

This document is confidential and is proprietary to the American Chemical Society and its authors. Do not copy or disclose without written permission. If you have received this item in error, notify the sender and delete all copies.

**Reliable, Resilient and Sustainable Urban Drainage Systems:
an Analysis of Robustness under Deep Uncertainty**

Journal:	<i>Environmental Science & Technology</i>
Manuscript ID	es-2018-011939.R1
Manuscript Type:	Policy Analysis
Date Submitted by the Author:	19-May-2018
Complete List of Authors:	Casal-Campos, Arturo; Ofwat Sadr, Seyed; University of Exeter College of Engineering Mathematics and Physical Sciences Fu, Guangtao; University of Exeter, Engineering Butler, David; University of Exeter, Centre for Water Systems

SCHOLARONE™
Manuscripts

1 Reliable, Resilient and Sustainable Urban Drainage

2 Systems: an Analysis of Robustness under Deep

3 Uncertainty

4 *Arturo Casal-Campos^{†*}, Seyed M. K. Sadr[†], Guangtao Fu^{†*} and David Butler[†]*

[†]Centre for Water Systems, College of Engineering, Mathematics and Physical Sciences, University of
Exeter, North Park Road, Harrison Building, Exeter, EX4 4QF, United Kingdom

5 * Corresponding authors: Casal-Campos, A.: a.casal@mail.com, Fu, G.: g.fu@exeter.ac.uk

7 **ABSTRACT**

8 Reliability, resilience and sustainability are key goals of any urban drainage system. However, only a few
9 studies have recently focused on measuring, operationalizing and comparing such concepts in a world of
10 deep uncertainty. In this study, these key concepts are defined and quantified for a number of gray, green
11 and hybrid strategies, aimed at improving the capacity issues of an existing integrated urban wastewater
12 system. These interventions are investigated by means of a regret-based approach, which evaluates the
13 robustness (that is the ability to perform well under deep uncertainty conditions) of each strategy in terms
14 of the three qualities through integration of multiple objectives (i.e. sewer flooding, river water quality,
15 combined sewer overflows, river flooding, greenhouse gas emissions, cost and acceptability) across four
16 different future scenarios. The results indicate that strategies found to be robust in terms of sustainability
17 were typically also robust for resilience and reliability across future scenarios. However, strategies found

18 to be robust in terms of their resilience and, in particular, for reliability did not guarantee robustness for
19 sustainability. Conventional gray infrastructure strategies were found to lack robustness in terms of
20 sustainability due to their unbalanced economic, environmental and social performance. Such limitations
21 were overcome, however, by implementing hybrid solutions that combine green retrofits and gray
22 rehabilitation solutions.

23 **1. INTRODUCTION**

24 Emerging threats affecting urban areas now and in the future may significantly contribute to the
25 deterioration of the level of service delivered by critical infrastructure, such as urban drainage systems (or
26 urban wastewater systems). Indeed, climate change, population growth, urbanization, and other changing
27 factors could be particularly damaging when simultaneously acting upon any of these systems, posing an
28 important challenge to their future performance.^{1,2} In addition to this, the deep uncertainty nature of future
29 conditions may question the robustness of conventional and alternative solutions to adapt to future
30 changes, given their unknown magnitude and extent of impacts over the long-term.³

31 Under these circumstances, urban wastewater infrastructure may need to undergo adaptive
32 improvements in order to become less vulnerable to future conditions, whether these are typical or
33 extraordinary.⁴ Indeed, it is expected that the urban wastewater system is reliable, able to minimize failure
34 frequency and deliver a satisfactory level of service most of the time, while behaving resiliently in order
35 to reduce the duration and magnitude of a failure when this eventually happens.⁵

36 At the same time, these adapted systems should also pursue sustainability in the long-term, i.e. to
37 achieve economic, environmental and social goals altogether. However, adaptation strategies that provide
38 a high level of technical performance (i.e. are reliable and resilient) may not necessarily be financially
39 viable, environmentally balanced (e.g. protecting the aquatic environment at the expense of other
40 environmental issues), or socially equitable.⁶

41 Understanding the attributes and existing relationships between these operational (reliability and
42 resilience) and strategic (sustainability) qualities thus becomes paramount in devising strategies likely to

43 be successful now and in the future. In spite of revived efforts to understand these attributes (in particular
44 from a resilience point of view)⁶ in the context of urban wastewater systems,⁷⁻¹⁰ there currently exists
45 limited knowledge regarding the extent and qualities of such relationships, especially as related to their
46 metrics and methods of assessment. Consequently, it is largely uncertain how to ensure that any drainage
47 strategy can maintain its reliability, resilience or sustainability qualities over time and in the face of
48 changing conditions. Moreover, it is unclear what type of solutions (e.g. centralized or decentralized
49 alternatives) are capable of delivering reliable, resilient and sustainable outcomes now and in the
50 future.¹¹⁻¹³

51 In addition to this, the need to satisfy an ever increasing variety of objectives, whether these are related
52 to economic (e.g. capital and operational costs), social (e.g. acceptability, equity) or environmental (e.g.
53 water quality, carbon emissions) drivers, may further challenge our aspirations to plan for robust solutions
54 that satisfy levels of service under a wide range of changing circumstances.

55 The aim of this paper is therefore to analyze, quantify and compare the robustness of urban drainage
56 system enhancement strategies in terms of reliability, resilience and sustainability when subject to
57 uncertain future changes (e.g. climate change, population growth). Such analysis is carried out using a
58 regret-based approach that relatively assesses the multi-objective performance metrics (reliability,
59 resilience and sustainability) of conventional gray infrastructure strategies and green infrastructure
60 retrofits across four future scenarios. The present study builds on the work developed by Casal-Campos et
61 al.,¹⁴ which developed a regret-based approach to compare the relative performance of green and gray
62 strategies on an integrated catchment. This paper applies the same regret-based method to investigate the
63 robustness of green, gray and hybrid strategies in delivering reliable, resilient and sustainable wastewater
64 services in the future. The outputs of the research provide further insight into the ability of adaptive
65 wastewater management policies to ensure enhanced levels of performance in the future.

66

67 2. MATERIALS AND METHODS

68 2.1. Overview

69 Each strategic intervention applied to the case study is assessed under four different future scenarios in
70 terms of its relative performance regarding multiple objectives (sewer and river flooding, river water
71 quality, operational Greenhouse Gas (GHG) emissions, Combined Sewer Overflow (CSO) spills, whole-
72 life costs and acceptability). A brief description of the model, scenarios and interventions is provided
73 below; the reader should refer to Casal-Campos et al.¹⁴ for further details.

74

75 2.2. Case Study and Future Scenarios

76 2.2.1. Case Study Overview

77 The urban wastewater catchment employed for this investigation is a semi-hypothetical benchmark case
78 originally defined by Schütze et al.¹⁵. The integrated case study consists of three main sub-systems,
79 namely: urban catchment and sewer system, wastewater treatment plant system (WWTP), and river
80 system; see the Supporting Information (SI), **Figure S1**. The catchment system is defined by 15
81 individual sub-catchments, served by a simplified combined sewer network (main trunk sewers with 1.2
82 meters diameter pipes). The excess flows forwarded from the sewer network are stored in a storm tank
83 (off-line pass-through tank, 6750 m³), overflowing to the river system. The wastewater flow entering the
84 treatment process follows a typical activated sludge arrangement. The river defined for the case study is a
85 hypothetical 40-km river divided in 40 equal stretches. The river base flow is 1.5 m³/s (129,600 m³/d),
86 which results in a 1:5 dilution factor of dry-weather treatment plant discharges to the river.

87 The catchment is modeled as an integrated urban wastewater system (IUWWS) using SIMBA 6.0.¹⁶
88 This is a modeling tool based on the MATLAB/SIMULINK environment, which allows users to construct
89 and develop specific modeling modules tailored to their needs (e.g. wastewater treatment processes,
90 elements of the catchment and sewer network, etc.) This included using SWMM 5 to model the sewer
91 network and river systems, as well as IWA's state-of-the-art Activated Sludge Model No.1 (ASM1) to

92 model the WWTP. The original integrated model has been extensively used to report on the benefits of
93 integrated real time control strategies to improve river water quality (i.e. dissolved oxygen and ammonia)
94 through control of diurnal patterns of WWTP and CSO discharges.¹⁷⁻¹⁹ Detailed information on the case
95 study and the simulation model can be found in the SI, Section S1 (Pages S5-S9) in the study of Casal-
96 Campos et al.¹⁴.

97 Using an IUWWS that simulates the different parts of the wastewater system allows evaluating the
98 performance of any intervention holistically, reducing the risk of partially assessing any strategies (e.g. by
99 emphasizing good performance on one sub-system while masking poor performance on another). Further,
100 such an approach permits enriching the operationalization of concepts such as reliability, resilience and
101 sustainability, for each concept can be described by multiple metrics affecting different sub-systems.

102 **2.2.2. Future Scenarios**

103 Climate change is one of the major uncertainties that affect urban wastewater systems planning. This
104 obliges water utilities to develop more reliable, resilient and sustainable urban systems under future
105 uncertainties²⁰. To this end, four future socio-economic scenarios were used to test the IUWWS under a
106 range of different conditions: Markets, Innovation, Austerity and Lifestyles. The future scenarios were
107 mainly defined based on the planning horizon used in the UK and Western Europe²¹⁻²⁷ as well as
108 previous work on UK/EU water-associated future scenario planning exercises^{3,22,28-31}. Such alternative
109 future conditions are constituted by an ensemble of nine different parameters (see **Table 1**), representative
110 of the range of uncertain circumstances facing the IUWWS, and their influence on system performance is
111 simulated in the integrated model. **Figure S2** (in the SI) summarizes the main characteristics of each
112 future scenario. The scenarios are depicted based on two drivers: (1) governance (economic growth vs.
113 environmental awareness); and (2) values (consumerism vs. conservationism)¹⁴. These two drivers are
114 often used as key features (for their ability) to facilitate a more diverse and transparent possibility space
115³².

116 Based on the two drivers, each of the four future scenarios is characterized by four key scenario factors
117 associated with the management of the IUWWS: (I) regulation: level of regulatory control of stormwater

118 and wastewater management activities; (II) centralized maintenance: the level of activity in each scenario
 119 aimed at preserving the existing wastewater infrastructure; (III) public attitudes: public willingness
 120 towards the decentralization of responsibilities concerning urban drainage; and (IV) technology: the level
 121 of technological development occurring under each scenario.

122

123 **Table 1:** Parameter estimates affecting case study conditions under each future scenario (adapted from
 124 Casal-Campos et al.¹⁴, Copyright (2015) American Chemical Society).

