LT/ 55°50' N, 22°17' E	Q, F (b)	4700-1500 BCE **
64. Vitas (2010)	LT/ 55°48' N, 21°53' E	F, A, Q, B (b)	3800-1290 BCE **
65. Ward et al. (1987)	GB / 56°40' N, 4°18' W	P (b)	4500-1000 BCE **

\*Approximate coordinates as the chronology/chronologies are based on material from several study sites

\*\*Approximate ages as the chronology/chronologies are dated by radiocarbon

\*\*\*Approximate ages as of individual subfossil trees (no tree-ring chronologies) dated by radiocarbon

**Table 2.** Papers in which comparative studies based on peat soil and mineral soil trees have been performed. For further information, see Fig 1 and 8. Study areas include Great Britain (GB), Lithuania (LT), Poland (PL), Russia (RU), and Sweden (SE). Tree species included are *Pinus sylvestris* (PiSy) and *Pinus sibirica* (PiSi).

<b>Id. No. and References</b>	<b>Country / Site name</b>	<b>Latitude, Longitude</b>	<b>Species</b>
66. Blanchet et al. (2016)	RU / Mukhrino	60°54'N, 68°42'E	PiSy, PiSi
67. Cedro & Lamentowicz (2011)	PO / Słowinskie Błoto	54°21'N, 16°29'E	PiSy
68. Edvardsson et al. (2015a, b)	LT / Aukštumala	55°23'N, 21°22'E	PiSy
69. Edvardsson et al. (2015a, b)	LT / Kerėplis	54°27'N, 24°32'E	PiSy
70. Edvardsson et al. (2015a, b)	LT / Rėkyva	55°51'N, 23°15'E	PiSy
71. Edvardsson & Hansson (2015)	SE / Store Mosse	57°14'N, 13°55'E	PiSy
72. Linderholm (2001)	SE / Lake Ånn	63°15'N, 12°30'E	PiSy
73. Linderholm et al. (2002)	SE / Augerusjärvi	67°38'N, 21°46'E	PiSy
74. Linderholm et al. (2002)	SE / Stortjäderberget	54°32'N, 19°00'E	PiSy
75. Linderholm et al. (2002)	SE / Årsön	63°15'N, 12°30'E	PiSy
76. Linderholm et al. (2002)	SE / Bredmossen	60°08'N, 16°05'E	PiSy
77. Linderholm et al. (2002)	SE / Hanvedsmossen	59°09'N, 17°55'E	PiSy
78. Linderholm et al. (2002)	SE / Anebymossen	57°51'N, 14°38'E	PiSy
79. Moir et al. (2011)	GB / Eilean Subhainn	58°41'N, 5°29'W	PiSy
80. Moir et al. (2011)	GB / Abernethy	57°14'N, 3°40'E	PiSy
81. Edvardsson (unpublished)	LT / Čepkeliai	54°07'N, 24°30'E	PiSy

