Applying and advancing the economic resource scarcity potential (ESP) method for rare earth elements

Abstract
A number of studies have identified rare earth elements (REE) as critical metals due to their high economic importance combined with a high risk of supply disruption (Du et al, 2011; Nassar et al, 2015; Schneider et al, 2014). The current methods used to calculate resource depletion in life cycle assessments (LCA) neglect socio-economic, regulatory and geopolitical aspects, nor do they include functionalities such as material recycling or reuse that control the supply of raw materials. These are important factors in determining criticality and are the controlling factors on REE availability rather than geological availability. The economic scarcity potential (ESP) method introduced by Schneider et al. (2014) provides a framework to calculate criticality. This paper reviews the ESP method and advances the method based on recent developments in material criticality. ESP criticality scores for 15 REE with the addition of Au, Cu, platinum-group metals (PGM), Fe and Li are measured. The results highlight that Nd and Dy are the most critical REE, owing mainly to the high demand growth forecast for these two elements. A pathway is presented for incorporating these calculated scores into the ReCiPe life cycle impact assessment (LCIA) method of a LCA.

1. Introduction
Life cycle assessment (LCA) is an important tool to quantify the environmental performance of a product or a process such as rare earth element (REE) production. A LCA can detail potential impacts that this process will have on human health, natural environment and natural resources. However there are limitations and problems for assessing abiotic resource depletion during a life cycle impact assessment (LCIA) (Drielsma et al, 2016). Abiotic depletion potential has been used as an indicator, calculating future exhaustion of resources based on current production levels. Advances were made to this approach by Vieira et al. (2016) with the surplus cost potential method, which calculates the increased cost of extracting raw materials due to depleting resources providing a cost per unit of metal extracted in the future. Both methods are useful in understanding the long-term availability of resources but fail to consider a range of factors which control the supply of critical raw materials. In order to correctly assess the criticality of materials, it is necessary to have an
indicator that takes into account several impact categories for supply risk and economic importance rather than just resource depletion. Otherwise, the assessment categorizes cerium (which is as abundant in the crust as copper) as highly critical along with dysprosium, praseodymium and the other heavy REE. This paper examines how an alternative method to assess mineral resource inputs can be devised and used for critical metals such as the REE.

Rare earth elements include the lanthanides and the chemically similar elements yttrium (Y) and scandium (Sc). The elements are often divided into two groups, the light rare earths elements (LREE) and heavy rare earth elements (HREE). The LREE include La, Ce, Pr, Nd, and Sm. The HREE include the elements from Eu to Lu in the Periodic Table as well as Y. The REE have strategic importance, with uses in a number of emerging low-carbon technologies. Specific physical properties of individual REE are necessary for efficient electric vehicles, and direct drive wind turbines, such as Nd in NdFeB high strength magnets. The addition of Dy is used to maintain the performance of these magnets at high temperatures. Other REE such as La and Ce are used in catalysts for fluid catalytic cracking of crude oil and production of transportation fuels; and Ce and La are used as emissions catalysts in petrol fueled vehicles. Total industrial demand of REE, excluding Y, is small with an estimated use of 159,500 tonnes in 2016 (USGS, 2016), but REE have a large positive economic contribution to downstream industries. One of the major challenges of REE supply is ‘the balance problem’; the misbalance between the economic market demand and the supply of individual REE\(^8\). There is often high demand for REE that are minor constituents of a REE ore (such as Pr), while the demand for the major constituents (such as La and Ce) may be much lower.

