

1 **Near-future carbon dioxide levels impair the olfactory system of a marine fish**

2 Cosima S. Porteus^{1*}, Peter C. Hubbard², Tamsyn M. Uren Webster³, Ronny van
3 Aerle⁴, Adelino V. M. Canário², Eduarda M. Santos¹, Rod W. Wilson^{1*}

4

5 ¹Biosciences, College of Life & Environmental Sciences, University of Exeter, Exeter,
6 Devon, EX4 4QD, UK

7 ²Centro de Ciências do Mar, Campus de Gambelas, Universidade do Algarve, Faro,
8 Algarve, 8005-139 Portugal

9 ³Biosciences, College of Science, Swansea University, Swansea, Wales, SA2 8PP,
10 UK

11 ⁴Centre for Environment, Fisheries and Aquaculture Science (Cefas), Weymouth,
12 Dorset, DT4 8UB, UK

13 *Corresponding authors:

14 Cosima Porteus, email: cosimaporteus@gmail.com

15 Rod W. Wilson, email: R.W.Wilson@exeter.ac.uk

16 **Survival of marine fishes exposed to elevated near-future CO₂ levels is**
17 **threatened by their altered responses to sensory cues. Here we demonstrate a**
18 **novel physiological and molecular mechanism in the olfactory system which**
19 **helps explain altered behavior under elevated CO₂. We combine**
20 **electrophysiology and transcriptomics with behavioral experiments to**
21 **investigate how elevated CO₂ affects the olfactory system of European sea**
22 **bass (*Dicentrarchus labrax*). Under elevated CO₂ (~1000 μatm) fish must be up**
23 **to 42% closer to an odor source for detection, compared with current CO₂**
24 **levels (~400 μatm), decreasing their chances of detecting food or predators.**
25 **Compromised olfaction correlated with the suppression in the transcription of**
26 **genes involved in synaptic strength, cell excitability, and wiring of the**
27 **olfactory system in response to sustained exposure to elevated CO₂. Our**
28 **findings complement the previously proposed impairment of γ-aminobutyric**
29 **acid receptors, suggesting both the olfactory system and central brain**
30 **function are compromised by elevated CO₂.**

31 Fish rely heavily on their olfaction for finding food^{1,2}, recognizing conspecifics and
32 predators^{1,3,4}, for the perception of reproductive status⁵ and homing towards suitable
33 habitats^{6,7,8,9}, including spawning grounds and larval settlement¹⁰. The predicted
34 end-of-the-century CO₂ levels (800-1000 μatm) cause ocean acidification (elevated
35 H⁺, reduced pH) and have been shown to have negative effects on the sensory-
36 related behavior and learning of many fish species studied to-date^{3,4,11,12,13} including
37 sharks^{14,15}. Moreover, coral reef fishes exposed to elevated CO₂ levels predicted for
38 the year 2100, show a 9-fold increase in mortality, when returned to the wild as
39 compared with fish exposed to current conditions⁴. This suggests the potential for a

40 major ecologically-relevant impairment of fitness of marine fishes as a consequence
41 of exposure to elevated CO₂.

42 To explain the effects of elevated CO₂ on fish behavior, changes in brain
43 neurotransmitter function have been proposed as the sole mechanism
44 responsible^{11,16,17,18}. To date, there has been no consideration of mechanisms
45 operating outside of the central processing of sensory information. A direct effect of
46 seawater pH or CO₂ on the peripheral olfactory system was dismissed in marine
47 fish¹⁹ without any empirical testing. We have challenged this view and hypothesized
48 that elevated CO₂ would directly affect the olfactory system of fish, given that the
49 olfactory epithelium is in intimate contact with sea water. We propose a novel
50 physiological mechanism by which ocean acidification can alter fish behavior and
51 learning directly via the olfactory system, and used the European sea bass as a
52 model to test this hypothesis.

53 **Results**

54 Firstly, we confirmed that the behavior of European sea bass was affected by
55 elevated CO₂ levels as demonstrated in other fish species. Juvenile sea bass were
56 exposed to current (~430 μatm; control) and predicted end of the century levels of
57 CO₂ (~1000 μatm; elevated CO₂) and their behavioral responses to a likely predator
58 odor²⁰ (bile from monkfish, *Lophius piscatorius*; dilution 1:1,000,000) was quantified.
59 Our data demonstrated that sea bass exposed to elevated CO₂ reduced their
60 baseline activity (swimming) by up to 40% compared with control fish (p=0.008, Fig
61 1a). This difference was independent of the duration of exposure (p=0.30).
62 Furthermore, control sea bass reduced their activity by 50% in the presence of the
63 predator odor, whereas sea bass exposed to elevated CO₂ reduced their activity by
64 only 20-27% (P=0.009, Fig. 1b). Both control and elevated CO₂ exposed fish

65 displayed freezing behavior (not moving for more than 5 seconds at a time) after 2
66 and 7 days of exposure (Fig. 1a,b). However, at 14 days, fish exposed to elevated
67 CO₂ spent significantly more time freezing (p=0.043) before and during exposure to
68 a predator cue (Fig. 1e).

69 Having established that end-of-century levels of CO₂ in the water result in
70 pronounced alterations of behavior, we tested if exposure to elevated CO₂ at the
71 olfactory epithelium alone was sufficient to reduce the detection of odorants. We
72 used electrophysiological recordings from peripheral sensory neurons of the
73 olfactory system, allowing us to isolate peripheral olfactory responses from central
74 brain processes. We measured changes in the activity of the olfactory nerve whilst
75 exposing the olfactory epithelium to sea water containing ten different olfactory
76 stimuli dissolved in either control or elevated CO₂ seawater, while fish were
77 maintained under control CO₂ levels. We tested the olfactory nerve response of sea
78 bass to a wide range of odorants: amino acids (L-cysteine, L-serine, L-alanine, L-
79 arginine and L-glutamate), as odorants principally mediating food detection²¹; bile
80 acids involved in chemically-mediated interactions between conspecifics and other
81 teleost species (cyprinol sulphate) and potentially predatory shark species (scymnol
82 sulphate)²²; body fluids (intestinal fluid, bile from conspecifics, and alarm cue), potent
83 chemical signals that can elicit behavioral responses vital for escape
84 from/awareness of predators or recognition of conspecifics^{23,24,25} (see Methods
85 section for a full description). The amplitude of the response indicates the change in
86 magnitude of the nerve activity in response to an odorant, and the detection
87 threshold is defined as the concentration of odorant that produced a detectable
88 response (above baseline). Overall, elevated CO₂ reduced the amplitude of the
89 response for 6 out of the 10 odorants compared to control (Fig. 2), and increased the