Fig. 1

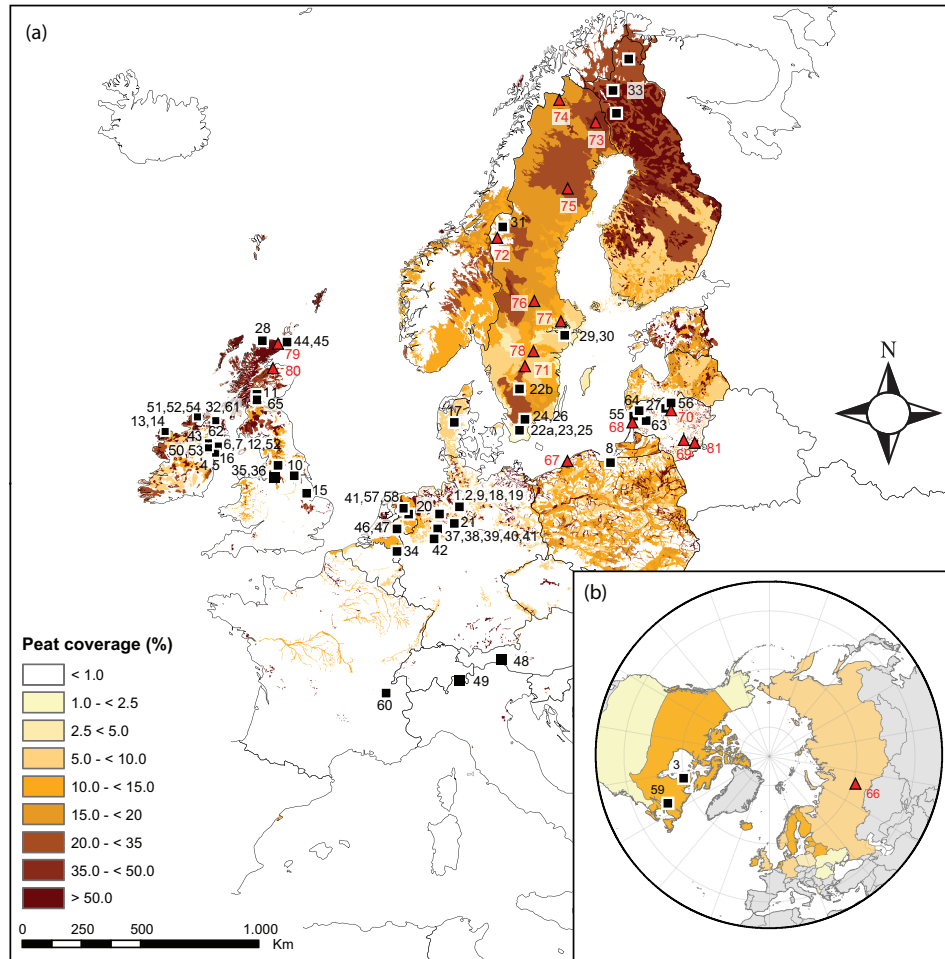


Fig. 2