The security of supply of REE has been a concern for import-dependent industrialized countries with ambitions to advance their low-carbon economy. China currently dominates the production of REE, excluding Y, accounting for 88% of total REE production in 2016 (USGS, 2016). There is a history of supply disruption of REE exports, this has fueled increased attention into the future availability of such elements. From 2007 to 2009 China reduced export quotas of REE by 25% (Binnemans et al, 2015). This resulted in significant price increases following the export restrictions which were put in place by China (Mancheri, 2015). Concerns about the future supply of REE and the monopolistic nature of production combined with the growing economic importance of downstream products has led to a number of studies identifying individual REE, or REE as a single group, as critical materials (NRC, 2008; Erdmann and Graedel, 2011; Nassar et al, 2015; BGS, 2015; Moss, 2013l Coulomb, 2015; Glöser et al, 2015).
A number of projects exist in various stages of development around the world that if moved into production would diversify the supply of REE. For example mining projects are in the prefeasibility or feasibility stage in Europe, with Sweden’s Norra Kärr project; in Africa with Malawi’s Songwe Hill, Namibia’s Lofdal, and South Africa’s Zandkopsdrift; in North America with Canada’s Ashram, and Nechalacho, USA’s Bear Lodge; Australia’s Nolans, Dubbo Zirconia project; South America has projects such as Araxá and Serra Verde, both in Brazil. However, there are a number of barriers making production outside China challenging. China currently possesses excess production capacity within the country, suppressing prices and reducing the chances of projects outside China from accessing funding. There is also a lack of proven processing technologies for the unconventional mineralogy in some of the new prospects and a lack of efficient and clean technology for separating and converting rare earth oxides to metals and alloys (USGS, 2018). These factors mean that a large amount of time and capital are required to bring in new operations online and diversify the supply.

Downstream uses of REE are often considered to have positive environmental impacts when they are used in generating clean energy or replacing conventional combustion engines in cars (Girardi, 2015). However, the mining, isolation and recovery of REE has a number of environmental and social impacts throughout the life-cycle (Zaimes et al, 2015, Koltun and Tharumarajah, 2014, Arshi et al, 2018, Du and Graedel, 2011, Haque et al, 2014, Sprecher et al 2014).

REE production and processing requires a large amount of energy and chemicals, and can produce greenhouse gas emissions, chemical pollutants, hazardous mine waste and wastewater, which can contain radioactive material and can cause extensive land transformation. Chemicals used in the refining process have been involved in REE bioaccumulation and pathological changes in local residents (Li et al, 2013). Contaminants associated with REE production, which include radionuclides and heavy metals, have been identified as having negative impacts on human, plant and livestock health (Rim, 2016).

It is important to understand and manage the environmental and social costs associated with REE production as we progress to a low-carbon economy and renewable energy generation, which is likely to require more metal and mineral raw materials per unit energy produced. When considering the sustainability of the raw materials that are produced for the low-carbon economy, it is important to consider risks to supply disruption, which could include market imbalances or governmental interventions such as export bans.
The aim of this paper is threefold. (i) To show that individual REE have unique supply risks and economic importance and therefore different levels of criticality. (ii) To provide a more appropriate impact category within LCIA for resource scarcity of critical metals (iii) Explain how criticality can be included in LCA frameworks and see what results would look like.

2. Review of REE criticality studies.

A variety of methodologies can be used to determine raw material criticality. The approaches may vary but share a common aim to define the supply risk of a raw material and its relative importance to the economy. The criticality calculation methodology typically contains an evaluation of the level of supply risk and the impact of said supply risk in a two-dimensional matrix (NRC, 2008; Erdmann and Graedel, 2011; Graedel et al, 2015). Environmental impacts can be used to create a third axis (Graedel, 2015).

Criticality studies are context dependent and can be carried out on a range of scales and for a range of stakeholders, which can be anything from a single company or technology, to a national or multi-national economy (Graedel et al, 2012). For example, a criticality study from the perspective of a country will be different from that of a company, and short-term risk of raw material criticality may not be the same in the medium or long-term. Criticality studies are connected to the concept of risk theory in a holistic way, including economic, societal or environmental risk (Helbig et al, 2016; Frenzel et al, 2017). A wide variety of factors are often considered in criticality assessments, including geological deposits, geographical concentration of deposit or processing facilities, social issues, regulatory structure, geopolitics, environmental issues, recycling potential, substitutability, and sustainability (Achzet and Helbig, 2013; Erdmann and Graedel, 2011).

Eight studies that include criticality of REE have been reviewed (Figure 1). Each study had a different context, with various spatial scales, from national to international and looked at different areas of the economy. For example Nassar et al. (2015) looked at the criticality of REE associated with the global economy, whilst Coulomb examined the criticality of REE in the context of the low-carbon economy. Where possible the studies looked at a medium-term time perspective of criticality.