90 detection threshold (i.e. reduced sensitivity) in 4 out of the 10 odorants (Fig. 3), but
91 had no effect on the remainder (i.e. elevated CO₂ did not increase amplitude or
92 sensitivity of the response to any of these odorants). Under elevated CO₂ the
93 responses to L-alanine, L-arginine and L-glutamate, cyprinol sulphate, scymnol
94 sulphate and alarm cue were up to 46% lower than those of controls (n=6-11 per
95 odorant per treatment, p=0.027, p=0.040, p=0.028, p=0.011, p=0.012, p=0.018
96 respectively, Fig. 2). The thresholds of detection for L-cysteine (p=0.049), L-alanine
97 (p=0.029) and L-glutamate (p=0.0047), and conspecific bile (p=0.029) were 2 to 5
98 fold higher in elevated CO₂ (Fig. 3). Therefore, under elevated CO₂ these odorants
99 would need to be present in the water at concentrations up to 5 times greater than in
100 current CO₂ conditions in order to be detected by sea bass.

101 The active space of a chemical is defined as the largest volume of water a fish
102 can occupy that still contains a concentration of odorant at or above the olfactory
103 threshold for detection^{25,26}. This is a useful parameter to help estimate how much
104 closer a fish would need to be (on average) to detect an odor-source under these
105 elevated CO₂ conditions. We assumed a homogeneous distribution of the odorant in
106 the water. Although this assumption may be simplistic for most natural environments
107 (due to the constant movement of water associated with currents, tidal movements,
108 etc.), it allows for a good estimate of the average change for most circumstances.
109 The largest reduction in threshold of detection (5 fold) was for glutamate (Fig. 3). We
110 calculated that the active space (i.e. a 3 dimensional volume) for glutamate detection
111 would be reduced by 80%, which would translate into fish having to be 42% closer to
112 the odorant source before detection occurred (i.e. based on the one dimensional
113 radius of the three dimensional active space sphere; Fig. 3 - see Methods for
114 calculation details and assumptions).

115 Lastly, RNA sequencing was used to elucidate the molecular mechanisms
116 underpinning the negative effects of elevated CO₂ on sea bass olfaction. Sea bass
117 were exposed for 2 and 7 days to either control (~450 µatm) or elevated CO₂ (~1000
118 µatm). Global gene transcription was measured in tissue samples taken from the
119 olfactory epithelium (n=6) and the olfactory bulb (n=4) using an Illumina HiSeq 2500
120 platform. *De novo* reference transcriptomes were constructed for each tissue using
121 the Trinity pipeline²⁷ (see Methods for full description). Transcript abundances were
122 calculated using RSEM²⁸ and differences in gene transcription were determined
123 using EdgeR²⁹ and a selection of scripts provided by Trinity.

124 Calcium/calmodulin-dependent protein kinase II beta 2 (CAMKII) directly
125 regulates α-amino-3-hydroxy-5-methyl-4-isoxazolepropionic acid (AMPA) glutamate
126 receptors, known to be involved in synaptic plasticity³⁰. In the olfactory bulb, fish
127 exposed to elevated CO₂ for 2 days showed a significant down-regulation of
128 *camk2ga* (gene encoding CAMKII) and *nptxr* (encoding for a protein involved in
129 synaptic plasticity and the clustering of AMPA receptors). This was followed by the
130 downregulation of *gria1b* (homologous to a gene encoding AMPA glutamate
131 receptors), downregulation of *map2k2a* (gene encoding mitogen activated protein
132 kinase kinase involved in olfactory learning in the olfactory bulb in rats³¹), and the
133 upregulation of *tmub1* (gene encoding a protein involved in the regulation of AMPA
134 receptors at the cell surface) at 7 days. Additionally, genes involved in
135 neurotransmitter re-uptake (*slc6a17* and *slc1a8*) including glutamate were
136 downregulated in the olfactory epithelium in fish exposed to elevated CO₂ for 7 days.
137 In the olfactory epithelium there was an upregulation of *slc4a8* (gene encoding the
138 sodium/bicarbonate cotransporter) at 2 days, similar to that observed for long-term
139 adjustments of bicarbonate (HCO₃⁻) transport in the gills of marine fish in response to

140 elevated CO₂³². Four chemosensory G-protein coupled receptor gene families have
141 been identified in teleosts: main olfactory receptors (ORs), trace amine-associated
142 receptors (TAARs), and vomeronasal receptors 1 and 2 (VR1 and VR2)³³ and all of
143 these were well represented in the olfactory epithelium transcriptome (see
144 Supplementary Table S6). In the olfactory epithelium two OR genes (*or* and *or142*)
145 were significantly down-regulated after 7 days of exposure; however, no OR genes
146 were up-regulated. Interestingly, in the olfactory bulb two additional OR genes (*or*
147 and *or120*), likely expressed in the axons of sensory neurons reaching the olfactory
148 bulb, were also downregulated in fish exposed to elevated CO₂ for 7 days. In
149 mammals, the presence of odorant receptor mRNA in the axons of sensory neurons
150 is well established, and these mRNAs are likely involved in the wiring of the olfactory
151 system³⁴. These results suggest that fish exposed to elevated CO₂ did not activate
152 compensatory molecular mechanisms to adjust for the loss of olfactory sensitivity
153 measured using nerve recording, and that the wiring of the olfactory system might be
154 affected by ocean acidification.

155 Genes involved in excitatory neurotransmission such as the nicotinic
156 acetylcholine receptor (*chrna7*) and glutamate receptor (*gria1b*) were down-
157 regulated in the olfactory bulb of fish exposed to elevated CO₂ levels for 7 days,
158 while a gene involved in decreasing neuronal excitability (calcium-activated
159 potassium channel, *kcnn3*) was up-regulated in the same tissue after 2 days of
160 exposure. In the olfactory epithelium, fish exposed to elevated CO₂ for 7 days
161 showed down-regulation of excitatory neurotransmission (*scn4ab*, *cacna2d*), and
162 neuronal growth and development (*zak* and *efnb2a*). These results indicate
163 mechanisms for decreased excitability of neurons in the olfactory epithelium and

164 bulb at both time points, and thus, a decrease in olfactory information being
165 transmitted to higher brain centers.