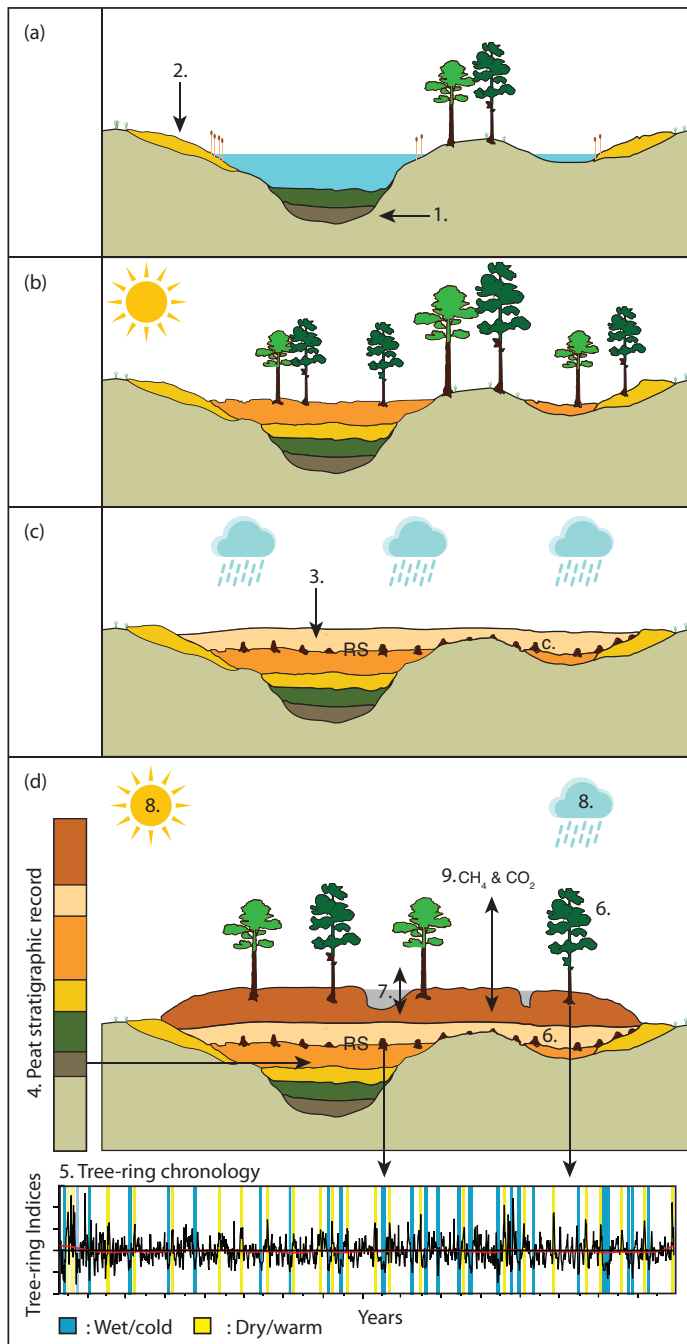


Fig. 3

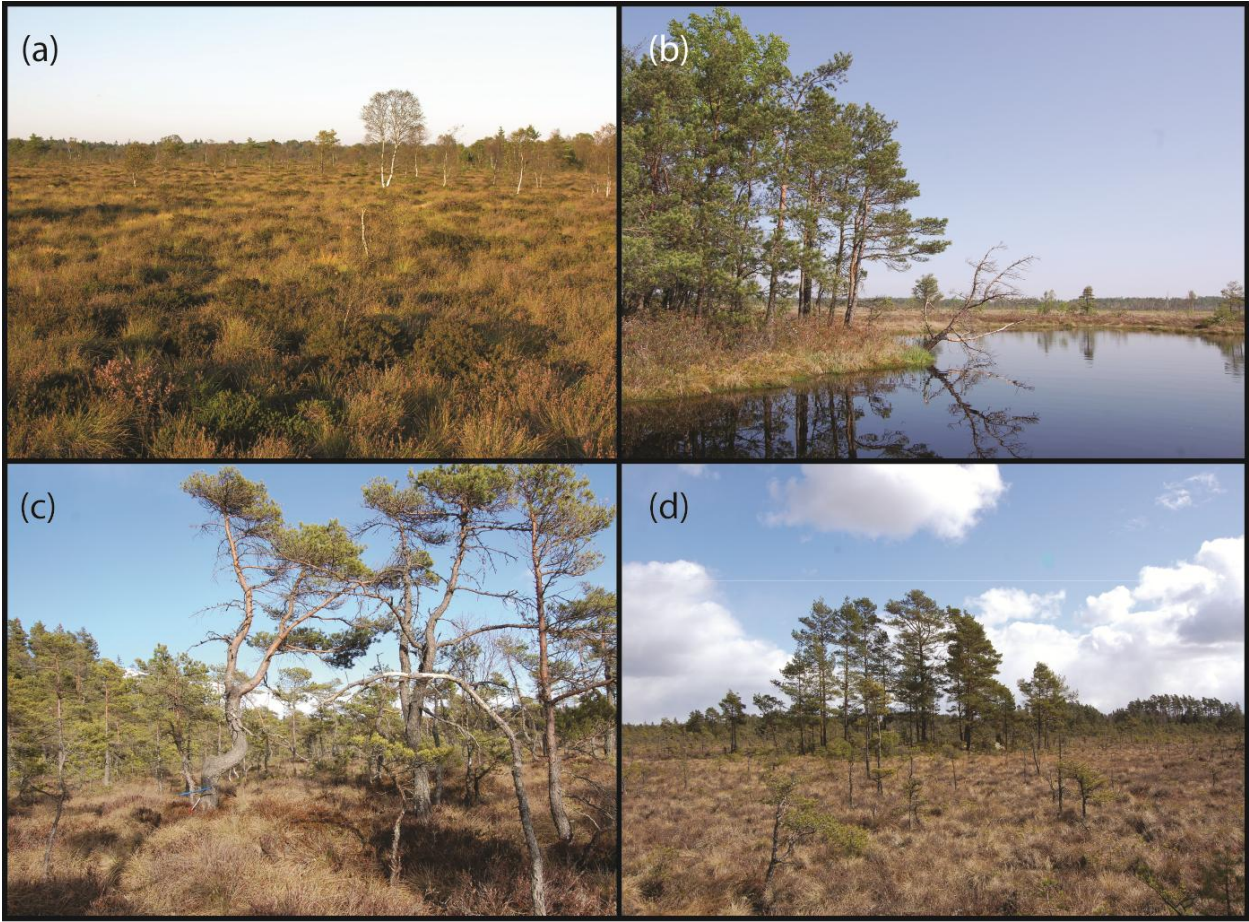