All but one study (BGS, 2015) included two-dimensions typical of criticality studies which could be translated into supply risk and economic importance. Figure 1 shows the supply risk of the REE on the left hand side of each box and to the right shows economic
importance of the REE from these studies. The relative criticality scores are normalized and given a colour scale between 1 (non-critical) to 6 (extremely-critical). The terms used in the study also varied meaning that this approach includes subjective judgement of the criticality scores. The white categories indicate gaps in the criticality study.

Figure 1. Criticality assessments for individual REE based on supply risk (green top half of each) and economic importance (blue bottom half) at various scales from national to global in a medium term time scale. White space means that the REE was not included in the criticality study (NRC, 2008; Erdmann and Graedel, 2011; Nassar et al, 2015; BGS, 2015; Moss, 2013; Coulomb, 2015; Glöser et al, 2015).
2.1. Life cycle impact indicators for abiotic resource depletion.

The concept of the Area of Protection was founded in the early 1990s by the Society of Environmental Toxicology and Chemistry (Fava et al, 1993). It is used in the LCA community to identify classes of endpoint category indicators that society deems important to protect, and allows a linkage between damages because of environmental intervention and societal values. The Area of Protection are divided into the protection of: Human Health, the Natural Environment and Natural Resources (Finnveden, 1997; Udo de Haes et al, 1999). The ILCD handbook defines these natural resources and that challenge as;

“The concern of natural resources is the removal of resources from the environment (and their use) which results in a decrease in the availability of the total resource stock, as non-renewable (usually abiotic) resources are finite”

This definition and the depletion of abiotic resources is a much disputed category within LCA as it crosses the economy-environment system boundary in combination with the
fact that there are different ways to define the depletion problem, and there are different ways 
of calculating these depletion definitions (Van Oers and Guinée, 2016).

For example Van Oers (2016) stated that the environmental impact of LCA should not 
strive to take into account the different aspects of a criticality assessment due to the varying 
temporal and spatial nature of each study. However this can be overcome with a clear 
definition during the goal and scope phase of a LCA and matching the criticality calculation 

to what is being measured. For example if the environmental performance of a mining project 
is being measured, it is possible to complete the criticality calculation for the life of the 

mining project with criticality scores in a global context.

Different approaches can be used to determine the decreasing availability of 
resources. Different approaches have distinct visions or cultural perspectives for abiotic 
resource depletion (De Schryver et al, 2018). The cultural perspective theory which has 
categorised visions on resource depletion as either individualist, hierarchist and egalitarian is 
explored is incorporated into different LCIA methodologies.

One approach to resource depletion which aims to remove the cultural perspective 
from the process is through the use of entropy or exergy as a basis for characterization, which 
considers the efficiency of extraction. A thermodynamic approach which can capture 
resources is a useful approach as is it has an established scientific basis. Exergy is a measure 
of available energy, whilst entropy in this context refers to the dispersal of energy within a 
system.

A common method that has been used and is considered individualist uses resource 
scarcity for the basis of characterization. This method calculates the long-term depletion of 
non-renewable resources. The depletion of resources is calculated and considers future 
resource scarcity as a result of current consumption. The impact from resource use is then 
calculated as an impact on human welfare due to reduced availability, increased competition, 
and limited accessibility driven by social and geopolitical factors (Finnveden, 2005; 
Sonnemann et al, 2015). These approaches have shortcomings. Firstly, calculations of 
physical resource availability or ‘depletion potential’ used in LCIA rely on a fixed stock 
paradigm, as described by Tilton (2002). The idea that there is a finite quantity of a resource, 
often described as a crustal abundance, fails to calculate the reuse or recycling rate of these 
materials and considers that materials are lost after use. There is also no clear definition for 
undiscovered resources (Vieira et al, 2016). The alternative method used is the opportunity 
cost paradigm, which states that if physical quantities reduce, or are more difficult to access, 
prices will increase and innovations and alternatives to that material will be sought, reducing
demand. LCIA practitioners have used both methods which have very different views on natural resources and can significantly alter LCIA results. In the fixed stock method, any use of natural resources results in reduced availability for the future, whereas in the opportunity cost view, natural resources are viewed as flows that need to be managed to meet human demands (Drielsma et al, 2016).