Parameter	Baseline	Markets	Innovation	Austerity	Lifestyles
Misconnections (L/s)	0	7.8	0.9	4.1	1.7
Urban creep (ha)	0	87.7	58.4	70.1	29.2
Water use (L/head/day)	155	165	125	140	110
Infiltration ⁽¹⁾ (L/s)	52.4	163.7	40.5	200.1	135.5
Siltation ⁽²⁾	0.97	0.92	1	0.84	0.92
Population (inhabitants)	181,000	262,450	244,350	217,200	226,250
CC precipitation uplift (%)	0	10	10	10	10
Impervious area in new developments (ha)	0	290.0	226.0	129.0	161.0
Acceptability preference ⁽³⁾	C	C	C/D	D	D
<p>(1) It refers to infiltration of groundwater into the sewer system.</p> <p>(2) The effect of siltation, which represented system capacity loss in sewer pipes due to deposited sediment, was modeled as the corresponding reduction in pipe diameter under each scenario (corresponding to full-pipe area reduction); 1: no reduction, 0: full reduction.</p> <p>(3) The acceptability of interventions under each scenario is assessed in terms of the preference for either centralized or decentralized options. The Innovation scenario shows a mixed preference for centralized interventions, where decentralization is also promoted.</p>					

125

126 Regulations under the Lifestyles and the Innovation scenarios are high (see **Table 2**) due to the
 127 environmental-awareness drive prevalent in these states of the world, whereas the regulatory climate
 128 under Markets and Austerity is low relative to the previous two scenarios to ensure low prices and austere
 129 policies, respectively. Innovation shows the highest level of technological development encouraged by
 130 strict regulation and a drive for sustainable outcomes. This is followed by Markets, which prioritizes

131 cheap and quick solutions over high-tech developments. Lifestyles and Austerity are the scenarios with
 132 the lowest level of technological development given the limited resources available under these states of
 133 the world.

134 **Table 2:** Qualitative strength (H: high, M: medium, L: low) of key scenario factors affecting the management of the
 135 IUWWS under the considered scenarios

Future Scenario	Regulation	Centralized Maintenance	Public Attitudes	Technology
Markets	L	M	L	M
Innovation	H	H	L	H
Austerity	L	L	M	L
Lifestyles	H	L	H	L

136
 137 The level of centralized maintenance is the highest for Innovation, due to high technological
 138 developments that allow a very cost-effective maintenance of the existing infrastructure. In contrast, the
 139 decentralized responsibilities encouraged under Lifestyles and Austerity importantly affect the level of
 140 centralized maintenance, which is low relative to those under Innovation. Markets can still maintain a
 141 medium level of maintenance of the centralized sewer infrastructure due to favorable economic
 142 conditions, but limited by regulatory commitments. The public attitudes toward decentralized drainage
 143 infrastructure are highest for Lifestyles, given the conservationist views of this scenario, which strongly
 144 favor the decentralization of responsibilities. Although decentralized responsibilities are also prevalent in
 145 the Austerity scenario, these are constrained by economic issues (e.g. upfront costs or running costs)
 146 which may limit the extent of decentralization. In contrast, centralized responsibilities are dominant under
 147 Innovation and Markets, given the consumerist views of these scenarios, resulting in a low level of
 148 attitudes to decentralization.

149
 150 The parameters, considered in this study, were mostly related to changes in catchment permeability and
 151 to the variation of sewer inflows, which could hinder system capacity in the future. Permeability changes

152 were represented by the rate of urban creep occurring in the baseline catchment (i.e. loss of pervious area
153 to impervious area in the original catchment) and by the increase in impervious area occurring as a
154 consequence of urbanization (i.e. new developments built consistently with population growth in each
155 scenario). Urban creep was modeled as the loss of a fraction of pervious area which was added to the
156 impervious area fraction in each sub-catchment. The urban creep fraction remained connected to the
157 combined sewer system, as opposed to the impervious area added due to new developments, which was
158 considered to be managed by separate sewers. More details on the characteristics of the future scenarios
159 can be found in Casal-Campos et al. ¹⁴.

160 Sewer inflows in each sub-catchment were determined by the combination of misconnections,
161 groundwater infiltration, and water use flow rates occurring under each future scenario. Foul sewers
162 misconnected to storm sewers were considered a factor that could deteriorate future background water
163 quality in the river, as wastewater is discharged untreated directly into the watercourse, along with surface
164 runoff from new developments and intermittent CSO spills. Misconnections only occurred as a
165 consequence of urban development (no misconnections in the baseline case), since the baseline river
166 quality was assumed to account for any existing background pollution. In each scenario, population
167 growth relative to the baseline determined the amount of misconnected foul sewers discharging into
168 surface sewers occurring in each sub-catchment (based on the rate of misconnections assumed under each
169 future scenario). The new dry-weather flow for each sub-catchment was then calculated by adding the
170 new domestic water users (i.e. new population) and deducting the misconnected flow. Misconnections
171 were assumed to have the same flow and pollutant concentration patterns as domestic wastewater.

172 Infiltrated groundwater was considered as an extraneous inflow evenly distributed throughout the
173 catchment. Groundwater flows infiltrating into combined sewers were modeled as a rate of the total dry
174 weather flow from each sub-catchment. This resulted in an external flow being added to each sub-
175 catchment throughout the year, independently of rainfall events, with an assumed pollutant load
176 equivalent to rainfall runoff.

177 Finally, the annual precipitation increase due to climate change was modeled using an annual average
 178 uplift for rainfall intensities. The rainfall time series used for the purpose of the study is a representative
 179 data series for annual precipitation (621.5 mm) in the catchment. This was introduced in the model in the
 180 form of a 5-minute intensity time varying data file and modified by applying the 10% uplift due to climate
 181 change, expected for annual average conditions in the year 2050 in the UK or Western Europe^{23,25}. The
 182 allocation of specific estimates for each parameter (to each scenario) is described in the SI (Section S2).

183

184 2.3. Strategic Interventions

185 A number of adaption strategies associated with the management of urban stormwater and wastewater
 186 have been proposed to ameliorate the impacts and consequences used to describe system performance in
 187 the IUWWS (see Table 3).

188

189 **Table 3:** Main characteristics of proposed stand-alone strategies.

Stand-alone strategies				
Strategy	Description	Area type/system served	Impervious area served as % of catchment	Strategy type
SCC	Permeable pavement	Urban creep in driveways	5-15	Decentralized
SCP	Bio-retention planters	Residential roads	28	Decentralized
SCR	Rain gardens	Residential roofs	44	Decentralized
SS	Separate sewers	All types (in 50% of the existing catchment)	50	Centralized
CS	Improved sewer capacity	Enhanced sewer pipes (ϕ 1.5m diameter)	100	Centralized
CST	Improved sewer capacity and storage	Sewer pipes (ϕ 1.5m) and storage (25,000 m ³)	100	Centralized
OT	On-site wastewater treatment	Half of new developments	NA (only wastewater)	Decentralized

190

191 These provide a wide portfolio of interventions that can be implemented in different parts of the
 192 IUWWS, namely: the existing sewer infrastructure, existing urban areas and new developments in the

193 urban catchment. Further, these attempt to capture two main groups of strategies that are usually proposed
 194 in order to address drainage issues in urban catchments: conventional gray infrastructure options that
 195 focus on end-of-pipe solutions (rehabilitation, sewer retrofits and new development options); and a range
 196 of alternative green retrofits that affect different urban area types (private driveways, roofs and public
 197 roads). Due to the scope mentioned above, we did not include all emerging concepts such as resources
 198 and nutrient recovery, which should be studied in the future. Regarding retrofit interventions, infiltration
 199 options (permeable pavement and bio-retention planters) have been prioritized due to their complete
 200 removal of stormwater, which would improve the hydraulic performance of combined sewers when
 201 compared to less effective alternatives (e.g. green roofs or rainwater harvesting intercept an initial fraction
 202 of the stormwater).

203 A number of hybrid strategies, combining interventions in existing (i.e. retrofit) and new developed
 204 areas, were proposed in addition to the above stand-alone solutions (see Table 4). Such “hybrid
 205 strategies” were considered potentially more feasible and achievable, given the reduced implementation
 206 rates for each of the considered strategy types. Hybrid strategies (in particular mixed centralized and
 207 decentralized technology options) may additionally provide a higher degree of flexibility and adaptability
 208 to urban water systems^{33–35}. These multi-concept strategies (as opposed to the previous mono-concept
 209 stand-alone strategies) become particularly important when considering robust solutions across different
 210 world views that aim at satisfying a number of stakeholders and objectives, finding compromises that
 211 reflect the complexities of water infrastructure policies³⁶.

212

213 **Table 4:** Main characteristics of proposed hybrid strategies.

Hybrid strategies				
Strategy	Description	Area type/system served	Impervious area served as % of catchment	Strategy type
H1	Rain gardens and <i>on-site</i> wastewater treatment	50% of residential roofs and 31.5% of new developments	22	Decentralized
H2	Rain gardens and	50% of residential roofs and 20%	22 + 20	Decentralized/

	<i>separate sewers</i>	<i>separation in the existing catchment</i>		<i>Centralized</i>
H3	Separate sewers and <i>on-site wastewater treatment</i>	20% separation in the existing catchment and <i>31.5% of new developments</i>	20	Centralized/ <i>Decentralized</i>
H4	Rain gardens and <i>improved sewer capacity</i>	All residential roofs and <i>combined sewer system improvement (ϕ1.5m)</i>	44 and 56	Decentralized/ <i>Centralized</i>
H5	Rain gardens and <i>combined sewer storage</i>	50% of residential roofs and <i>combined sewer system (25,000 m³ tank)</i>	22 and 78	Decentralized/ <i>Centralized</i>

214

215 Hybrid strategies were proposed as a combination of fractions of original stand-alone mono-concept
 216 strategies, namely: roof disconnection, sewer separation and on-site wastewater treatment. These three
 217 strategies were selected as representative for retrofit decentralized, retrofit centralized and new
 218 development solutions, respectively. The resulting hybrid strategies removed a similar annual volume of
 219 stormwater or wastewater from the wastewater system. This was calculated equivalent to: disconnecting
 220 half of the residential roofs in the catchment (50% of SCR), providing on-site wastewater treatment to an
 221 average of 31.5% of new developments across scenarios (i.e. similar to implementing 63% of the OT
 222 strategy), or introducing separate sewers for 20% of the existing catchment (i.e. 40% of the SS strategy).
 223 As many as two strategies were combined in order to better understand the contributing effect of each
 224 intervention to the hybrid option. Further, it was assumed that a higher number of combined strategies
 225 may be unfeasible to implement across all the considered future scenarios, given their differing views on
 226 centralized and decentralized interventions.

227 In addition to this, the disconnection of roofs (SCR) and the rehabilitation of combined sewers in the
 228 network (CS) have been also combined (H4) to compare the performance of decentralized infiltration (in
 229 the form of rain gardens) against centralized storage (large tunnel storage implemented in the CST
 230 strategy). A combination of centralized storage (CST strategy without sewer pipe rehabilitation) and roof
 231 disconnection was also considered (H5) to assess the extent to which decentralized runoff control could
 232 complement the installation of centralized sewer storage schemes while reducing sewer replacement
 233 requirements. Apart from the mono-concept and hybrid strategies, we considered a “do-nothing” strategy,

234 which represents the current system without any interventions, for comparison with other intervention
235 strategies. The “do-nothing” option was assumed a zero-cost (in terms of capital expenditure) and low-
236 acceptability alternative throughout all scenarios, since it is expected that improvements will be needed in
237 the system by the year 2050.