Fig. 4

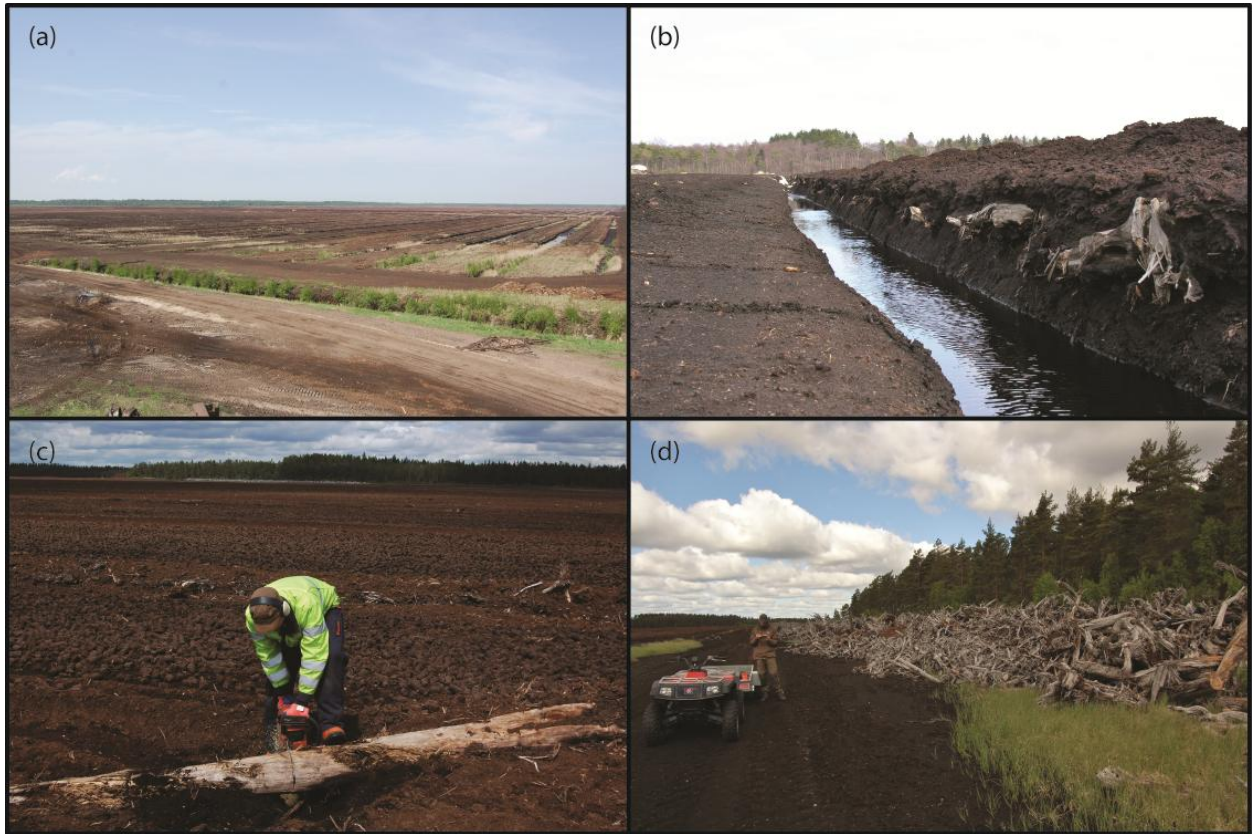


Fig. 5

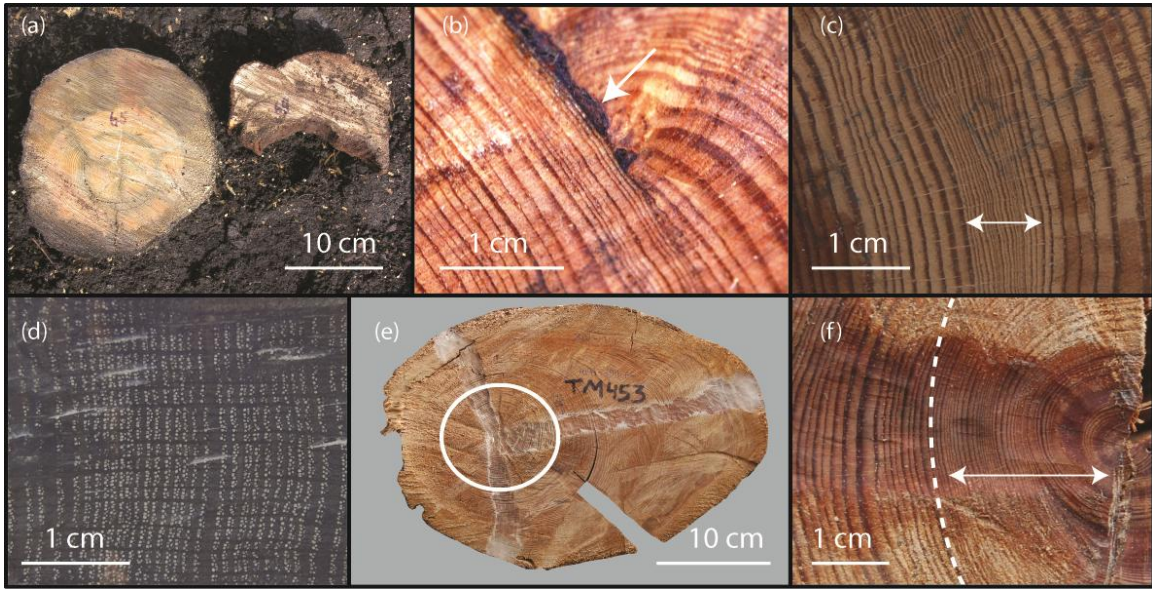