Different methods have different visions and methodologies. Many of these methods that are currently employed to not consider the socio-economic, regulatory and geopolitical aspects or functionalities such as material recycling or reuse.

3. Materials and Methods

The abiotic depletion potential method (Van Oers and Guinée, 2016) and the surplus cost potential method (Vieira et al, 2016) are used for comparison in this paper. The latter has been integrated into the ReCiPe methodology (Huijbregts et al, 2016). This method to calculate metal depletion provides scores for 75 mineral resources providing impact scores in relation to 1kg of Cu. Figure 2 provides a comparison of five mineral resources and categorizes rare earth elements as a single group.

![Figure 2. Mineral resource scarcity results (individualist) using the surplus cost potential approach](image)

LCIA is a step in a LCA which translates data such as emissions or resource uses from LCA studies to an easily understandable smaller number of impact scores. The method
of calculating these scores is referred to as characterization, and the results will produce an environmental impact per unit of stressor (e.g. per kg of resource). Schneider et al. (2014) identified that economic aspects of resource supply are neglected in current LCA methodologies and attempted to overcome this by introducing the economic resource scarcity potential (ESP) model.

Various data that contribute to scarcity of resources are included, expanding the Area of Protection for natural resources to include economic or socially derived scarcity. The factors that are included in ESP include reserves, recycling, and country and company concentration of mining activities, economic stability, demand growth, trade barriers, and companion metal fraction. Drielsma et al. (2016) highlighted that this method assesses short term availability of resources, and is a useful tool in identifying disruptions that may arise in this timeframe. Drielsma et al. (2016) also argued that the Area of Protection for natural resources is altered using this method as the ESP method aims to protect the product system being measured rather than the resources themselves. For example, the protection of the value that a resource has when being used rather than the resource itself.

Current LCIA methods, such as the ReCiPe approach only take into account geological availability and the increased cost of accessing raw materials in the future. The surplus cost potential method fails to take into account resource criticality. Additional methods, such as the ESP approach, would be a useful step to incorporate criticality factors into the life cycle sustainability assessment framework which would better represent impacts on the Area of Protection for Natural Resources (Sonnemann et al, 2015). The ESP method put forward by Schneider (2014) allows for a new characterization factor for resource use impact assessment. Using these characterization factors and a framework to incorporate criticality into the life cycle sustainability assessment context by Sonnemann et al. (2015) allows for integration of the ESP method into the LCA.

3.1. Methodology of ESP Calculations.

The factors that impact resource availability were suggested by Schneider et al. (2014) and have been highlighted in table 1. Equal weighting was used for all impact categories initially replicating the method used by Schneider (2014). This was followed by a comparison of results if the economic importance impact category was increased to represent 50 percent of the total ESP score. Production data were obtained by combining the USGS data with other project scale information. Individual REE data were obtained from individual
companies, and when not possible were estimated from literature. All sources of information and origins of data used in the study are included in the supplementary information.

Table 1. Overview of impact categories, indicators and thresholds used in the ESP calculations

<table>
<thead>
<tr>
<th>Impact category</th>
<th>Category indicators</th>
<th>Threshold</th>
</tr>
</thead>
<tbody>
<tr>
<td>Supply risk</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Reserve availability</td>
<td>Reserve/Annual production</td>
<td>Low&lt;0.4&lt;high</td>
</tr>
<tr>
<td>Recycling</td>
<td>New material content (%)</td>
<td>Low&lt;0.5&lt;high</td>
</tr>
<tr>
<td>Mining country concentration reserves</td>
<td>HHI index</td>
<td>Low&lt;0.15&lt;high</td>
</tr>
<tr>
<td>Production bottleneck (country concentration)</td>
<td>HHI index</td>
<td>Low&lt;0.15&lt;high</td>
</tr>
<tr>
<td>Production bottleneck (company concentration)</td>
<td>HHI index</td>
<td>Low&lt;0.15&lt;high</td>
</tr>
<tr>
<td>Governance stability</td>
<td>WGI(^1)</td>
<td>Low&lt;0.25&lt;high</td>
</tr>
<tr>
<td>Socioeconomic stability</td>
<td>HDI(^2)</td>
<td>Low&lt;0.12&lt;high</td>
</tr>
<tr>
<td>Trade barriers mine production</td>
<td>Share of mine production under trade barriers (%)</td>
<td>Low&lt;0.25&lt;high</td>
</tr>
<tr>
<td>Companion metal fraction</td>
<td>Production as companion metal (%)</td>
<td>Low&lt;0.2&lt;high</td>
</tr>
<tr>
<td>Trade</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Economic importance</td>
<td>Average production and cost per kg</td>
<td>Low&lt;0.1&lt;high</td>
</tr>
</tbody>
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The data incorporate 10 impact categories and can be aggregated to provide a single ESP value (Equation 2). Each category has been described in a glossary in the supplementary information. This allows for the comparison of the 15 REE studied as well as providing a comparison with Au, Cu, PGM, Fe and Li. Other elements were selected because they offered a range of supply risk and economic importance scores in previous criticality studies. They are used for comparison with the REE and to give a context to how REE perform. The criticality in the context of this paper is within a “global economy” and so not specific to a particular technology or group. This also allows for integration within the ReCiPe LCIA as this is on a global scale. It should be noted that it is possible to adjust the context through weighting the results or changing the thresholds. Thresholds used in this study are shown in Table 1 with justification for their values.