166 **Discussion**

167 **Elevated CO₂ affects behavior**

168 Adult sea bass spawn offshore and the newly-hatched larvae must navigate back to
169 the safety of coastal nursery habitats using their olfactory senses to home in on
170 these sites and to avoid predators³⁵, and this life history strategy is shared by many
171 fish species. This is the most vulnerable stage in the life cycle of sea bass being
172 associated with high mortality from predation³⁵. Here, we report that juvenile sea
173 bass exposed to levels of CO₂ predicted for the end of the century demonstrate
174 impaired behavior in response to a predator cue (Fig. 1). This response is similar to
175 those reported for other fish species studied to date and, importantly, an increased
176 effect was observed with exposure time^{3,4,11,12}. However, in sea bass the baseline
177 activity was lower after exposure to elevated CO₂ (Fig. 1), which is unlike most fish
178 species studied to date (generally higher activity was reported after exposure to
179 elevated CO₂)^{36,37}. Although a recent study found no effects of elevated CO₂ on the
180 routine swimming behavior of early juvenile sea bass³⁸, this could have been
181 influenced by much higher control CO₂ conditions (585 μ atm vs 430 μ atm in our
182 study). This higher level of CO₂ (585 μ atm) is predicted to be reached during the mid
183 21st century, and behavioral impairments were reported in some species following
184 exposure to this level of CO₂³⁹. In our study, juvenile sea bass exposed to elevated
185 CO₂ also spent more time freezing compared to those exposed to control CO₂ levels;
186 this is consistent with previous findings showing that rockfish exposed to elevated
187 CO₂ had elevated levels of anxiety compared to control fish¹¹.

188 **Elevated CO₂ affects olfactory sensitivity**

189 The decreased behavioral response to the predator odor was accompanied
190 by a decrease in olfactory sensitivity (either via response amplitude or detection
191 threshold) to 8 out 10 odorants tested, indicating that olfaction is generally impaired
192 in sea bass exposed to elevated CO₂. The sensitivity of some odorants was more
193 affected than that of others, suggesting that the quality of the perceived odor might
194 be altered in fish exposed to elevated CO₂. This may help to explain the
195 inappropriate (rather than simply inhibited) behavioural responses to complex
196 predator and home-reef odours previously described in the literature^{3,4}. Furthermore,
197 our results are consistent with studies in freshwater pink salmon (*Oncorhynchus*
198 *gorbuscha*) and marine shore crabs (*Carcinus maenas*) exposed to elevated CO₂
199 showing that olfactory sensitivity is impaired, even after prolonged exposure to
200 elevated CO₂, and can be restored within two hours or less of return to control
201 conditions^{40,41}, indicating that this effect persists during chronic exposure to elevated
202 CO₂, but is readily reversible. In freshwater fish and marine crabs exposed to
203 elevated CO₂/low pH the loss of response to some odorants, including alarm cue,
204 has been attributed to structural and functional changes of the chemical cues
205 themselves^{19,41}. This is also a potential explanation for how elevated CO₂ affects
206 odorant-receptor binding in the current study.

207 We estimate that under elevated CO₂ sea bass must be up to 42% closer to
208 the odorant source than in current day conditions to allow detection (Fig. 3). This
209 would increase the risk of predation or decrease the ability to find food, resulting in a
210 direct impact on survival. This is consistent with the observation that coral reef fish
211 raised in elevated CO₂ show a 9-fold increase in mortality in the wild⁴. The activity of
212 sea bass exposed to elevated CO₂ was significantly reduced (Fig. 1) likely resulting
213 in reduced energetic costs. Importantly, this reduced activity would also further

214 reduce their chances of encountering odors in elevated CO₂ conditions. Therefore, if
215 these changes in detection thresholds persist during longer term exposure to
216 elevated CO₂ (as found in freshwater salmon⁴⁰), these could have important
217 ecological consequences at the population level, affecting communication with
218 conspecifics, prey detection and, particularly, predator avoidance.

219 **Elevated CO₂ affects gene expression**

220 Only one recent study has investigated the effect of elevated CO₂ on global
221 gene expression patterns in the brain of fish, and found that in juvenile spiny
222 damselfish (*Acanthochromis polyacanthus*) genes associated with brain glucose,
223 serine and glycine metabolism were differentially expressed in fish exposed to
224 elevated CO₂⁴². Interestingly, in our study both electrophysiology and RNA-Seq
225 results indicated that the response to glutamate was most affected by elevated CO₂.
226 We show that in fish exposed to elevated CO₂ genes encoding CAMKII, MAPK
227 kinase and AMPA glutamate receptors were downregulated, and genes associate
228 with AMPA receptor cycling (*tmub1* and *nptx2*) were upregulated. These genes are
229 involved in long term depression (a long-lasting decrease in synaptic strength), a
230 process associated with a decline in learning and memory in higher brain centres³⁰.
231 The expression of genes involved in maintaining neuronal excitability (*chrna7*,
232 *gria1b*, *scn4ab*, *cacna2d*) also decreased in fish exposed to elevated CO₂. A
233 decrease in synaptic plasticity in the olfactory bulb and a decrease in neuronal
234 excitability in the olfactory system suggest that less olfactory information was being
235 sent to higher brain centers. Additionally, OR genes in the olfactory bulb were
236 downregulated in fish exposed to elevated CO₂. These genes have been shown to
237 be involved in the patterning of the olfactory bulb in mammals, particularly during the
238 development of the olfactory system³⁴. Interestingly, these findings are consistent

239 with impaired learning in responding to a predator odor in larval damselfish
240 (*Pomacentrus amboinensis*) exposed to elevated CO₂¹³, perhaps due to reduced
241 olfactory information reaching higher brain centers, compromising learning and
242 memory formation.