238

239 **2.4. Performance Metrics**

240 The performance impacts and socio-economic consequences derived from performance failures (e.g.
241 capacity exceedance) cover a broad range of objectives of interest. These are typical key objectives used
242 by the UK water industry to make strategic decisions regarding the improvement of the levels of service
243 and urban wastewater infrastructure (i.e. associated with cost, environmental impact, flood control,
244 customer acceptability). These objectives are represented by specific performance indicators, which
245 reflect the main attributes of reliable, resilient and sustainable IUWWS. Note that these metrics are
246 proposed for a holistic assessment of the integrated urban wastewater system, presenting advantages when
247 compared with the separate management and regulation of individual sub-systems with isolated objectives
248 (i.e. surface water drainage system, sewer system, wastewater treatment system and receiving water
249 system). This allows for the assessment of the performance of decentralized systems, such as bio-
250 retention planters (SCP), rain gardens (SCR) and on-site treatment (OT), from an integrated perspective
251 using a wide range of metrics. Performance indicators for reliability, resilience and sustainability affected
252 by a number of selected objectives are presented in **Table 5**.

253 It has been recognized that an urban drainage system must be first reliable under standard design
254 conditions and then built upon by resilience under exceptional conditions, with an ultimate aim to
255 improving and/or achieving sustainability in the long term⁵. Due to the pyramidal structure of reliability,
256 resilience and sustainability, the indicators are inevitably interlinked; however, they measure different
257 aspects of system performance, as shown in Table 5. The consideration of different objectives (reliability,
258 resilience and sustainability) responds to different conditions that decision-makers cannot be sure of in an
259 uncertain future (or under conditions of deep uncertainty). Such an approach would help decision makers

260 to evaluate their required levels of compliance/service with regards to the scale and degree of complexity
 261 of their problem. For example, in a low uncertainty situation, a decision-maker may be satisfied using
 262 reliability metrics without considering further failure impacts; however, they may still find value in the
 263 resilience and sustainability metrics to address additional needs under higher uncertainty conditions.

264 **Table 5:** Performance indicators used to describe performance objectives.

Objective	Reliability indicator	Resilience indicator	Sustainability Indicator
Sewer Flooding	% time free of flood	Summation of duration-weighted flood volumes [m ³]	Total flood volume [m ³]
River DO	% time DO >4 mg/L	Summation of duration-weighted DO minima [mg/L]	6-hour minimum dissolved oxygen [mg/L]
River AMM	% time AMM <4 mg/L	Summation of duration-weighted AMM maxima [mg/L]	99 percentile total ammonia [mg/L]
CSO	% time free of spills	Summation of duration-weighted spill volumes [m ³]	Total spill volume [m ³]
River Flooding	-	-	Total flood volume [m ³]
GHG Emissions	-	-	Operational emissions from pumping and treatment [tCO ₂]
Cost	-	-	Present value of whole-life costs [£]
Acceptability	-	-	Acceptability of strategies [1/2/3] (*)

265 (*) [1/2/3] represents the expected acceptability of strategies [high/med/low] based on future scenario narratives. In scenarios with a
 266 preference for centralized solutions (denoted by C in Table 1), decentralized strategies score poorly (low acceptability, score: 1), and vice versa.

267 It is important to note that IUWWSs are very site-dependent systems whose performance may be
 268 significantly affected by local characteristics (e.g. climate), monitoring regimes, operational conditions
 269 (e.g. influent pollutant concentrations) and the age of the system itself, often presenting a wide variability
 270 of values³⁷.

271

272 **2.4.1. Reliability Indicators**

273 In this study, reliability is defined as the degree to which the system minimizes level of service failure
 274 frequency over its design life when subject to standard loading⁵. Service failure here means failing to
 275 comply with the levels required by regulations, not considering mechanical failures, such as those from

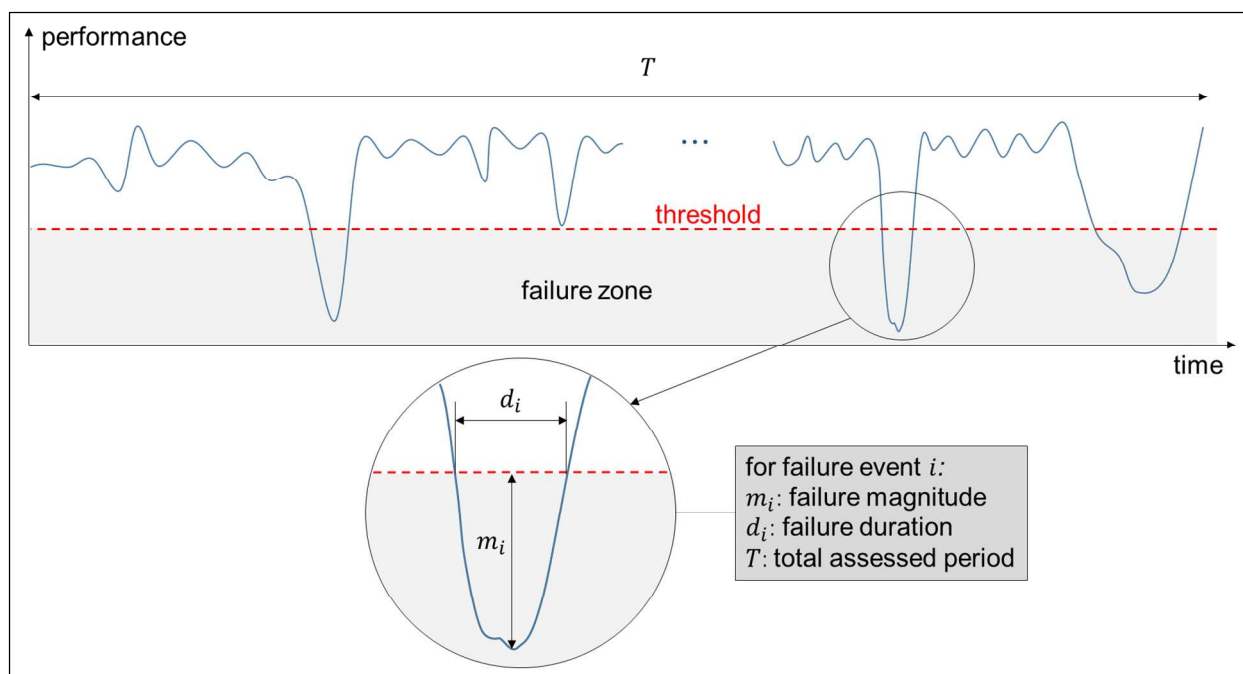
276 pumps and aerators, as the focus is on the long-term urban wastewater system planning under uncertain
277 future conditions.

278 The reliability indicators presented in **Table 5** are characterized by the consistency of acceptable levels
279 of service (used for the system design), measured as the probability that no failure occurs within a fixed
280 period of time³⁸ (i.e. a given threshold is not violated within the one year assessment). The general
281 mathematical expression describing reliability indicators is,

$$\text{Reliability} = 1 - \sum_i \frac{d_i}{T} \quad [1]$$

282 Where d_i represents the duration of each failure occurring within the total assessed period T (i.e. one
283 year). Reliability therefore denotes the annual fraction of time free of failure (See **Figure 1**), and is
284 mainly used for system design.

285 Failure thresholds for the sewer flooding and CSO objectives were set at zero, so that only a complete
286 avoidance of spill or flood events could translate into full reliability under these objectives. Water quality
287 failure thresholds were defined as critical concentrations of river dissolved oxygen and river total
288 ammonia (i.e. 4mg/L), after Schütze et al.¹⁵.



289
 290 **Figure 1:** A generic description of failure and its components as used in this study. When system
 291 performance falls below a predefined threshold, a failure occurs. Magnitude and duration are key
 292 attributes describing the operational implications of a failure.

293 **2.4.2. Resilience Indicators**

294 To face future challenges, it must be ensured that drainage solutions operate safely (i.e. fail-safe,
 295 reliably) as far as practicably possible, but that they also respond safely to failure (i.e. safe-to-fail,
 296 resiliently)⁴⁰. These systems can behave more flexibly and recover quickly in order to reduce damage
 297 (and the disruption to the level of service) when failure occurs⁴¹. To capture the attributes described
 298 above, resilience is here defined as the degree to which the system minimizes level of service failure
 299 magnitude and duration when subject to exceptional conditions, represented here by a threat or
 300 combination of threats⁵.

301 The same four objectives used for reliability are employed to measure resilience as listed in **Table 5**;
 302 however, they are calculated differently, with a combination of both the magnitude and the duration of
 303 failure events occurring within the assessed time period (See **Figure 1**). This resembles current
 304 definitions of failure severity in urban water systems (or better, the complement of resilience) which refer

305 to measures of the loss of functionality in the system^{9,42}. The expression proposed here is the weighted
306 summation (relative to failure duration) of failure magnitudes,

$$\text{Severity} = \sum_i \frac{m_i \times d_i}{T} \quad [2]$$

307 where m_i and d_i represent the magnitude and duration of the failures occurring within the total assessed
308 period T (i.e. one year), respectively.

309 The severity of each failure event (and its units) was therefore described by the main magnitude of
310 interest for that failure, namely: volume for sewer flood and CSO failures; minimum in-river
311 concentration for dissolved oxygen; and maximum in-river concentration for total ammonia. Resilience
312 was thus understood as the complementary attribute of severity, so that maximum severity translated into
313 minimum resilience and vice versa⁴³. Equation [2] is a simplified expression that combines the
314 magnitudes and durations of the failures expected in a typical year for comparative purposes in the
315 assessments carried out in this study. Thus, this does not intend to be an accurate representation of the
316 absolute severity of failures but rather a proxy for comparing the relative failure severity of different
317 options under the same annual conditions.

318 **2.4.3. Sustainability Indicators**

319 The adopted definitions of reliability and resilience refer to the operational performance of urban
320 wastewater systems when facing future change, whether this is before or after a failure occurs. Indeed,
321 these are attributes of how a system endures, responds to stress and recovers from failure to minimize any
322 potential impacts.

323 In contrast, the concept of sustainability is defined here as “the degree to which the system maintains
324 levels of service in the long-term whilst maximizing social, economic and environmental goals”⁵. Thus
325 sustainability is associated with the performance of the system in the long term (even beyond the design
326 life), including failure and non-failure periods. When impacts derived from failure (i.e. from how reliable
327 and resilient the system is) extend to the wider social, environmental and economic systems in urban

328 areas, they become consequences that affect the system's sustainability ⁵. As operational impacts (e.g.
329 magnitude or duration of flooding events) interact with the three pillars of sustainability in the form of
330 consequences (e.g. damages to society, the environment and the economy), threats are transferred to the
331 recipients of water services (i.e. society, natural environment and economic systems).

332 Sustainability indicators in **Table 5** are related to the likely consequences of failure to the economy, the
333 environment and society. These wider implications include four new objectives (river flooding, GHG
334 emissions, cost and acceptability) not accounted for in the reliability and resilience indicators. Their
335 omission in reliability and resilience illustrates the operational nature of reliability and resilience, since
336 there is no clear association of any of these additional objectives with operational performance failure.
337 Instead, these were considered "consequential" objectives, which not only derived the operational
338 performance of the system, but also became paramount when looking at the overall long-term economic,
339 environmental and social (i.e. consequences) trade-offs of investment decisions ⁴⁴ initially triggered by
340 operational drivers (e.g. flooding or water quality objectives). GHG emissions are calculated from the
341 energy use required for wastewater pumping and treatment processes during one year operational
342 period¹⁴. Whole-life costs are calculated as the present values of capital and maintenance costs during an
343 operational life of 35 years with a discount rate of 3.5% ¹⁴, see the SI (Section S3).