The aggregation of the supply risk and economic importance impact factors is given equal weighting. Individual category indicator results (impact factor x LCI) give an indication for
the magnitude of the risk. However, the results only provide a comparison of the resources
studied. A greater number of resources used for this method will allow for a more
comprehensive estimation of supply risk and provide a better basis for decision making.

As noted by Schneider (2014), to produce a supply risk perspective for the resource
availability requires each category indicator to be placed in relation to a target. This method
is described in detail by the distance-to-target method by Frischknecht et al. (2008). The
resulting impact factors ($I$) provide a threshold, above which high risk of supply disruption is
expected. This was calculated for comparison for the 15 REE together with gold, copper,
platinum group metals (PGM), iron ore and lithium ($i$) and each impact category ($j$). The
ratio of current to critical flows is squared allowing large impact values (above the target
value) to be weighted above proportional (Frischknecht and Büsser Knöpfel, 2013; Drielsma
et al., 2014). The indicators are scaled from 0-1, with order being inverted when necessary to
ensure high score corresponds to high risk. All values below the value of “1” are deemed
uncritical and have no impacting score.

\[
I_{i,j} = \text{Max} \left\{ \left( \frac{\text{indicator value}_{i,j}}{\text{threshold}_{i,j}} \right)^2 ; 1 \right\}
\]

Equation 1.

\[
\text{ESP}_i = \prod_j (I_{i,j})
\]

Equation 2.

The resulting economic scarcity potential score for each element which includes the
impact categories from both supply risk and economic importance is a dimensionless quantity
determined by the ratio of the current indicator value to the determined threshold linked to
the LCI.

4. Results
The performance of individual REE compared to Au, Cu, PGM, Fe and Li has been calculated and highlighted in Figure 3.

Figure 3. Individual impact category scores for 10 categories. Data based on (Buijs et al, 2012; NRC, 2016; Graedel et al, 2015; Nassar et al, 2015; Angerer et al, 2009).

4.1. Reserve availability

The 15 REE included in the study had a lower score for reserve availability than Au, Cu, PGM, Fe and Li. These other metals had higher impact scores because of their high level of production relative to REE; being produced in thousands or millions of tonnes per annum compared to REE which have a total production of the 126,000 tonnes in 2016. This, combined with the large reserves of REE, calculated as 120,000,000 (USGS, 2016) t based on their continued availability and typical metallurgical recoveries means the reserve availability of REE is higher than the other metals in the study leading to a low impact score. Of the REE, Y, Gd, Tb, Dy and Ho had the highest impact score whilst Ce and La had the lowest. These results can be explained by the fact that HREE are less abundant in the earth’s crust and also less abundant in REE deposits, whilst consumption of some of these elements remains relatively high, such as Dy and Tb in permanent magnets. Er, Tm, Yb and Lu are not abundant in deposits but are exploited at very low rates leading to a moderate impact score.