Fig. 6

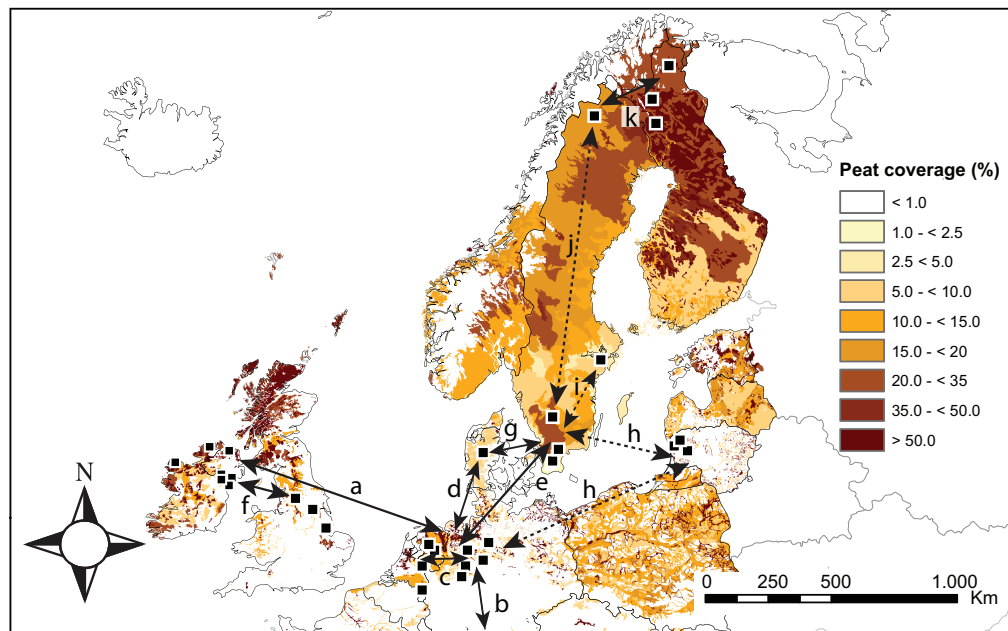


Fig. 7