344

345 **2.5. Robustness Indexes for Reliability, Resilience and Sustainability**

346 The scenarios presented earlier in the text described some of the numerous uncertainties that may affect
347 our predictions concerning IUWWSs in the future. Indeed, a characteristic of long-term policy analysis is
348 working under conditions of *deep uncertainty*, i.e. where analysts do not know or the parties to a decision
349 cannot agree on: (1) the appropriate conceptual models that describe the relationships among the key
350 driving forces that will shape the long-term future; (2) the probability distributions used to represent
351 uncertainty about key variables and parameters in the mathematical representations of these conceptual
352 models; or (3) how to value the desirability of alternative outcomes ⁴⁵. This is fundamentally different

353 from other types of uncertainty which could be quantified using various approaches such as probabilities,
 354 imprecise probabilities, intervals and fuzzy sets.^{3,45}

355 The variety of alternatives considered and the uncertainty over future conditions recommends the
 356 exploration of robust strategies. These are strategies that perform reasonably well compared to other
 357 alternatives across a wide range of plausible scenarios⁴⁶. In a context of deep uncertainty, a robust
 358 strategy will generally trade optimal performance for less sensitivity to broken assumptions, performing
 359 satisfactorily (although sub-optimally) over a range of possible futures^{47,48}.

360 **2.5.1. Performance regrets**

361 The approach used in this study evaluated the robustness of strategies by assessing their relative
 362 performance loss (i.e. regret) across all the objectives and future scenarios previously described¹⁴. Such
 363 performance loss was assessed through each of the performance indicators presented in **Table 5**,
 364 representing the relative impacts and consequences of failure in the form of performance regrets.

365 The concept of regret (or opportunity loss), as introduced by Savage⁴⁹, was used to make decision
 366 recommendations on mutually exclusive strategies. The regret of strategy s under a future state f is
 367 defined as the difference between the performance of s (for objective i) and that of the best-performing
 368 strategy s' for the same future scenario f and objective i ⁴⁷,

$$\text{Regret}_i(s, f) = |\max_{s'} [\text{Performance}_i(s', f)] - \text{Performance}_i(s, f)| \quad [3]$$

369 The regret of selecting a specific drainage strategy s , as calculated in Equation [3], is understood as the
 370 missed opportunity to choose an alternative strategy which would have resulted in a more beneficial
 371 outcome once the future is materialized under scenario f ⁵⁰. Thus, the basis of regret-based approaches is
 372 to select the strategy that minimizes the opportunity loss or regret accrued from all the considered
 373 objectives across all future scenarios.

374 As discussed in the previous section, we consider different ranges of indicators and objectives to assess
 375 reliability, resilience and sustainability of the system in different future states. Some of the indicators are
 376 evaluated on an operational basis; few others (e.g. costs) are assessed over the lifetime of the

377 system/intervention. In order to ensure consistency in the results and credibility of the approach,
 378 performance regrets concerning any objective i under any future scenario f were normalized relative to
 379 the most regrettable alternative s^* in that objective and scenario (i.e. the one with the largest regret
 380 according to [3]). Thereby, the proposed equations ([4], [5] and [6]) worked as a utility function that
 381 assigned normalized regret scores according to performance (i.e. between 0 and 1, from best to worst
 382 performance) for each strategy and future scenario ¹⁴.

$$Rel_i(s, f) = \frac{\text{Regret}_i(s, f)}{\max_{s^*}[\text{Regret}_i(s_{rel}, f)]} \quad \text{for } i = 1, \dots, 4 \quad [4]$$

$$Res_j(s, f) = \frac{\text{Regret}_j(s, f)}{\max_{s^{**}}[\text{Regret}_j(s_{res}, f)]} \quad \text{for } j = 1, \dots, 4 \quad [5]$$

$$Sus_k(s, f) = \frac{\text{Regret}_k(s, f)}{\max_{s^{***}}[\text{Regret}_k(s_{sus}, f)]} \quad \text{for } k = 1, \dots, 8 \quad [6]$$

383 $Rel_i(s, f)$ represents the normalised performance regret of strategy s under scenario f for the i th
 384 reliability objective (one for each of the five reliability indicators in **Table 5**). An analogous description
 385 of $Res_j(s, f)$ and $Sus_k(s, f)$ applies to the j^{th} resilience indicator and the k th sustainability indicator in
 386 **Table 5**, respectively. The worst performing strategies in each case are represented by s_{rel} , s_{res} and s_{sus} .

387 **2.5.2. Reliability, Resilience and Sustainability Indexes**

388 By using Equations [7]-[9], the reliability, resilience and sustainability of each strategy under each
 389 future scenario can be encapsulated in a single multi-criteria regret index. Each of these indexes has
 390 therefore been used to compare the relative overall performance (in terms of reliability, resilience or
 391 sustainability) of each strategy within each future state. This reduces the problem of assessing multiple
 392 utilities (i.e. five normalised performance regrets for reliability and resilience indexes, or eight for the
 393 sustainability index) into one of assessing a one-dimensional weighted utility ⁵¹.

$$\overline{Rel}(s, f) = \sum_i \left(w_i^f Rel_i(s, f) \right) \quad \text{for } i = 1, \dots, 4 \quad [7]$$

$$\overline{Res}(s, f) = \sum_j \left(w_j^f Res_j(s, f) \right) \quad \text{for } j = 1, \dots, 4 \quad [8]$$

$$\overline{Sus}(s, f) = \sum_k \left(w_k^f Sus_k(s, f) \right) \quad \text{for } k = 1, \dots, 8 \quad [9]$$

394 $\overline{Rel}(s, f)$ represents the reliability index of strategy s under future scenario f as the weighted
 395 summation of reliability normalised performance regrets $Rel_i(s, f)$. Analogous descriptions apply to
 396 resilience index $\overline{Res}(s, f)$ and sustainability index $\overline{Sus}(s, f)$. w_i^f , w_j^f and w_k^f represent the relative
 397 weights of the i th, j th and k th objectives (associated with reliability, resilience and sustainability,
 398 respectively) in future scenario f ; with $\sum_i w_i^f = \sum_j w_j^f = \sum_k w_k^f = 1$.

399 Weights for each future scenario (**Table 6**) were calculated using the Analytical Hierarchy Process
 400 (AHP) method with help of a panel of (four) decision makers (from both academia and water authorities
 401 in the UK). A pairwise comparison between the criteria was carried out with the ultimate goal of
 402 determining the relative weights of different indicators. The pairwise comparison was implemented by
 403 establishing a reciprocal matrix wherein scores are assigned based on the relative importance of one
 404 objective relative to another⁵². More details on the AHP method and how the pairwise comparison was
 405 carried out are provided in the SI (Sections S4 and S5).

406

407 **Table 6:** Weights applied to performance objectives for each future scenario (first row refers to reliability
 408 and resilience weights w_i^f, w_j^f ; second row to sustainability weights w_k^f).

$w_i^f = w_j^f$	Sewer	River	River	CSO	River	GHG	Cost	Accept.
w_k^f	Flooding	DO	AMM		Flooding	Emissions		
Markets	0.53	0.12	0.12	0.23	-	-	-	-
	0.19	0.06	0.06	0.10	0.19	0.04	0.27	0.10
Innovation	0.51	0.11	0.11	0.27	-	-	-	-
	0.22	0.08	0.08	0.14	0.20	0.05	0.19	0.04
Austerity	0.43	0.10	0.10	0.37	-	-	-	-
	0.17	0.06	0.06	0.12	0.22	0.04	0.29	0.04
Lifestyles	0.10	0.36	0.36	0.20	-	-	-	-
	0.05	0.19	0.19	0.11	0.05	0.19	0.03	0.19

409

410 **2.5.3. Robustness Indexes**

411 There has been some debate about the definition and evaluation of robustness, since some authors
 412 define this term as the contrary of vulnerability^{53,54} while others consider it a characteristic attribute of
 413 resilient systems⁵⁵⁻⁵⁷. Schoen et al.⁵⁶ defines robustness as “strength, or the ability of the system to
 414 withstand a given level of stress or demand without suffering degradation or loss of function”. In this
 415 study, the robustness of a strategy in the future is defined as its ability to perform well regardless of future
 416 conditions. A “robustly reliable” (or robustly resilient or robustly sustainable) drainage strategy is
 417 therefore one that is reliable (or resilient or sustainable) under a number of future scenarios.

418 Equations [10]-[12] express how robust the qualities of reliability, resilience and sustainability are
 419 for each strategy across all future scenarios. The four reliability indexes $\overline{Rel}(s, f)$ obtained through [7]
 420 were combined to calculate the reliability robustness index for strategy s or $Rob_{Rel}(s)$ (see [10]).
 421 Resilience indexes $\overline{Res}(s, f)$ and sustainability indexes $\overline{Sus}(s, f)$ are similarly merged into a resilience
 422 robustness index $Rob_{Res}(s)$ and a sustainability robustness index $Rob_{Sus}(s)$, respectively (see [11] and [
 423 12] below).

$$424 \quad Rob_{Rel}(s) = \frac{\sum_f (\overline{Rel}(s, f))}{4} \quad \text{for } f = 1, \dots, 4 \quad [10]$$

$$425 \quad Rob_{Res}(s) = \frac{\sum_f (\overline{Res}(s, f))}{4} \quad \text{for } f = 1, \dots, 4 \quad [11]$$

$$426 \quad Rob_{Sus}(s) = \frac{\sum_f (\overline{Sus}(s, f))}{4} \quad \text{for } f = 1, \dots, 4 \quad [12]$$

427 The arithmetic mean applied to either set of four scenario indexes (reliability, resilience or
 428 sustainability indexes) was considered an adequate representation of overall regret, providing an integral
 429 picture of robustness across performance objectives and scenarios for each strategy. Consequently, the
 430 strategies with the smallest reliability, resilience and sustainability robustness index (i.e. the smallest
 mean value) were regarded as the most robust alternatives in reliability, resilience and sustainability
 terms.

431 3. RESULTS AND DISCUSSION

432 As mentioned in Section 2, 12 strategies, including 7 stand-alone and 5 hybrid options, are assessed
433 under the proposed future scenarios (i.e. Markets, Innovation, Austerity, and Lifestyles). Note that these
434 strategies are not designed for a specific future scenario due to the deep uncertainties involved. On the
435 basis of the performance assessment carried out, the regrets of each strategy are calculated for each
436 scenario and robustness indexes determined across all scenarios, as shown below separately for reliability,
437 resilience and sustainability. Additional details for reliability, resilience and sustainability performance
438 indicators are provided in the SI.