4.2. Recycling

More work needs to be carried out to quantify the rate of recycling of different REE because the published data used for the calculations in this study does not represent the quantity of recycled material reentering the system.
4.3. Country concentration of reserves

The country concentration of reserves impact score was high for PGM compared to the other raw material in this study. This is because of the dominance of South Africa in holding the reserves of PGM. In contrast reserves of Au, Cu and Fe appear the most widespread as they have the lowest score in this category. The REE had moderate scores in this area with slightly increasing impact scores of the HREE because of the dominance of China in holding much of the HREE reserves. The country concentration of reserves indicated that although the reserves of rare earths are relatively widespread, there is a high concentration of Sm and Eu in China, whilst Ho, Er, Tm, Yb and Lu in reserves is more geographically widespread.

4.4. Country concentration of production

The impact score for the country concentration of production was high for all REE compared to Au, Cu, PGM, Fe and Li owing to the dominance of REE production from China. The HREE had the highest impact score for this section. Li was highest scoring in this category for the non-REE.

4.5. Company concentration of mine production

The company concentration of mine production impact category displays the dominance of Northern Rare Earth (Group) High-Tech Co., Ltd, China even when put in in the context of other raw materials, with Ce, Pr, Nd, Sm, Eu, Gd, Tb having the highest scores for this section. The lower impact score for the LREE can be explained by production from Lovozerskiy GOK in Russia, Mount Weld in Australia and mineral sands in India, which are all LREE-enriched deposits.

4.6. Governance stability

The impact scores were high for the REE, with highest scores being seen with the HREE that are produced almost exclusively in China. Li had a low impact score in this category is explained by its production in Australia and Chile. PGM and Au had high scores in these categories highlighting that there are risks associated with the stability of governments in regions where these materials are mined.
4.7. Socioeconomic stability

Au was the highest scoring element, followed by Cu and then the Ce, Pr, Nd, Sm, Eu, Gd, Tb. The low socioeconomic stability of the countries producing Au are highlighted as well as the moderate socioeconomic score of China. For REE the lowest impact scores were Ce and Eu. This is owing to the combination of elevated levels of production of these elements from Mt Weld, Australia and Australia’s higher performance in government stability and socioeconomic stability.

4.8. Policy potential index

The 15 REE studied had a high score for the policy potential index. However it is PGM that had the highest score in this category, whilst Fe had a similar score to the REE. The policy potential index impact score was the highest for Tb, whilst Ho had the lowest score. Many of the REE received moderate scores in this impact category indicating that there was only a small amount of variation in the impact scores for the REE.

4.9. Companion metal fraction

REE have a high risk associated with the fact that they are commonly exploited as a by-product of each other and of other raw materials (such as iron ore at Bayan Obo, China) among others. The other raw materials used in comparison had low impact scores in this category indicating that they are commonly extracted as the main component at a mine. The companion metal fraction impact scores were relatively similar to each other. Pr had the highest score whilst Y had the lowest.

4.10. Economic importance

In the economic importance category the REE have a low score. This category is dominated by Cu and to a lesser extent Au. These are the two raw materials that have the greatest economic importance during the raw material extraction phase. Of the REE, Nd had a markedly higher economic importance impact score than the other REE. This is owing to the use of Nd in NdFeB magnets, which are predicted to drive demand growth until 2022 (Roskill, 2016). Dy and Pr were calculated as having the next highest economic importance scores. All other REE have low economic importance scores.
4.2. Overall ESP

The final ESP results are presented on a logarithmic scale to better display the relative performance of individual elements. The ESP scores displayed in Figure 4 show how the REE compared to Au, Cu, PGM, Fe and Li. Giving equal weighting to each category and using the methodology described above resulted in PGM having the highest ESP score and so these elements are considered the most critical in this context. The factors driving the PGM score up are the high policy potential index score, the high governance stability score as well as a high country concentration of reserves. Dy scores second highest for ESP. It is interesting to note that as a greater number of raw materials are included in the study, the relative performance of elements can change, as in this case where Dy has overtaken Nd in terms of relative ESP score. This is because the economic importance was an important factor in driving Nd’s ESP score up in the REE comparison, but as more raw materials are added with a greater economic importance, this distinction becomes less important. Nd is the next highest scoring element, followed by Tb, Pr, Gd and Y. Au, Cu, Fe and Li all have lower ESP scores than the REE.