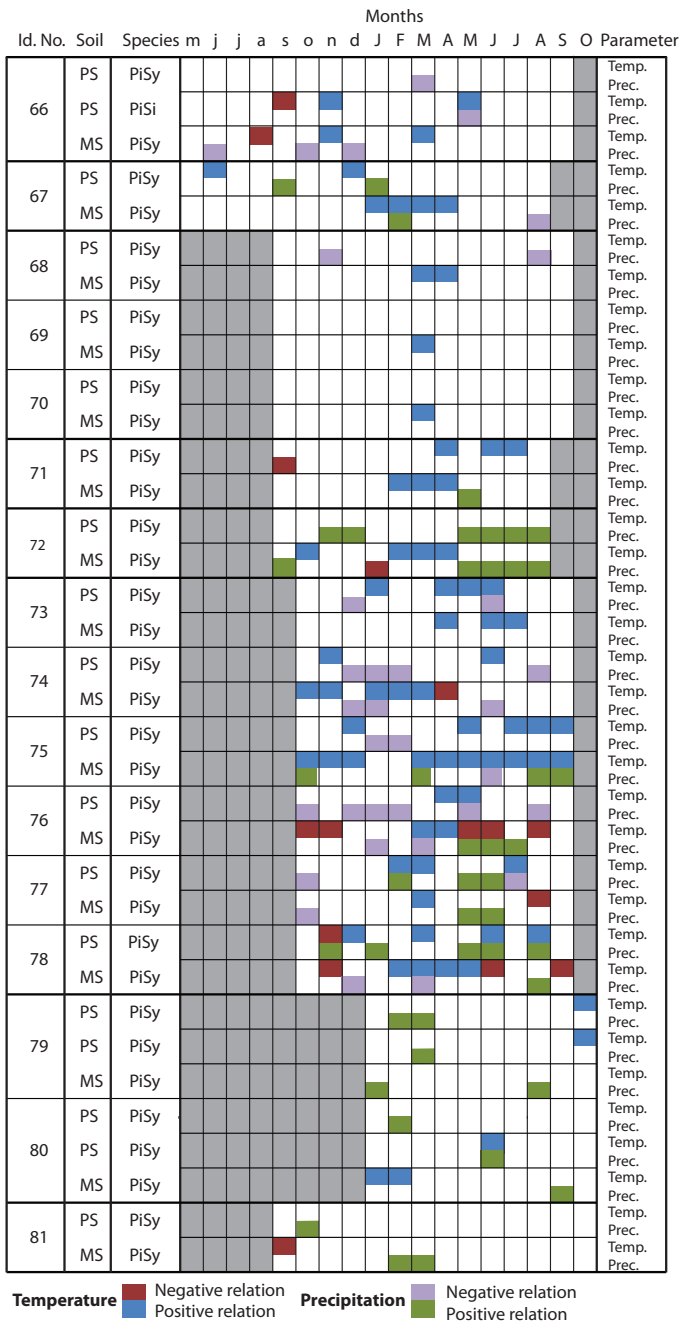


Fig. 9

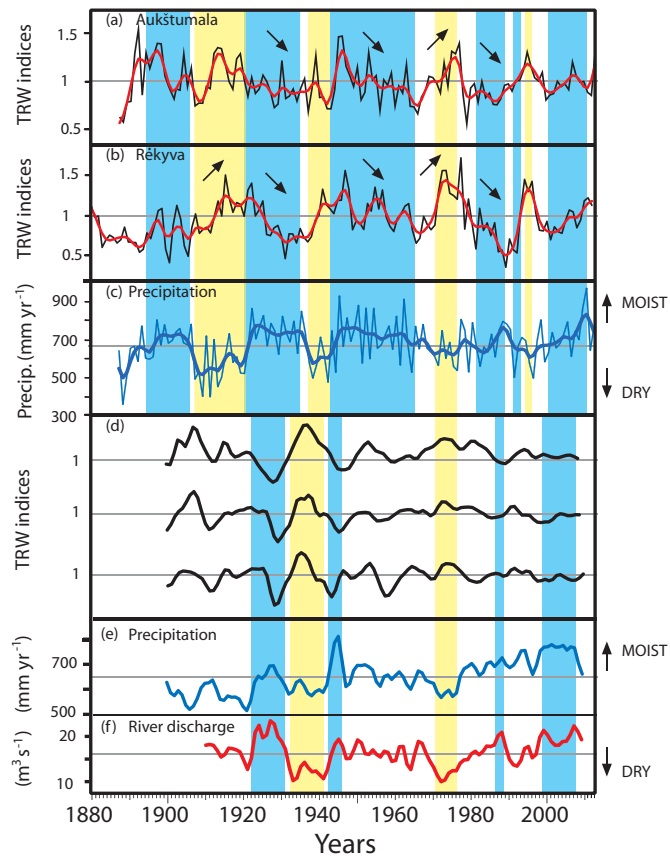