439

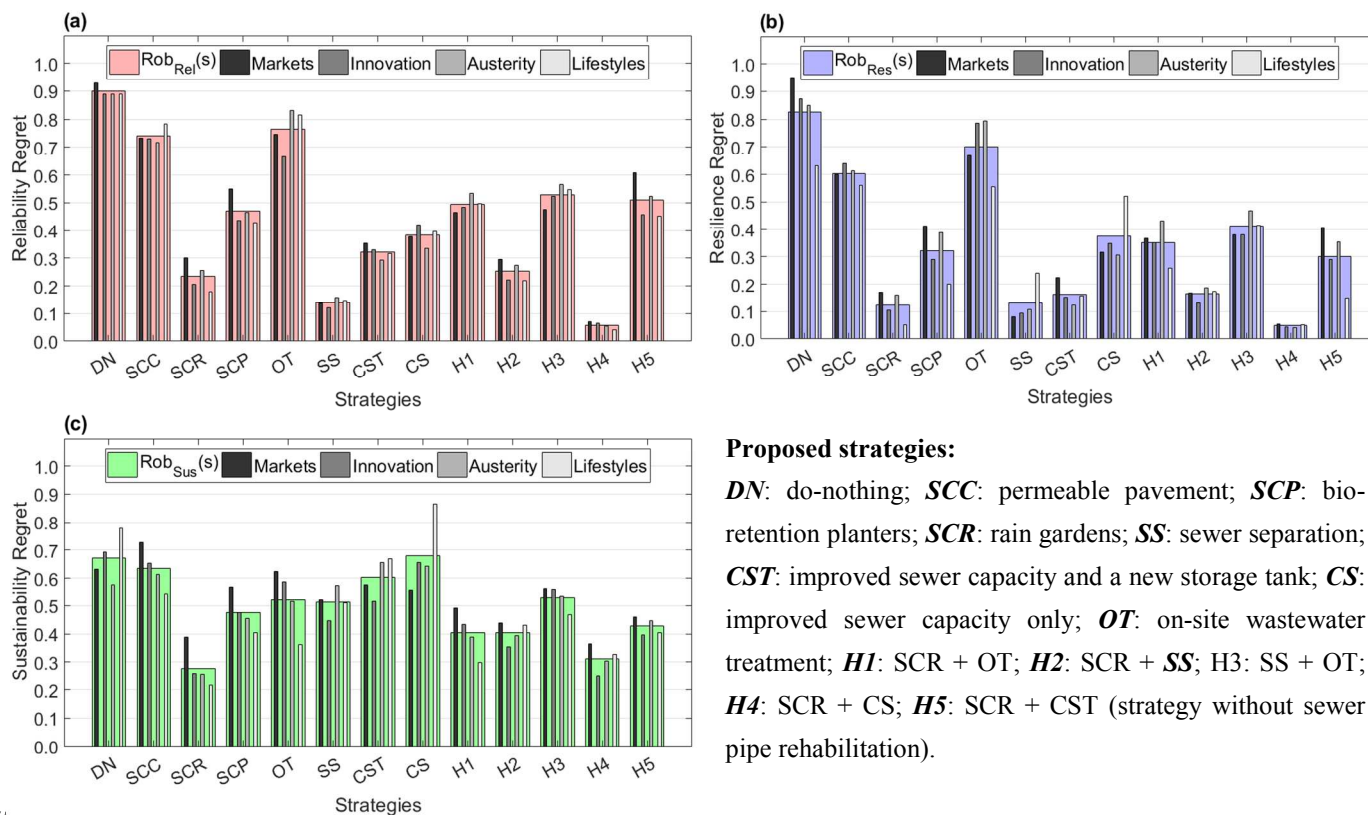
440 3.1. Reliability Robustness

441 The reliability robustness index for each strategy implemented ($Rob_{Rel}(s)$) is presented in **Figure 2(a)**,
442 along with their specific scenario reliability indexes ($\overline{Rel}(s, f)$). The most robust strategy regarding
443 reliability (i.e. most “robustly reliable” option) was the mixed implementation of sewer rehabilitation and
444 decentralized retrofit rain gardens for roofs (H4 strategy). This strategy was followed by the stand-alone
445 disconnection of roofs using rain gardens (SCR). The reliability index of H4 (CS + SCR) under Markets
446 shows the worst performance compared to those of this strategy under the other three future scenarios,
447 although better than the reliability index of SCR. Therefore, the implementation of SCR could be enough
448 if future conditions are lenient towards Markets, avoiding the expansion of sewers included in H4 and
449 requiring less investment effort.

450

451 The least robust alternatives were “do-nothing” (i.e., no improvements in the system) and on-site
452 wastewater treatment for new developments (OT); the latter with a reliability robustness index similar to
453 the mitigation of urban creep using permeable pavement (SCC). The high reliability regret of these
454 strategies illustrate the limited failure duration improvements obtained relative to “do-nothing” across
455 scenarios.

456

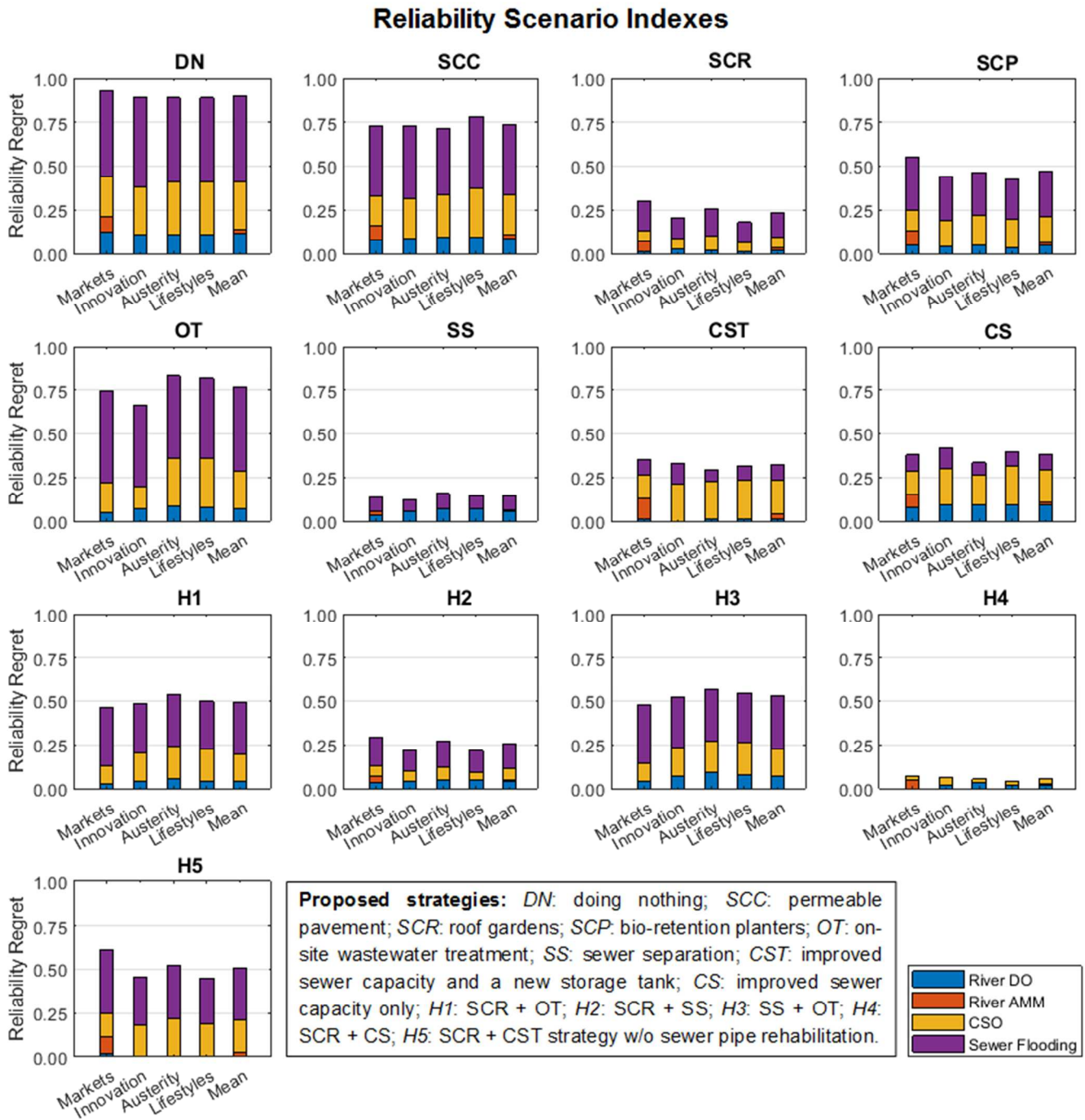


457
 458 **Figure 2:** (a) Reliability robustness index (red bars) and individual scenario reliability indexes (gray bars
 459 within); (b) Resilience robustness index (blue bars) and individual scenario resilience indexes; (c)
 460 Sustainability robustness index (green bars) and individual scenario sustainability indexes for the
 461 proposed strategies.

462 The robustness of reliability attributes in retrofit decentralized strategies (SCR, SCP and SCC)
 463 remained proportionate to the disconnected impermeable area in the existing catchment (i.e. more
 464 disconnected area, thus better performance). The scenario reliability indexes for SCR and SCP followed
 465 very similar patterns, with regrets in Markets and Austerity above the average regret (i.e. reliability
 466 robustness index) and Innovation and Lifestyles below average regret. The scenario reliability indexes for
 467 SCC presented a flatter profile, providing a more consistent amount of reliability regret across scenarios.
 468 This suggests that the mitigation of urban creep was not effective in improving the reliability of the

469 system to failures in the least favorable scenarios (e.g. high water use, high development), while low
470 creep mitigation rates in more favorable scenarios prevented larger reliability gains.

471



472 **Figure 3:** Performance regrets of $\overline{Rel}(s, f)$ and $Rob_{Rel}(s)$

473

474 **Figure 3** illustrates a breakdown (stacked chart) of the future scenarios and robustness indexes for

475 reliability, where the performance of each strategy (with respect to each indicator/objective) is presented

476 (the stacked charts for resilience and sustainability are illustrated in the SI, **Figures S3** and **S4**,
477 respectively).

478 **Figure 3** shows that the most dominant color is purple (representing the sewer flooding objective),
479 followed by gold color (CSO objective). In fact, the strategies with the worst reliability indexes did not
480 perform well under these two objectives (e.g. DN, SCC, and OT). Concerning gray infrastructure
481 strategies, sewer rehabilitation and storage (CST) was ranked within the low-regret end of the reliability
482 scale, followed by sewer separation (SS) and, falling back into lower reliability positions, the stand-alone
483 rehabilitation of sewer pipes (CS). These last two strategies were not effective in enhancing the reliability
484 of the system regarding river flooding and dissolved oxygen failure probability (see

485 **Figure 3**). The comparison of regret indexes for H4 (i.e. rain gardens and improved sewer capacity)
486 and CST (i.e. improved sewer capacity and storage) indicates that the reliability of source control
487 techniques such as rain gardens is higher than those of centralized storage schemes, whether or not these
488 included system capacity rehabilitation such as sewer pipe enlargement. In a similar way, multi-concept
489 hybrid strategies (H1 to H5) performed most reliably when retrofit roof disconnection was involved in the
490 interventions (i.e. H4, H2 and H5). In contrast, hybrid strategies influenced by on-site wastewater
491 treatment in new developments (H1 and H3) resulted in higher-regret reliability indexes, partly reflecting
492 on the low reliability performance of the OT strategy.

493

494 **3.2. Resilience Robustness**

495 The resilience robustness indexes presented in **Figure 2(b)** resulted in low regrets for the H4 strategy
496 and the SCR strategy (similar to the results of reliability robustness shown in **Figure 2(a)**), occupying the
497 most robust positions when compared with the rest of strategies. As in the reliability case, “do-nothing”
498 and on-site treatment (D-N and OT strategies, respectively) obtained the worst resilience robustness
499 indexes as well as the most regretful individual scenario resilience indexes. Retrofit decentralized
500 strategies (SCR, SCP and SCC) maintained their rank positions relative to their reliability robustness

501 ranking. This situation contrasted with the resilience robustness of gray infrastructure alternatives (CST,
502 SS and CS), which were displaced by the improved robustness of hybrid strategies (H2, H5 and H3,
503 respectively).