The economic scarcity potential approach used in this study provides results that greater reflect the reality of resource availability until 2021 when compared to the abiotic depletion potential or surplus cost potential approach, which are more suited to understanding
the long-term availability of resources. It considers socio-economic, regulatory and
geopolitical aspects or functionalities such as material recycling or reuse in the calculations
rather than geological availability. This is an area that is currently missing in the LCA
approach but has an impact on low-carbon technology development and proliferation. Nd and
Dy are the highest scoring REE using this approach, highlighting the need to broaden the
supply chain for these two elements in particular, whilst Ce has a low economic scarcity
potential score and is overproduced. New uses of Ce, which is cheap because of the
oversupply, would help to even up requirement for REE and help supply of Nd and Dy.

A simplified calculation was used for economic importance, looking only at demand
growth, production volume and value of material produced. Improvements could be made to
this calculation. A novel empirical approach has been presented by Mayer and Gleich (2015)
which looks at risk associated with future price increases of raw materials. The approach
which uses a compounding framework to calculate net present values and volatility is a
potential avenue to include under these calculations which may provide more realistic
economic importance impact scores.

The method used in this study only looks at the impact categories associated with the
mining and dissolution phase and fails to consider the larger production chain of final
products which can be in a number of forms such as rare earth oxides, misch-metals or
separated metals and transport. Future work could look at the different processing stages and
see how this would alter the economic scarcity potential scores for different elements. Recent
work has examined the role of primary processing (first post-mining stage) in the supply risk
of critical metals (Nansai et al, 2017). Understanding the role of different processing stages in
raw material availability is an important area of research, especially for REE production
which has a long and complex production chain. Future work should cover all elements from
the periodic table using the economic scarcity potential approach to calculate scores for the
global economy for the short to medium term. Using improved economic importance
calculations would make the approach a useful addition to the LCIA results. Annual updates
on production would allow the method to be up to date and have practical use.

4.3. Adjusting the weighting of economic importance

Criticality studies are context dependent. The ESP results above use an equal
weighting for each impact category. However, it is possible to adjust the level of an impact
category or categories to represent a different context. Figure 5 shows this with the blue bars indicating the results of the ESP scores with equal weighting for the impact factors. The orange bars calculate the ESP score by giving all the supply risk impact categories (reserve availability, recycling, country concentration of reserves, country concentration of mine production, company concentration of mine production, governance stability, socioeconomic stability, trade barriers to mine production, companion metal fraction) equal weighting and giving the economic importance impact category the same weighting as the combined supply risk impact categories.

Figure 5. Economic scarcity potential scores for calculated using 10 categories for each individual element. Blue bars are ESP scores with equal weighting for the impact factors. The orange bars calculate the ESP score by giving all the supply risk impact categories (reserve availability, recycling, country concentration of reserves, country concentration of mine production, company concentration of mine production, governance stability, socioeconomic stability, trade barriers to mine production, companion metal fraction) equal weighting and giving the economic importance impact category the same weighting as the combined supply risk impact categories.

The results indicate an increased ESP score for Au and Cu, which is the highest scoring element in this context, because of their high economic importance score. A small reduction in the ESP score for PGM, which is the second highest scoring element, and Li which has a small reduction in ESP score. Fe has a large decrease. The REE have a substantial decrease in their ESP score owing to their relatively low economic importance.
using the simple calculation in this study when compared to the other elements. Nd is highest
scooring of the REE, followed by Dy owing to their relative high economic importance
compared to other REE.

Increasing the weighting of economic importance (Figure 5) highlights the flexibility
of criticality studies. For example, giving equal weighting Cu was considered one of the
lowest scoring elements in comparison, but when economic importance was increased to 50% of the total ESP score it became the highest scoring element in the study. Criticality studies
can be used to compare the relative levels of criticality of raw materials in different scenarios,
but these need to be clearly defined. This study used a global spatial scale for the whole
economy and used a medium term time scale, but it is possible to adjust the criteria for a
number of scenarios. The weighting of the impact categories will be different depending on
the context of the study. For example a study of the criticality of raw materials for the low-
carbon economy, would give a higher economic importance to the raw materials used in the
relevant technologies than has been given in this study. A valuable area of research would be
to develop understanding of appropriate weighting for the impact categories under different
scenarios. Understanding the importance of different processes of raw material availability
would be a useful step in developing a robust method and would be important in its
successful integration into the LCA approach.