243 **Novel physiological mechanism**

244 We propose a novel mechanism based in the olfactory system to explain how
245 elevated CO₂ alters the behavior of fish (Fig. 5). First, we show that elevated CO₂
246 can have a direct effect on the sensitivity of olfactory reception to various odorants in
247 sea bass, likely by reduced affinity of odorant-receptor binding in the olfactory
248 epithelium. Our electrophysiology data show that fewer impulses are sent to the
249 olfactory bulb in response to most odorants, regardless of concentration. This would
250 result in a decrease in the activity of olfactory bulb synapses, detected by a change
251 in the timing and the frequency of calcium cycling in these neurons, a process that
252 can lead to a decrease in synaptic plasticity³⁰. Indeed, gene expression results show
253 that sea bass exposed to elevated CO₂ downregulate genes involved in synaptic
254 plasticity and maintaining the excitability of both peripheral olfactory receptor
255 neurons and central olfactory bulb neurons, supporting our hypothesis. Therefore,
256 we propose that under future levels of elevated CO₂ fish may sense less information
257 through their olfactory receptors, and this would be compounded by less peripheral
258 olfactory information being transmitted to higher brain centers. Additionally, we also
259 found decreases in the expression of genes involved in in the wiring of the olfactory
260 system, an important developmental process for juvenile fish. These physiological
261 and molecular changes are consistent with the altered behavior observed in this
262 study and others and have strong implications for fitness in the wild.

263 The mechanism of altered neurotransmitter function previously hypothesized to
264 explain the impairments of sensory-induced behaviors observed in coral reef fish is
265 limited to alterations at the level of the brain⁴³. It proposes that extracellular acid-
266 base regulatory changes that fish undergo in response to exposure to elevated CO₂
267 lead to changes in gradients for HCO₃⁻ and Cl⁻ ions across neuronal cell membranes
268 in the brain. In turn these changes are suggested to interfere with the normal
269 functioning of the gamma-aminobutyric acid A (GABA_A) receptor, causing increased
270 excitation rather than inhibition of the nervous system and the observed downstream
271 behavioral impairments^{18, 43}. However, not all fish are good acid-base regulators, and
272 some do not regulate extracellular pH at all when facing elevated CO₂
273 environments⁴⁴. The mechanism proposed here is independent of any changes in
274 blood acid-base chemistry but is instead dependent on the external (seawater)
275 changes in CO₂/H⁺. This raises the possibility that all fish species exposed to
276 elevated CO₂ are potentially susceptible to the direct impairment of peripheral
277 olfactory sensitivity proposed here, whereas the central brain impairment of sensory
278 behavior will principally be relevant to species that are good acid-base regulators.

279 An apparent discrepancy is that in some studies behavioral abnormalities
280 previously shown for fish exposed to ocean acidification are not evident for the first
281 24 hours of exposure⁴. However, these previous observations are based on fish
282 exposed to strong odors, probably well above the threshold of detection^{3,4,43}. Thus,
283 even a 50% reduction in the olfactory sensitivity at these high odorant concentrations
284 would not prevent fish from smelling these strong predator odors, giving rise to some
285 form of behavioral response under elevated CO₂ conditions. By contrast, the
286 peripheral mechanism proposed here would impair olfaction following any duration of
287 elevated CO₂ conditions, particularly when odorants are close to their detection

288 threshold, a more realistic scenario in a natural environment. Secondly, the
289 behavioral responses documented previously are downstream of the central brain
290 GABA-regulated processes that should only be impaired secondary to acid-base
291 regulation and changes in blood chemistry⁴⁵. It is also important to recognize that the
292 two models (the one proposed here based in the olfactory system and the previously
293 proposed impairment of GABA receptor function) are not mutually exclusive. Indeed,
294 it seems likely that they would operate together during exposures lasting longer than
295 24 h, in particular for acid-base regulators, impacting sensory behavior through two
296 distinct physiological mechanisms and ultimately impairing fitness.

297 **Conclusions**

298 Recent studies, including ours, indicate that behavioral responses persist, or
299 become more pronounced, with prolonged experimental exposure to elevated
300 CO₂^{4,12} and in fish that live in naturally high CO₂ environments (near CO₂ seeps)⁴⁶.
301 Additionally, it is not known if the relatively fast change in CO₂ predicted for this
302 century would allow sea bass and other fishes to acclimate or adapt to a high CO₂
303 world, but one generation is apparently not enough to mitigate the effects of elevated
304 CO₂⁴⁷. We propose that the impairment of sensory behavior is induced via not one,
305 but two complementary physiological mechanisms, acting on the olfactory system
306 and on the GABA receptor function in the brain. In essence, fish are impacted at two
307 distinct levels of the sensory-behavioral system, both at the periphery and the central
308 nervous system affecting their behavior. This suggests that complete adaptation may
309 require phenotypic modification at both of these targets of CO₂ exposure. In turn this
310 could either increase selection pressure on this sensory pathway or increase the
311 time required for selection compared to if there was only one target mechanism, and
312 thus complicate predictions about the length of time required for adaptation to occur.

313 Ultimately, it is becoming clear that an elevated CO₂ environment has the potential
314 for major negative impacts on olfactory-mediated behavior of fish across a wide
315 range of habitats and latitudes. This highlights the potential for ecologically
316 significant population-level impacts on fishes, and perhaps other marine fauna,
317 including on economically and ecologically important species.

318 **Funding**

319 This study was supported by grants from Association of European Marine Biology
320 Laboratories [227799], the Natural Environment Research Council [RWW;
321 NE/H017402/1], the Biotechnology and Biological Sciences Research Council
322 [RWW; BB/D005108/1], Fundação para a Ciência e Tecnologia (Portuguese Science
323 Ministry) [UID/Multi/04326/2013] and a Royal Society Newton International
324 Fellowship to CSP. CSP is also a beneficiary of a Starting Grant from AXA.

325 **Author contributions**

326 CSP and RWW designed the behavior experiments. CSP performed the experiments
327 and analyzed those data; CSP, PCH and RWW designed the electrophysiology
328 study, CSP and PCH performed the electrophysiology experiments. CSP, TMUW,
329 RvA, and EMS designed the transcriptomics experiments, CSP performed the
330 experiments and constructed the libraries. CSP performed the bioinformatics
331 analysis and interpreted the results with help from TMUW, RvA, and EMS; all
332 authors contributed to and provided feedback on various drafts of the paper.