504

505 In general, a reduction in the resilience robustness index of strategies was observed relative to their
506 reliability robustness indexes (see **Figure 2(a)** and **Figure 2(a)**). The weighting of failure duration
507 through failure magnitudes was mainly responsible for this, increasing the significance of the most severe
508 annual failures and thus reducing that of the most common (i.e. least severe) ones. In this sense, the role
509 of gray infrastructure solutions in enhancing the conveyance capacity of the system did not have a
510 negative impact on failure durations (in particular those affecting river conditions downstream) but it
511 demonstrated to have a more acute effect on failure magnitudes (see **Figure S3**, in the SI). This situation
512 was most relevant for stand-alone gray infrastructure options such as SS and CS (see **Figure S3**, in the
513 SI), whose resilience regrets under the Lifestyles scenario (i.e. high infiltration to sewers, low
514 maintenance and low technology) increased beyond those of any other strategies relative to the reliability
515 case. The extreme conditions defined in Markets (i.e. high creep, high population, high water use) caused
516 an increase in river flooding regret and water quality regrets, especially for gray CST (see **Figure S3**, in
517 the SI), that could not be fully compensated by reducing regrets in other scenarios and objectives. This
518 meant that gray infrastructure strategies deteriorated failure conditions during adverse scenario
519 circumstances when compared to other alternatives.

520

521 **3.3. Sustainability Robustness**

522 The implementation of retrofit rain gardens for roof disconnection (SCR strategy) resulted in the most
523 robust alternative for sustainability, followed by the multi-concept strategy H4 (retrofit rain gardens plus
524 rehabilitation of sewer pipes). The assessment of sustainability robustness indexes (see **Figure 2(c)**)
525 introduced river flooding, cost, GHG emissions and acceptability criteria, which favored the stand-alone
526 disconnection of roofs with rain gardens (SCR) due to its low cost and acceptability in scenarios where

527 these criteria were highly valued (namely: Austerity and Lifestyles), as opposed to H4 (more costly and
528 less acceptable in those scenarios); see **Tables S6** and **S7** in the SI.

529 “Do-nothing” (DN) is the least robust strategy in sustainability terms, with sewer pipe rehabilitation
530 (CS) being penultimate in the robustness hierarchy, despite obtaining a better resilience robustness
531 indexes compared to SCC (permeable pavement), SS (sewer separation), CS (improved sewer capacity)
532 and CST (improved sewer capacity and storage) under Austerity scenario, mainly due to high costs and
533 GHG emissions, and low acceptability of these four strategies in this scenario (i.e. lack of economic,
534 environmental, and social performance). As expected DN has the best performance (i.e. the lowest regret)
535 with respect to cost in all future scenarios (see **Figure S4**). Should investments be made in the future, the
536 DN approach allows for flexibility to implementing new strategies when future needs become clear.

537 Hybrid alternatives showed increased sustainability robustness when compared to mono-concept
538 interventions and, in particular, gray infrastructure stand-alone strategies (i.e. CST, SS and CS). In this
539 sense, hybrid interventions involving retrofit roof disconnection with rain gardens (i.e. H4, H1, H2 and
540 H5) improved their sustainability robustness at the expense of that of mono-concept gray infrastructure
541 options, which failed to maintain low regrets across future scenarios.

542 Gray infrastructure strategies were generally penalized by their cost burden in scenarios that otherwise
543 favored centralized solutions in the acceptability objective (e.g. CST in Markets), as well as by
544 environmental and social issues in scenarios where cost was a less important factor to decisions (e.g. CS
545 in Lifestyles). This situation contrasted with that of hybrid alternatives, which attained a more balanced
546 performance, given their even distribution of regrets across objectives and scenarios.

547 Retrofit decentralized strategies (i.e. SCR, SCP and SCC) showed a wide variety of robustness in
548 sustainability terms (as in reliability and resilience), mostly linked to its main trade-off; i.e. the balance
549 between cost and operational performance. In most objectives, unpronounced trade-offs for these strategy
550 types ensured a balanced accumulated regret for most scenarios, proportionate to the level of
551 impermeable area intervened. However, the cost of permeable pavement implementation (SCC) and bio-
552 retention planters (SCP) highly constrained their obtaining low-regret sustainability scenario indexes and

553 thus low-regret robustness indexes (**Figure S4**). In particular creep mitigation using permeable pavement
554 (SCC) proved ineffective in outweighing the cost regret in the Markets scenario, or in achieving low
555 regrets in more operationally lenient scenarios (i.e. Lifestyles) given the limited benefits derived from low
556 creep removal rates.

557 On-site treatment of part of the wastewater from new developments (OT) showed a significant
558 improvement in sustainability robustness relative to the reliability and resilience regrets presented above.
559 In spite of not directly addressing stormwater management issues, the OT strategy compensated these
560 high regrets with a modest cost trade-off and large improvements in GHG emissions. This was also a
561 factor which contributed to the improved performance of some hybrid strategies, such as H1.

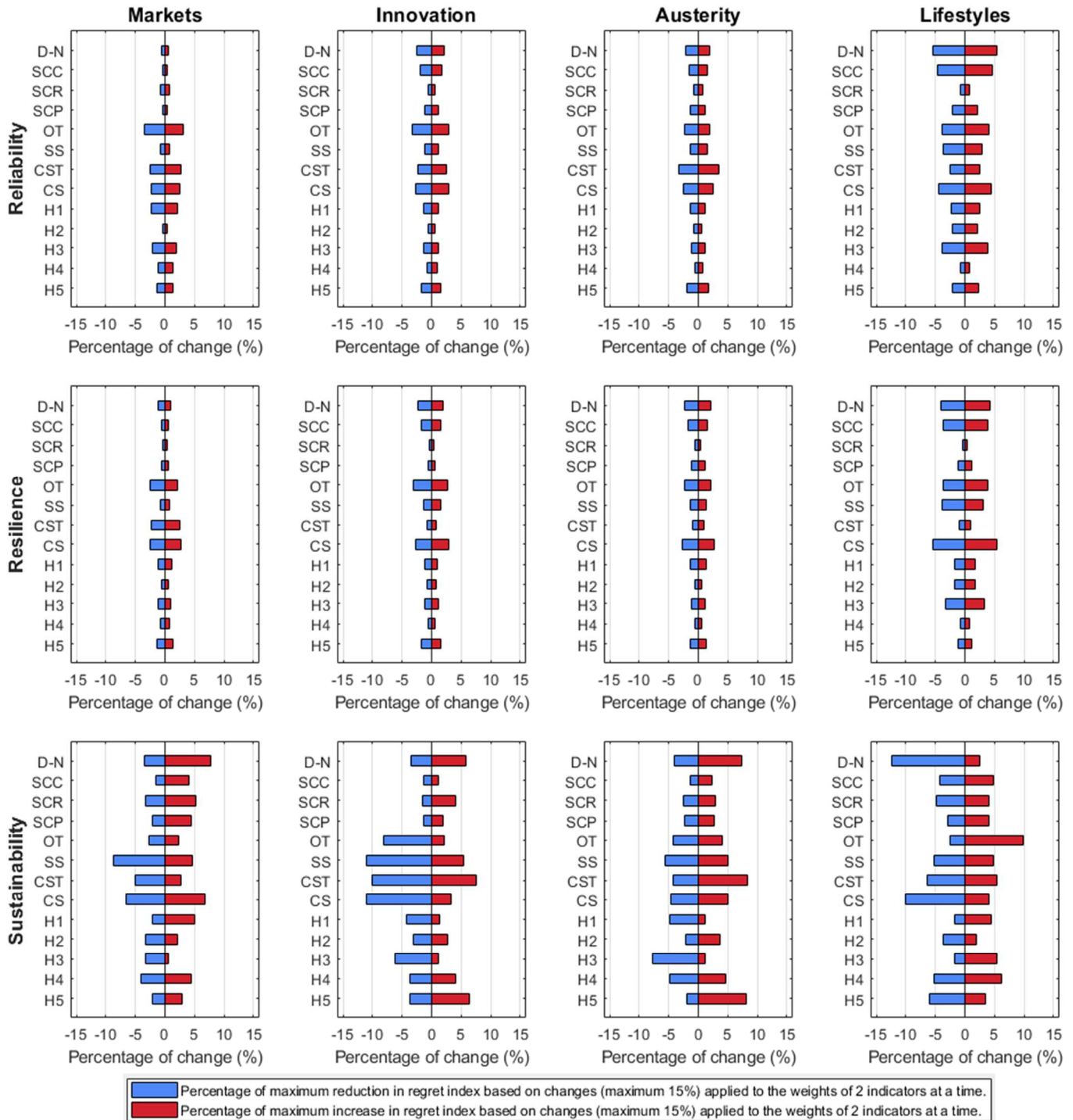
562

563 **3.4. Sensitivity Analysis**

564 A two-at-a-time sensitivity analysis was carried out with a focus on the importance weight of each
565 assessment objective (i.e. sewer flooding, river water quality, combined sewer overflow (CSO), river
566 flooding, GHG emissions, cost and acceptability). The assigned weights were altered by $\pm 15\%$ in each
567 future scenario (i.e. Markets, Innovation, Austerity and Lifestyles) for the calculation of reliability,
568 resilience and sustainability indexes (**Figure 4**).

569 The sensitivity was calculated as the percentage variation in the reliability index ($\overline{Rel}(s, f)$), resilience
570 index ($\overline{Res}(s, f)$) and sustainability index ($\overline{Sus}(s, f)$) relative to the value obtained with the originally
571 assigned weights. **Figure 4** shows that the reliability indexes are approximately within a variation of $\pm 3\%$
572 in Markets, Innovation and Austerity, and $\pm 5\%$ in Lifestyles. It is also shown that sensitivity associated
573 with the resilience indexes is very similar to that of the reliability; the only difference is shown in the
574 results of Markets scenarios where the maximum difference from the original resilience indexes is $\pm 2\%$.

575



576 **Figure 4:** Sensitivity of individual scenario reliability indexes, resilience indexes and sustainability indexes to a
 577 two-at-a-time alteration of objectives' weights ($\pm 15\%$) for the proposed strategies. DN: doing nothing; SCC:
 578 permeable pavement; SCP: bio-retention planters; SCR: rain gardens; SS: sewer separation; CST: improved sewer

579 capacity and new storage tank; CS: improved sewer capacity only; OT: on-site wastewater treatment; H1: SCR +
580 OT; H2: SCR + SS; H3: SS + OT; H4: SCR + CS; H5: SCR + CST strategy without sewer pipe rehabilitation.

581
582 On the other hand, the sustainability indexes present more sensitivity to changes in the objectives'
583 weights, with maximum variations of approximately +9% and -12%. The sensitivity varies from strategy
584 to strategy, with lower extremes around -12% for the “do-nothing” strategy (DN) (for sustainability index
585 under Lifestyles) and -11% for on-site treatment (OT) and for improved sewer capacity (CT) (for
586 sustainability indexes under Innovation), and an upper extreme of +9% for OT (for sustainability indexes
587 under Lifestyles). **Figure 4** illustrates that hybrid strategies are generally less sensitive to changes applied
588 to the original weights when compared to mono-concept strategies. On average, the highest degree of
589 confidence, when the original weights were altered by $\pm 15\%$, can be seen in the results of SCR ($\pm 1.5\%$),
590 H2 ($\pm 1.5\%$), SCP ($\pm 1.6\%$), H4 ($\pm 2.0\%$) and H1 ($\pm 2.0\%$), whereas, the highest variations are observed in
591 the results of CS, DN, OT, and CST (with average variations of $\pm 4.3\%$, $\pm 3.6\%$, $\pm 3.5\%$ and $\pm 3.4\%$,
592 respectively).