4.4. Integration into LCA

The scores of the individual elements will be calculated against the reference element
of copper. Figure 6 provides a simulation of resource depletion results using three different
calculation methodologies (economic scarcity potential, surplus cost potential, abiotic
depletion potential) with the example using a 1 kg NdFeB magnet. Simplified inventory data
were used (Jin et al, 2016), and is shown in table 2. A comparison of results is highlighted
results using the abiotic depletion potential approach, the surplus cost potential approach and
the economic scarcity potential approach.

<table>
<thead>
<tr>
<th>Element</th>
<th>Weight %</th>
</tr>
</thead>
<tbody>
<tr>
<td>Fe</td>
<td>66.88</td>
</tr>
<tr>
<td>Nd</td>
<td>18.0</td>
</tr>
</tbody>
</table>
Dy 6.15
Pr 4.6
Cu 0.18

Figure 6. Comparison of resource depletion calculation methodology on the results for the components of NdFeB magnet.

Figure 7. Elemental contribution to resource depletion calculation scores for economic
scarcity potential, surplus cost potential, and abiotic depletion potential for components of NdFeB magnet.

The results show that there is an increased score (kg Cu-eq/kg ore for the economic scarcity potential calculation method. This is because the REE components, Pr, Nd and Dy have a high economic scarcity potential score as elements. Cu is the reference value for all methods which explains the equal score with each method. Fe has a lower score using the economic scarcity potential approach as it has been calculate to have low criticality. Figure 7 highlights how the economic scarcity potential approach places greater emphasis on elements that have higher criticality scores and are more susceptible to supply disruption in the short to medium term. This information could prove useful in comparative LCA when examining the performance of a product and process and provides and additional metric for which to compare. Such as scenario could exist when comparing the environmental performance of two mining operations. Results for environmental performance could be included alongside criticality data for a better comparison.

5. Conclusions

The ESP approach is particularly useful when trying to understand the availability of critical metals. This is important as they play a key role as raw materials for the low-carbon economy. This is important as they play a key role as raw materials for the low-carbon economy. This paper aimed to compare the performance of individual REE and put it in context with other raw materials. The results indicate that REE need to be considered as distinct elements with different criticality associated with each of them. For example Dy and Nd had the highest economic scarcity potential scores, whilst Lu and Ce had the lowest of the REE. One of the reasons for Ce having a low score is its overproduction. The excess availability and low criticality means that companies have an opportunity to find new uses for Ce. For example the Critical Materials Institute have developed aluminum-cerium alloys (Sims et al., 2016). The high scores for Nd and Dy are due to the increase in demand of NdFeB magnets in hybrid and electric vehicles until 2026 (Goodenough, 2017). Whilst projections for Sm, Tm and Lu suggest that growth and production volume will remain low, keeping the economic importance of these elements low. All REE have higher economic scarcity potential scores than Au, Cu, Fe and Li, whilst PGM had the highest score of all the elements included in the study. The high score for PGM was due to its concentration of reserves and production in South Africa, which has a low score in the governance stability
and policy potential index. Although further work needs to be done and more elements need to be included in the method before its integration into LCIA results, this study provides a guideline for the approach.

A major challenge for this approach, as with all raw material studies is the availability of data. An inconsistent amount of data are available for the calculations of the economic scarcity potential impact categories. There is a lack of reliable production data for the REE, and this would also be the case for other raw materials. USGS and BGS are useful sources of data, and they are clear about the uncertainty of some production data. For example the high level of illegal mining in REE in China has been ignored (Rao, 2016).

The development of economic and supply risk indicators that can fit alongside or within LCA should be further explored and methods such as the approach shown here can be considered complimentary to other resource depletion methods currently employed.

6. Acknowledgements

Research was funded by NERC SoS RARE consortium (NE/M011429/1), www.sosrare.org with University of Exeter co-funding for R. Pell’s PhD project. The authors appreciate the funding assistance from the Camborne School of Mines Trust to visit and present this research in USA and gain insight from the Critical Metals team from Yale University and Nedal Nassar and his team at USGS.

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