333 **Acknowledgements**

334 The authors are grateful to Drs Lee Hagey and Alan Hofmann (UCSD) for their kind
335 gift of cyprinol sulphate and scymnol sulphate. We would also like to thank the
336 Aquatic Research Centre (ARC) staff at the University of Exeter for their assistance
337 with fish husbandry and experimental setup, Bas Verbruggen for helpful
338 bioinformatics advice, and Louisa Salisbury for help with tissue sampling.

339 **References**

- 340 1. Velez Z, Hubbard P, Welham K, Hardege J, Barata E, Canário AM.
341 Identification, release and olfactory detection of bile salts in the intestinal fluid
342 of the Senegalese sole (*Solea senegalensis*). *J Comp Physiol A* **195**, 691-698
343 (2009).
- 344 2. Yacoob SY, Browman HI. Olfactory and gustatory sensitivity to some feed-
345 related chemicals in the Atlantic halibut (*Hippoglossus hippoglossus*).
346 *Aquaculture* **263**, 303-309 (2007).
- 347 3. Munday PL, Dixson DL, Donelson JM, Jones GP, Pratchett MS, Devitsina GV,
348 *et al.* Ocean acidification impairs olfactory discrimination and homing ability of
349 a marine fish. *Proc Natl Acad Sci USA* **106**, 1848-1852 (2009).
- 350 4. Munday PL, Dixson DL, McCormick MI, Meekan M, Ferrari MCO, Chivers DP.
351 Replenishment of fish populations is threatened by ocean acidification. *Proc*
352 *Natl Acad Sci USA* **107**, 12930-12934 (2010)
- 353 5. Yambe H, Kitamura S, Kamio M, Yamada M, Matsunaga S, Fusetani N, *et al.*
354 L-Kynurenine, an amino acid identified as a sex pheromone in the urine of
355 ovulated female masu salmon. *Proc Natl Acad of Sci* **103**, 15370-15374
356 (2006).
- 357 6. Arvedlund M, McCormick MI, Fautin DG, Bildsøe M. Host recognition and
358 possible imprinting in the anemonefish *Amphiprion melanopus* (Pisces:
359 Pomacentridae). *Mar Ecol Prog Ser* **188**, 207-218 (1999).
- 360 7. Arvedlund M, Takemura A. The importance of chemical environmental cues
361 for juvenile *Lethrinus nebulosus* Forsskål (Lethrinidae, Teleostei) when
362 settling into their first benthic habitat. *J Exp Mar Biol Ecol* **338**, 112-122
363 (2006).

- 364 8. Atema J, Kingsford MJ, Gerlach G. Larval reef fish could use odour for
365 detection, retention and orientation to reefs. *Mar Ecol Prog Ser* **241**, 151-160
366 (2002).
- 367 9. Gerlach G, Atema J, Kingsford MJ, Black KP, Miller-Sims V. Smelling home
368 can prevent dispersal of reef fish larvae. *Proc Natl Acad of Sci* **104**, 858-863
369 (2007).
- 370 10. Vrieze LA, Sorensen PW. Laboratory assessment of the role of a larval
371 pheromone and natural stream odor in spawning stream localization by
372 migratory sea lamprey (*Petromyzon marinus*). *Can J Fish Aquat Sci* **58**, 2374-
373 2385 (2001).
- 374 11. Hamilton TJ, Holcombe A, Tresguerres M. CO₂-induced ocean acidification
375 increases anxiety in Rockfish via alteration of GABA_A receptor functioning.
376 *Proc R Soc Lond B Biol Sci* **281**, 20132509 (2014).
- 377 12. Jutfelt F, Bresolin de Souza K, Vuylsteke A, Sturve J. Behavioural
378 disturbances in a temperate fish exposed to sustained high-CO₂ levels. *PLoS*
379 *ONE* **8**, e65825 (2013).
- 380 13. Ferrari MCO, Manassa RP, Dixon DL, Munday PL, McCormick MI, Meekan
381 MG, *et al.* Effects of ocean acidification on learning in coral reef fishes. *PLoS*
382 *ONE* **7**, e31478 (2012).
- 383 14. Dixon DL, Jennings AR, Atema J, Munday PL. Odor tracking in sharks is
384 reduced under future ocean acidification conditions. *Global Change Biol* **21**,
385 1454–1462 (2015).
- 386 15. Green L, Jutfelt F. Elevated carbon dioxide alters the plasma composition and
387 behaviour of a shark. *Biol Lett* **10**, 20140538 (2014).

- 388 16. Chivers DP, McCormick MI, Nilsson GE, Munday PL, Watson S-A, Meekan
389 MG, *et al.* Impaired learning of predators and lower prey survival under
390 elevated CO₂: a consequence of neurotransmitter interference. *Global*
391 *Change Biol* **20**, 515-522 (2013).
- 392 17. Watson S-A, Lefevre S, McCormick MI, Domenici P, Nilsson GE, Munday PL.
393 Marine mollusc predator-escape behaviour altered by near-future carbon
394 dioxide levels. *Proc R Soc B Biol Sci* **281**, 20132377 (2014).
- 395 18. Heuer RM, Welch MJ, Rummer JL, Munday PL, Grosell M. Altered brain ion
396 gradients following compensation for elevated CO₂ are linked to behavioural
397 alterations in a coral reef fish. *Scientific Reports* **6**, 33216 (2016).
- 398 19. Leduc AOHC, Munday PL, Brown GE, Ferrari MCO. Effects of acidification on
399 olfactory-mediated behaviour in freshwater and marine ecosystems: a
400 synthesis. *Philos Trans R Soc Lond B Biol Sci* **368**, 1-14 (2013).
- 401 20. Fariña AC, Azevedo M, Landa J, Duarte R, Sampedro P, Costas G, *et al.*
402 *Lophius* in the world: a synthesis on the common features and life strategies.
403 *ICES Journal of Marine Science* **65**, 1272-1280 (2008).
- 404 21. Hara T. The diversity of chemical stimulation in fish olfaction and gustation.
405 *Rev Fish Biol Fisheries* **4**, 1-35 (1994).
- 406 22. Buchinger TJ, Li W, Johnson NS. Bile salts as semiochemicals in fish. *Chem*
407 *Senses*: bju039v031 (2014).
- 408 23. Leduc AOHC, Roh E, Macnaughton CJ, Benz F, Rosenfeld J, Brown GE.
409 Ambient pH and the response to chemical alarm cues in juvenile Atlantic
410 salmon: mechanisms of reduced behavioral responses. *Trans Am Fish Soc*
411 **139**, 117-128 (2010).