593

594 3.5. Implications

595 The concept of regret is a useful tool for the comparative assessment of a large array of performance
596 indicators that, given their different natures (quantitative or qualitative) and scales, are often difficult to
597 compare and normalize. The integrative nature of regret permits analysts to operationalize complex
598 concepts such as reliability, resilience or sustainability into indexes that can illustrate the overall
599 robustness of the proposed strategies. The case study investigated here is in the context of UK
600 regulations; however, the scenarios, strategies and performance metrics are typically found in Western
601 Europe and many other countries alike. Therefore, the findings of this study can be broadly applicable to
602 those countries and provide decision makers and utility managers with enhanced insight into the
603 development of more reliable, resilient and sustainable urban wastewater systems using gray, green and
604 hybrid options.

605 The results presented in this paper demonstrate that strategies that are robust for sustainability in the
606 case study are likely to be robust for both resilience and reliability across future scenarios (e.g. SCR and
607 H4), whereas robustness for resilience and, in particular, for reliability cannot ensure robustness for
608 sustainability. In this sense, the behavior of some strategies (e.g. H1 and OT) appeared to contradict this
609 view, since their low robustness in reliability and resilience terms later translated into higher
610 sustainability robustness; however, these were relatively far from low-regret robustness indexes at the top
611 of the hierarchy. Additional objectives, not accounted for in reliability and resilience assessments (river
612 flooding, cost, GHG emissions and acceptability), made up a significant part of the enhanced performance
613 of these strategies, while benefitting from the low performance of other options in these objectives.
614 Therefore, reliability, resilience, and sustainability indicators cannot be used interchangeably and should
615 be looked at and analyzed depending upon the purpose of the decision making exercise. Indeed,
616 reliability, resilience and sustainability approaches need to be used proportionately to the complexity and
617 scale of the problem to be solved. As mentioned earlier in the text, decision-makers may be interested in
618 satisfying a limited number of objectives in a low uncertainty problem, where reliability is sought and a
619 sustainability-led analysis may excessively complicate (or even hinder) their decision. Instead, they may
620 prefer to approach a highly complex and uncertain problem from a sustainability point of view in order to
621 better balance the potentially critical trade-offs present in a much more challenging decision exercise.
622 Robustness for sustainability is regarded as a more demanding attribute as it focuses on economic,
623 environmental and socio consequences, therefore, a larger number of criteria are involved in its definition
624 and more trade-offs can affect the sustainability index.

625 It is important to note that such relationships are not categorical since they are dependent on the
626 selected performance indicators and how their regrets are traded between objectives within each
627 robustness assessment. For example, there are less known adverse impacts of gray and green options on
628 human health (e.g. pathogen-related risks) from an integrated systems' perspective^{58,59}, and the objective
629 tradeoffs might be different when these impacts are considered. This and other emerging issues should be
630 further investigated once such data and knowledge become available. In this sense, the multi-criteria and

631 multi-scenario assessment conditions presented in the case study favored “balanced” strategies; those
632 without marked performance trade-offs which were generally responsible for larger regrets across future
633 scenarios.

634 Gray infrastructure alternatives where environmental, economic and social objectives are more difficult
635 to be simultaneously satisfied (e.g. cost, water quality issues) are not robust for sustainability, even in
636 cases where they show robustness concerning both reliability and resilience. Such trade-offs are
637 compensated for in hybrid strategies combining gray infrastructure interventions with green retrofit
638 strategies. In this context, hybrid strategies can mitigate loss in performance by diversifying the number
639 of interventions and by complementing the benefits and strengths obtained from each mono-concept
640 strategy. These multi-concept strategies demonstrate that the robustness of gray infrastructure strategies
641 for reliability, resilience and sustainability can be enhanced by using green retrofits as these are able to
642 better negotiate their performance regrets. In this sense, decentralized infiltration through rain gardens
643 proves to be more robust than large storage tunnel interventions when combined with sewer rehabilitation
644 schemes across future scenarios.

645 Green retrofits provide consistent levels of robustness under a variety of scenarios, achieving low-
646 regret robustness for reliability, resilience and sustainability altogether. The mitigation of urban creep
647 using permeable pavements cannot compensate its cost regret with small operational gains in the
648 IUWWS. Such limitation recommends: 1) the implementation of this strategy in combined sewer areas
649 where this can be integrated with the disconnection of adjacent roofs and paved areas as to enhance its
650 catchment benefits; 2) supporting more cost-effective alternatives for urban creep mitigation in separate
651 areas and elsewhere.

652 In urban wastewater system modeling, a wide range of parameters can be used, including population,
653 urban growth and rainfall intensities. In this study, uncertainties are mapped into and captured within a
654 large discrete scenario space, because of the difficulties in assigning likelihoods or intervals to these
655 parameters in a world of deep uncertainty. In the case study, robustness is calculated using a limited
656 number of future scenarios. However, this study represents a first step to explore how to make a long-

657 term strategic plan for the complex integrated urban wastewater system under deep uncertainty. This is
658 essential to provide insight into which strategies are best suited to which futures, what indicators should
659 be monitored in order to meet future compliances, and which strategy is most robust (i.e. less regrettable)
660 no matter how the future unfolds.

661

662 **ASSOCIATED CONTENT**

663 **Supporting Information (SI).** Case Study Schematic Diagram. Future Scenarios. Assessment Details of
664 Reliability, Resilience and Sustainability. Computation of indicators' relative weights using the AHP
665 method and Pair-Wise Comparison between Objectives Under Different Future Scenarios. Reliability,
666 Resilience and Sustainability Results. Scenario and Robustness Indexes for Reliability, Resilience and
667 Sustainability. This material is available free of charge via the Internet at <http://pubs.acs.org>.

668

669 **Author Contributions**

670 All authors have given approval to the final version of the manuscript.

671

672 **ACKNOWLEDGEMENTS**

673 This study was funded by the UK Engineering and Physical Sciences Research Council through
674 STREAM (EP/G037094/1) with Northumbrian Water Limited, BRIM (EP/N010329/1) and the final
675 author's fellowship Safe & SuRe (EP/K006924/1).

676

677 **REFERENCES**

- 678 1. Zimmerman, J. B., Mihelcic, J. R. & Smith, J. Global Stressors on Water Quality and Quantity.
679 *Environ. Sci. Technol.* 4247–4254 (2008). doi:10.1021/es0871457

- 680 2. Kleidorfer, M., Möderl, M., Sitzenfrei, R., Urich, C. & Rauch, W. A case independent approach on
681 the impact of climate change effects on combined sewer system performance. *Water Sci. Technol.* **60**,
682 1555–64 (2009).
- 683 3. Urich, C. & Rauch, W. Exploring critical pathways for urban water management to identify robust
684 strategies under deep uncertainties. *Water Res.* **66**, 374–389 (2014).
- 685 4. Ferguson, B. C., Frantzeskaki, N. & Brown, R. R. A strategic program for transitioning to a Water
686 Sensitive City. *Landscape Urban Plan.* **117**, 32–45 (2013).
- 687 5. Butler, D., Ward, C., Sweetapple, C., Astaraie-Imani, M., Diao, K., Farmani, R., & Fu, G. Reliable,
688 Resilient and Sustainable Water Management: The Safe & SuRe Approach. *Glob. Chall.* 1–15
689 (2016). doi:10.1002/gch2.1010
- 690 6. Xu, L., Marinova, D. & Guo, X. Resilience thinking: a renewed system approach for sustainability
691 science. *Sustain. Sci.* **10**, 123–138 (2014).
- 692 7. Ward, S. & Butler, D. Rainwater Harvesting and Social Networks: Visualising Interactions for Niche
693 Governance, Resilience and Sustainability. *Water* **8**, 526 (2016).
- 694 8. Blackmore, J. & Plant, R. Risk and resilience to enhance sustainability with application to urban
695 water systems. *J. Water Resour. Plan. Manag.* **134**, 224–233 (2008).
- 696 9. Mugume, S. N., Gomez, D. E., Fu, G., Farmani, R. & Butler, D. A global analysis approach for
697 investigating structural resilience in urban drainage systems. *Water Res.* **81**, 15–26 (2015).
- 698 10. Juan-García, P. Butler, D., Comas, J., Darch, G., Sweetapple, C., & Thornton, A. Resilience theory
699 incorporated into urban wastewater systems management . State of the art. *Water Res.* **115**, 149–161
700 (2017).
- 701 11. Scholten, L., Maurer, M. & Lienert, J. Comparing multi-criteria decision analysis and integrated
702 assessment to support long-term water supply planning. *PLOS ONE* **12**, e0176663 (2017).
- 703 12. Kousky, C., Olmstead, S. M., Walls, M. A. & Macauley, M. Strategically Placing Green
704 Infrastructure : Cost-Effective Land Conservation in the Floodplain. *Environ. Sci. Technol.* **47**, 3563–
705 3570 (2013).

- 706 13. Eggimann, S., Truffer, B. & Maurer, M. To connect or not to connect? Modelling the optimal degree
707 of centralisation for wastewater infrastructures. *Water Res.* **84**, 218–231 (2015).
- 708 14. Casal-Campos, A., Fu, G., Butler, D. & Moore, A. An Integrated Environmental Assessment of
709 Green and Gray Infrastructure Strategies for Robust Decision Making. *Environ. Sci. Technol.* **49**,
710 8307–8314 (2015).
- 711 15. Schütze, M. R., Butler, D. & Beck, M. B. *Modelling, Simulation and Control of Urban Wastewater*
712 *Systems*. (Springer, 2002).
- 713 16. Ifak. *Simulation of Sewer Systems Integrated in SIMBA. SIMBA 6.0 Sewer User's Guide*. (Institute
714 for Automation and Communication, 2007).
- 715 17. Butler, D. & Schütze, M. R. Integrating simulation models with a view to optimal control of urban
716 wastewater systems. *Environ. Model. Softw.* **20**, 415–426 (2005).
- 717 18. Fu, G., Butler, D. & Khu, S.-T. Multiple objective optimal control of integrated urban wastewater
718 systems. *Environ. Model. Softw.* **23**, 225–234 (2008).
- 719 19. Fu, G., Khu, S.-T. & Butler, D. Optimal Distribution and Control of Storage Tank to Mitigate the
720 Impact of New Developments on Receiving Water Quality. *J. Environ. Eng.* **136**, 335–342 (2010).
- 721 20. The Government Office for Science (UK-GOS). *International Dimensions of Climate Change*. (UK-
722 GOS, 2011).
- 723 21. Environment Agency. *Planning ahead for an uncertain future. Water in the 2050s*. (UK Environment
724 Agency, 2009).
- 725 22. Environment Agency. *Water: Planning ahead for an uncertain future. Water in the 2100s*. (UK
726 Environment Agency, 2010).
- 727 23. European Environment Agency. *Projected change in annual and summer precipitation*. 1–7
728 (European Environment Agency, 2014).
- 729 24. European Environment Agency. *Key findings - Climate change, impacts and vulnerability in Europe*
730 *2016*. (European Environment Agency, 2017).