Fig. 10

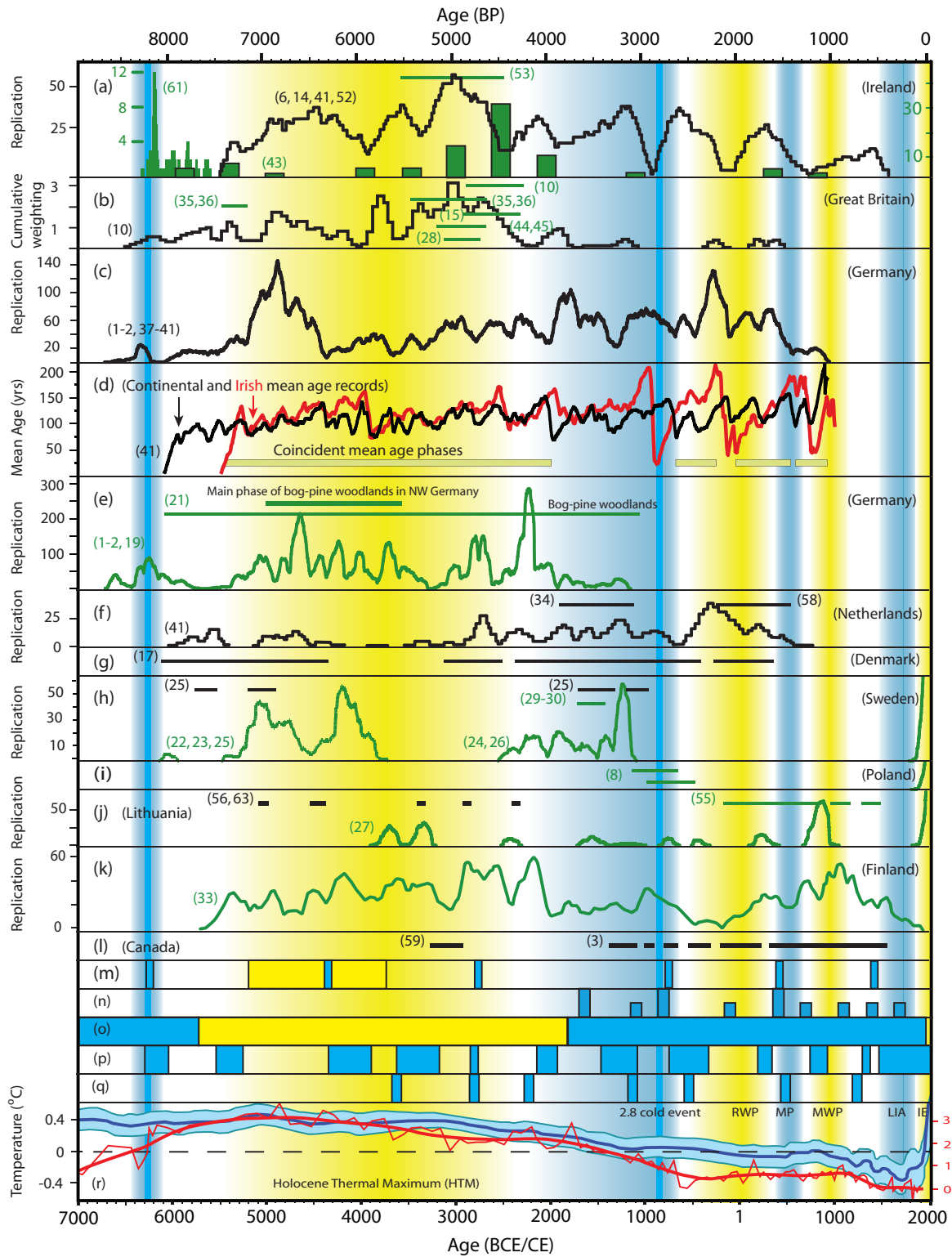


Fig. 11