- 412 24. Lönnstedt OM, McCormick MI. Chemical alarm cues inform prey of predation
413 threat: the importance of ontogeny and concentration in a coral reef fish. *Anim*
414 *Behav* **82**, 213-218 (2011).
- 415 25. Frade P, Hubbard PC, Barata EN, Canario AVM. Olfactory sensitivity of the
416 Mozambique tilapia to conspecific odours. *J Fish Biol* **61**, 1239-1254 (2002).
- 417 26. Hubbard PC, Barata EN, Canário AVM. Olfactory Sensitivity of the gilthead
418 seabream (*Sparus auratus* L) to conspecific body fluids. *J Chem Ecol* **29**,
419 2481-2498 (2003).
- 420 27. Haas BJ, Papanicolaou A, Yassour M, Grabherr M, Blood PD, Bowden J, *et*
421 *al.* De novo transcript sequence reconstruction from RNA-seq using the Trinity
422 platform for reference generation and analysis. *Nat Protocols* **8**, 1494-1512
423 (2013).
- 424 28. Li B, Dewey CN. RSEM: accurate transcript quantification from RNA-Seq data
425 with or without a reference genome. *BMC Bioinformatics* **12**, 323 (2011).
- 426 29. Robinson MD, McCarthy DJ, Smyth GK. edgeR: a Bioconductor package for
427 differential expression analysis of digital gene expression data. *Bioinformatics*
428 **26**, 139-140 (2010).
- 429 30. Malenka RC, Nicoll, A. R. Long-Term Potentiation--A Decade of Progress?
430 *Science* **285**, 1870-1874 (1999).
- 431 31. Zhang JJ, Okutani F, Inoue S, Kaba H. Activation of the mitogen-activated
432 protein kinase/extracellular signal-regulated kinase signaling pathway leading
433 to cyclic AMP response element-binding protein phosphorylation is required
434 for the long-term facilitation process of aversive olfactory learning in young
435 rats. *Neuroscience* **121**, 9-16 (2003).

- 436 32. Deigweiher K, Koschnick N, Pörtner H-O, Lucassen M. Acclimation of ion
437 regulatory capacities in gills of marine fish under environmental hypercapnia.
438 *American Journal of Physiology - Regulatory, Integrative and Comparative*
439 *Physiology* **295**, R1660-R1670 (2008).
- 440 33. Hashiguchi Y, Furuta Y, Nishida M. Evolutionary patterns and selective
441 pressures of odorant/pheromone receptor gene families in teleost fishes.
442 *PLoS ONE* **3**, e4083 (2008).
- 443 34. Dubacq C, Fouquet C, Trembleau A. Making scent of the presence and local
444 translation of odorant receptor mRNAs in olfactory axons. *Developmental*
445 *Neurobiology* **74**, 259-268 (2014).
- 446 35. Benhaïm D, Péan S, Lucas G, Blanc N, Chatain B, Bégout M-L. Early life
447 behavioural differences in wild caught and domesticated sea bass
448 (*Dicentrarchus labrax*). *Applied Animal Behaviour Science* **141**, 79-90 (2012).
- 449 36. Ferrari MCO, Dixon DL, Munday PL, McCormick MI, Meekan MG, Sih A, *et*
450 *al.* Intrageneric variation in antipredator responses of coral reef fishes affected
451 by ocean acidification: implications for climate change projections on marine
452 communities. *Global Change Biol* **17**, 2980-2986 (2012).
- 453 37. Cripps IL, Munday PL, McCormick MI. Ocean acidification affects prey
454 detection by a predatory reef fish. *PLoS ONE* **6**, e22736 (2011).
- 455 38. Duteil M, Pope EC, Pérez-Escudero A, de Polavieja GG, Fürtbauer I, Brown
456 MR, *et al.* European sea bass show behavioural resilience to near-future
457 ocean acidification. *Royal Society Open Science* **3**,
458 <https://doi.org/10.1098/rsos.160656> (2016).

- 459 39. Munday PL, Pratchett MS, Dixon DL, Donelson JM, Endo GGK, Reynolds
460 AD, *et al.* Elevated CO₂ affects the behavior of an ecologically and
461 economically important coral reef fish. *Marine Biology* **160**, 2137-2144 (2013).
- 462 40. Ou M, Hamilton TJ, Eom J, Lyall EM, Gallup J, Jiang A, *et al.* Responses of
463 pink salmon to CO₂-induced aquatic acidification. *Nat Clim Change* **5**, 950-
464 955 (2015).
- 465 41. Roggatz CC, Lorch M, Hardege JD, Benoit DM. Ocean acidification affects
466 marine chemical communication by changing structure and function of peptide
467 signalling molecules. *Glob Chang Biol* **22**, 3914-3926 (2016).
- 468 42. Schunter C, Welch MJ, Ryu T, Zhang H, Berumen ML, Nilsson GE, *et al.*
469 Molecular signatures of transgenerational response to ocean acidification in a
470 species of reef fish. *Nature Clim Change* **6**, 1014-1018 (2016).
- 471 43. Nilsson GE, Dixon DL, Domenici P, McCormick MI, Sorensen C, Watson S-
472 A, *et al.* Near-future carbon dioxide levels alter fish behaviour by interfering
473 with neurotransmitter function. *Nat Clim Change* **2**, 201-204 (2012).
- 474 44. Brauner CJ, Wang T, Wang Y, Richards JG, Gonzalez RJ, Bernier NJ, *et al.*
475 Limited extracellular but complete intracellular acid-base regulation during
476 short-term environmental hypercapnia in the armoured catfish, *Liposarcus*
477 *pardalis*. *J Exp Biol* 2004, **207**(19): 3381-3390.
- 478 45. Esbaugh A, Heuer R, Grosell M. Impacts of ocean acidification on respiratory
479 gas exchange and acid-base balance in a marine teleost, *Opsanus beta*. *J*
480 *Comp Physiol B* **182**, 921-934 (2012).
- 481 46. Munday PL, Cheal AJ, Dixon DL, Rummer JL, Fabricius KE. Behavioural
482 impairment in reef fishes caused by ocean acidification at CO₂ seeps. *Nat*
483 *Clim Change* **4**: 487-492 (2014).