- 731 25. Ofwat. *Future Impacts on Sewer Systems in England and Wales. Summary of a Hydraulic Modelling*
732 *Exercise Reviewing the Impact of Climate Change, Population and Growth in Impermeable Areas up*
733 *to Around 2040.* (Ofwat, 2010).
- 734 26. Panagos, P. Ballabio, C., Meusburger, K., Spinoni, J., Alewell, C., & Borrelli, P. Towards estimates
735 of future rainfall erosivity in Europe based on REDES and WorldClim datasets. *J. Hydrol.* **548**, 251–
736 262 (2017).
- 737 27. UKCP09. UK Climate Projections: annual mean precipitation projections, 2020s, 2050s and 2080s.
738 *UK Climate Projections* (2014). Available at:
739 <http://ukclimateprojections.metoffice.gov.uk/23995?emission=medium>. (Accessed: 22nd February
740 2018)
- 741 28. Evans, E. P., Ashley, R., Hall, J.W., Penning-Rowsell, E.C., Saul, A., Sayers, P.B., Thorne, C.R., &
742 Watkinson, A.R. *Future Flooding Volume 1: Future Risks and their Drivers.* (UK Government
743 Office for Science, 2004).
- 744 29. Farmani, R. Butler, D., Hunt, D.V.L., Memon, F.A., Abdelmeguid, H., Ward, S., & Rogers C.D.F.
745 Scenario-based sustainable water management and urban regeneration. *ICE Publ.* **165**, 89–98 (2012).
- 746 30. Lombardi, D. Rogers, C.D.F, Aston, R., Barber, A., Boyko, C., Brown, J., Bryson, J., Butler, D.,
747 Caputo, S., Caserio, M., Coles, R., *et al.* *Designing Resilient Cities: A Guide to Good Practice.* (IHS
748 BRE Press, 2012).
- 749 31. Makropoulos, C. K., Memon, F. A., Shirley-Smith, C. & Butler, D. Futures: an exploration of
750 scenarios for sustainable urban water management. *Water Policy* **10**, 345–373 (2008).
- 751 32. Berkhout, F. & Hertin, J. Foresight Futures Scenarios. Developing and applying a participative
752 strategic planning tool. *Greener Manag. Int.* **37**, 37–52 (2002).
- 753 33. Makropoulos, C. K. & Butler, D. Distributed water infrastructure for sustainable communities. *Water*
754 *Resour. Manag.* **24**, 2795–2816 (2010).
- 755 34. Libralato, G., Volpi Ghirardini, A. & Avezzi, F. To centralise or to decentralise: An overview of the
756 most recent trends in wastewater treatment management. *J. Environ. Manage.* **94**, 61–68 (2012).

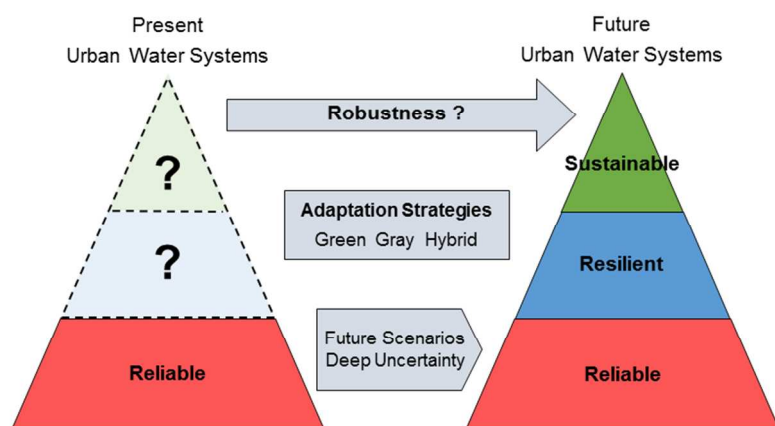
- 757 35. Dobbie, M. F., Brown, R. R. & Farrelly, M. A. Risk governance in the water sensitive city:
758 Practitioner perspectives on ownership, management and trust. *Environ. Sci. Policy* **55**, 218–227
759 (2016).
- 760 36. Poustie, M. S., Deletic, A., Brown, R., Wong, T., De Haan, F.J., & Skinner R. Sustainable urban
761 water futures in developing countries: the centralised, decentralised or hybrid dilemma. *Urban Water*
762 **12**, 543–558 (2014).
- 763 37. Scholes, L., Revitt, D. M. & Ellis, J. B. A systematic approach for the comparative assessment of
764 stormwater pollutant removal potentials. *J. Environ. Manage.* **88**, 467–478 (2008).
- 765 38. Hashimoto, T., Stedinger, J. R. & Loucks, D. P. Reliability, resiliency, and vulnerability criteria for
766 water resource system performance evaluation. *Water Resour. Res.* **18**, 14–20 (1982).
- 767 39. Bao, Y. & Mays, L. W. Model for water distribution system reliability. *J. Hydraul. Eng.* **116**, 1119–
768 1137 (1991).
- 769 40. Ahern, J. From fail-safe to safe-to-fail: Sustainability and resilience in the new urban world. *Landsc.*
770 *Urban Plan.* **100**, 341–343 (2011).
- 771 41. Wang, C. & Blackmore, J. M. Resilience Concepts for Water Resource Systems. *J. Water Resour.*
772 *Plan. Manag.* **135**, 528–536 (2009).
- 773 42. Hwang, H., Lansey, K. & Quintanar, D. R. Resilience-based failure mode effects and criticality
774 analysis for regional water supply system. *J. Hydroinformatics* **17**, 193–210 (2015).
- 775 43. Hwang, H., Forrester, A. & Lansey, K. Decentralized Water Reuse: Regional Water Supply System
776 Resilience Benefits. *Procedia Eng.* **70**, 853–856 (2014).
- 777 44. Hallegatte, S., Shah, A., Brown, C., Lempert, R. & Gill, S. Investment Decision Making Under Deep
778 Uncertainty: Application to Climate Change. *World Bank Policy Res. Work. Pap.* 6193 (2012).
779 doi:doi:10.1596/1813-9450-6193
- 780 45. Lempert, R. J., Popper, S. W. & Bankes, S. C. *Shaping the Next One Hundred Years: New Methods*
781 *for Quantitative, Long-Term Policy Analysis.* (RAND Corporation, 2003).
782 doi:10.1016/j.techfore.2003.09.006

- 783 46. Lempert, R. J. & Schlesinger, M. E. Robust strategies for abating climate change. *Clim. Change* **45**,
784 387–401 (2000).
- 785 47. Lempert, R., Groves, D., Popper, S. & Bankes, S. A general, analytic method for generating robust
786 strategies and narrative scenarios. *Manag. Sci.* **52**, 514–528 (2006).
- 787 48. Lempert, R. J. & Collins, M. T. Managing the risk of uncertain threshold responses: comparison of
788 robust, optimum, and precautionary approaches. *Risk Anal.* **27**, 1009–26 (2007).
- 789 49. Savage, L. J. *The Foundations of Statistics*. (Wiley & Sons, 1954).
- 790 50. Willows, R., Reynard, N., Meadowcroft, I. & Connell, R. *Climate adaptation : Risk, uncertainty and*
791 *decision-making. Part 2*. (UKCIP Technical Report, 2003).
- 792 51. Keeney, R. L. & Raiffa, H. *Decisions with multiple objectives: Preferences and value tradeoffs*.
793 (John Wiley & Sons, 1976).
- 794 52. Sadr, S. M. K., Mashamaite, I., Saroj, D., Ouki, S. & Ilemobade, A. Membrane assisted technology
795 appraisal for water reuse applications in South Africa. *Urban Water J.* **16**, 1–17 (2016).
- 796 53. Aven, T. On Some Recent Definitions and Analysis Frameworks for Risk, Vulnerability, and
797 Resilience. *Risk Anal.* **31**, 515–522 (2011).
- 798 54. Scholz, R. W., Blumer, Y. B. & Brand, F. S. Risk, vulnerability, robustness, and resilience from a
799 decision-theoretic perspective. *J. Risk Res.* **15**, 313–330 (2012).
- 800 55. Ayyub, B. M. Systems Resilience for Multihazard Environments: Definition, Metrics, and Valuation
801 for Decision Making. *Risk Anal.* **34**, 340–355 (2014).
- 802 56. Schoen, M. Hawkins, T., Xue, X., Ma, C., Garland, J., & Ashbolt, N.J. Technologic resilience
803 assessment of coastal community water and wastewater service options. *Sustain. Water Qual. Ecol.* **6**,
804 75–87 (2015).
- 805 57. Xue, X., Schoen, M.E., Ma, X., Hawkins, T.R., Ashbolt, N.J., Cashdollar, J., & Garland J. Critical
806 insights for a sustainability framework to address integrated community water services: Technical
807 metrics and approaches. *Water Res.* **77**, 155–169 (2015).

- 808 58. Murphy, H. M., Meng, Z., Henry, R., Deletic, A. & McCarthy, D. T. Current Stormwater Harvesting
809 Guidelines Are Inadequate for Mitigating Risk from *Campylobacter* During Nonpotable Reuse
810 Activities. *Environ. Sci. Technol.* **51**, 12498–12507 (2017).
- 811 59. Schoen, M. E., Xue, X., Hawkins, T. R. & Ashbolt, N. J. Comparative Human Health Risk Analysis
812 of Coastal Community Water and Waste Service Options. *Environ. Sci. Technol.* **48**, 9728–9736
813 (2014).

814

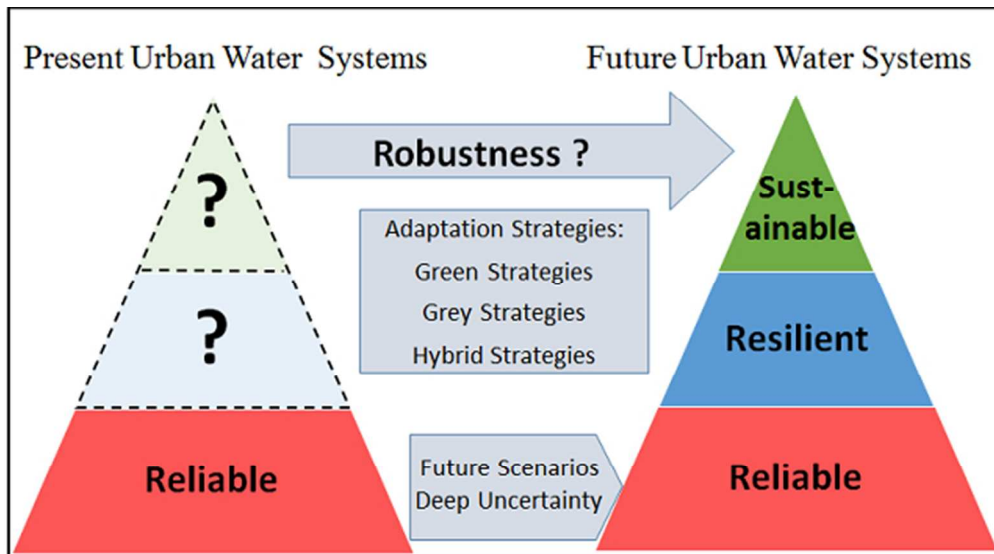
815

816 **TOC/Abstract ART**

817

818

819



169x93mm (96 x 96 DPI)