484 47. Welch MJ, Watson S-A, Welsh JQ, McCormick MI, Munday PL. Effects of
485 elevated CO₂ on fish behaviour undiminished by transgenerational
486 acclimation. *Nat Clim Chang* **4**: 1086-1089 (2014).

487

488

489

490 **Methods**

491 *1.1 Fish maintenance*

492 Juvenile European sea bass (*D. labrax*; 4–8 g, 7 cm total length) were obtained from
493 Ecloserie Marines de Gravelines, France and initially held within a stock tank
494 containing recirculating artificial seawater (15°C) in the University of Exeter Aquatic
495 Resource Centre. Photoperiod was controlled (12:12 light:dark) and fish were fed
496 twice daily *ad libitum* with Perla MP pellets (Skretting, Shay Lane, Longridge,
497 Preston, UK).

498 *1.2 Exposure to simulated ocean acidification*

499 All experiments were conducted under approved protocols according to the UK
500 Home Office regulations for use of animals in scientific procedures, and approved by
501 the University of Exeter Ethics committee. Sea bass were exposed to simulated
502 ocean acidification for 2, 7 and 14 days as described previously⁴⁸. For water
503 chemistry parameters during exposure for behavioral and the transcriptomics
504 experiments see Supplementary Table S1.

505 *1.3 Behavioral testing*

506 Fish were fed 1% body mass of Perla MP pellets (Skretting, Shay Lane, Longridge,
507 Preston, UK) before behavioral experiments to prevent hunger from over-riding
508 behavioral responses to a predator⁴⁹. Sea bass were then transferred by scoop
509 (without air exposure) to a flume (60 cm x 16 cm x 8 cm) using flow-through control
510 or elevated CO₂ (same as exposure; see Supplementary Table S2). Sea bass were
511 left for 30 min in the flume as preliminary experiments indicated this was sufficient
512 time for them to resume normal activity and behavior after transfer. Fish behavior
513 was recorded from above using a video camera (Sony Handycam, DCR-SR190).
514 The bottom of the flume had a black grid in order to quantify activity. Activity was

515 recorded for 5 min to determine baseline activity, after which diluted (1:1,000,000)
516 monkfish (*Lophius piscatorius*) bile was added to the inflowing water for a further 5
517 min. Two fish were tested at the same time (one from each treatment) and they were
518 randomly assigned.
519 Video analysis was used to quantify the number of squares visited by each fish in the
520 minute before monkfish bile and during the 5 min in which monkfish bile was present.
521 Video analysis was carried out blind regarding CO₂ treatment. The total number of
522 seconds spent freezing was quantified (expressed as % of time) in the 5 min prior to
523 and during presence of monkfish bile. A fish was considered to be freezing when it
524 spend more than 5 sec without moving.

525 *Statistical analysis*

526 To test for the effects of treatment and time (day) on baseline activity and
527 change in activity a two-way analysis of variance was performed (SigmaPlot, version
528 9.2). A Holm-Sidak *post hoc* test was used to test for differences from control and
529 between different times. To test for the effects of treatment and presence of predator
530 odor on freezing behavior a two-way repeated measures analysis of variance was
531 performed, followed by Holm-Sidak *post hoc* test when significant differences were
532 detected (at 7 and 14 days). A significance level of $P < 0.05$ was used throughout the
533 analysis.

534 *1.4 Transcriptomics*

535 *RNA extraction, library preparation and sequencing*

536 Fish were humanely sacrificed on day 2 and 7 of the exposure period (see above) by
537 a lethal dose of benzocaine (0.5 g L⁻¹; Sigma-Aldrich). Brains and olfactory
538 epitheliums were dissected and snap frozen in liquid nitrogen and stored at -80°C

539 prior to transcript profiling. Transcript profiling was conducted in the brain and
540 olfactory epithelium of 6 and 4 fish per treatment group, respectively.
541 RNA was extracted using an RNeasy mini kit (QIAGEN, Catalog # 74104), and on-
542 column DNase treatment was performed, according to the manufacturer's
543 instructions. The concentration, purity and integrity of RNA were determined using a
544 NanoDrop ND-1000 Spectrophotometer (NanoDrop Technologies, USA) and an
545 Agilent 2100 Bioanalyzer (Agilent Technologies, Inc., USA). All RNA input to library
546 construction was of high quality with A_{260}/A_{280} and A_{260}/A_{230} ratios > 2.0 and RIN
547 scores > 8.5 . ERCC spike-in control mixes (Ambion) were added to all individual
548 RNA samples, according to the manufacturer's instructions to allow for analysis of
549 the accuracy of the transcript quantification and dynamic range (Supplementary
550 Table S3-S4). cDNA libraries for all samples were then prepared using the NextFlex
551 Rapid RNA-Seq Kit (Cat # 5138-02), multiplexed with 12 and 16 samples per lane for
552 the olfactory epithelium (OE) and bulb (OB), respectively (4 lanes total). The
553 samples were sequenced using an Illumina HiSeq 2500 in standard mode, to
554 generate 100 bp paired reads. Sequencing of the OE and OB produced 1007 and
555 349 million paired 100 bp reads, respectively, averaging 41.9 and 21.8 million reads
556 per library, respectively.

557 Data analysis

558 All analyses were carried out on a local server running the NEBC Bio-Linux 7
559 environment⁵⁰. Raw sequence reads were quality-trimmed using Trimmomatic⁵¹
560 (v0.32). The first 12 bp were removed from each read (to remove 5' bias caused by
561 non-random hexamer priming) and a 4-base wide sliding window was used to cut
562 when the average Phred score quality per base was below 20. Furthermore, bases
563 were trimmed of the end of the reads when the Phred score quality was less than 15

564 and reads shorter than 36 bp were discarded. Digital normalization was performed to
565 remove highly duplicated reads using the normalize-by-median.py script part of the
566 khmer package described by Brown et al.⁵², with the recommended k-mer value of
567 20 and a coverage threshold of 30. Following digital normalization, a total of 69.4
568 and 47.2 million paired reads originating from all the OE and OB sequence libraries,
569 respectively, were retained and input into the *de novo* transcriptome assemblies. The
570 assemblies were constructed using Trinity⁵³ (version 2.1.1) and the default
571 parameters, specifying a minimum contig length of 200 bp (see Supplementary
572 Table S5 for comparative summary statistics for each assembly). All transcripts in
573 the final assemblies were annotated using Blastx⁵⁴ against the Ensembl peptide
574 databases⁵⁵ (Release 79; March 2015) using an E-value cut-off of $1e^{-15}$ and 'best
575 hits' were assigned to transcripts in the following preferential order to facilitate
576 assigning Gene Ontology (GO) terms for subsequent functional annotation analysis:
577 zebrafish (*Danio rerio*); human (*Homo sapiens*) and mouse (*Mus musculus*);
578 stickleback (*Gasterosteus aculeatus*), medaka (*Oryzias latipes*), Japanese pufferfish
579 (*Takifugu rubripes*), tilapia (*Oreochromis niloticus*), green spotted puffer fish
580 (*Tetraodon nigroviridis*), cod (*Gadus morhua*) and spotted gar (*Lepisosteus*
581 *oculatus*). Additional annotation was performed using Diamond blastx⁵⁶ (v0.8.22.84;
582 E-value < $1e^{-10}$) against the full NCBI nr protein database (March 2016). Transcripts
583 that were not annotated or that showed similarity to non-metazoan genes only
584 (n=826 and n=964 for the olfactory epithelium and olfactory bulb, respectively) were
585 removed from the assembly prior to differential expression analysis, using MEGAN
586 ⁵⁷ (version 5.11.3, January 2016). All four families of olfactory receptors were well
587 represented in the olfactory epithelium transcriptome (Supplementary Table S6).

588 Reads from individual samples were mapped against the Trinity transcriptome
589 assemblies for each tissue and transcript abundances were calculated using RSEM
590 and the `align_and_estimate_abundance.pl` script supplied by Trinity (version 2.1.1).
591 Differential expression analysis was carried out at the gene level using EdgeR and
592 the `run_DE_analysis.pl` script provided by Trinity. Genes were considered
593 differentially expressed with a FDR < 0.1 (Benjamini-Hochberg correction; see
594 Supplementary File 2 and Figures S2-S4). Hierarchical clustering was performed on
595 all differentially expressed genes using the `analyze_diff_expr.pl` script. Functional
596 analysis was then performed for differentially expressed genes from each tissue
597 using the Database for Annotation, Visualization and Integrated Discovery⁵⁸ (DAVID
598 v6.8, Oct 2016), using the zebrafish gene identifiers and the *de novo* transcriptomes
599 for each tissue as a background (see Supplementary Tables S7-S10). Kyoto
600 Encyclopedia of Genes and Genomes (Kegg) pathways (<http://genome.jp/kegg/>) and
601 Gene Ontology (GO) terms (<http://geneontology.org/>) for Biological Process, Cellular
602 Component and Molecular Function were considered significantly over-represented
603 when adjusted $P < 0.1$ (Benjamini-Hochberg).

604 qRT-PCR validation

605 RNA-Seq expression was validated using Quantitative Realtime PCR (qRT-PCR) for
606 a selection of key genes that were shown to be differentially expressed in the
607 olfactory epithelium using RNAseq (see Table S11). We used a combination of the
608 same samples used for the RNA-Seq and additional replicates from the same
609 experiments (n=5-8 replicates per treatment). Primers were designed for the
610 assembled transcripts using OligoArchitect Online (Sigma-Aldrich) and blasted using
611 the NCBI BLAST tool to verify specificity. RNA concentration and purity were
612 assessed with a NanoDrop ND-1000 Spectrophotometer (NanoDrop Technologies,

613 Wilmington, USA). cDNA was synthesized according to manufacturer's instructions
614 from 2 µg of total RNA treated with RQ1 DNase (Promega, Southampton, UK) using
615 random hexamers (MWG-Biotech) and M-MLV reverse transcriptase (Promega).
616 cDNA was diluted 1:2 and RT-qPCR was performed in triplicate using a CFX96
617 Touch Real-Time PCR Detection System (Bio-Rad Laboratories, Hercules, CA)
618 using SYBR Green chemistry (Bio-Rad). Each reaction was performed in a total
619 volume of 15 µl. PCR conditions consisted of 10 minutes at 95 °C, followed by 40
620 cycles of 10s at 95°C and 30s at the optimized annealing temperature for each
621 primer pair. High resolution melt analysis (60-95°C, in 0.5 °C increments) was then
622 performed to verify the presence of single amplification products of the expected size
623 and T_m. A template-minus negative control was run in triplicate on each plate to
624 verify the absence of cDNA contamination, and replicates displaying non-specific
625 amplification were removed from the analysis. Annealing temperature for each
626 primer pair was optimized by performing a standard curve. The linear correlation (R²)
627 between the mean C_q and the logarithm of the cDNA dilution was > 0.98 in each
628 case, and efficiencies were between 1.92-2.11. The primer sequences, PCR product
629 sizes, annealing temperatures and PCR efficiencies for each primer pair are shown
630 in Table S11. Efficiency-corrected relative expression levels of each target gene
631 were determined according to Pfaffl⁵⁹ and by normalizing to the control gene, *rpl8*.
632 This control gene was chosen because it remained constant across all treatments
633 and tissues in the RNA-Seq experiments and has been widely used as a control
634 gene for qPCR experiments in fish⁶⁰. Unpaired two-sample t-tests were used to test
635 for significant differences between treatments.

636 *1.5 Electrophysiology*

637 *Experimental animals*

638 Fish care and experimentation complied with the guidelines of the European
639 Union Council (86/609/EU) and Portuguese legislation for the use of laboratory
640 animals under a “Group-1” license issued by the Veterinary General Directorate of
641 the Ministry of Agriculture, Rural Development and Fisheries of Portugal. Juvenile
642 European sea bass (*Dicentrarchus labrax*; 16.6 - 36.0 g) were obtained as 2 month-
643 old larvae from the Ifremer experimental aquaculture station in Palavas-Iles-Flots,
644 France, and raised in flow-through natural sea water from the Ria Formosa (sand-
645 filtered and UV sterilized, ~36 psu) at the Ramalhete Aquatic Facilities, Faro,
646 Portugal. Sea bass were transferred to Laboratório Experimental de Organismos
647 Aquáticos and held in recirculated natural sea water (as above) at 20°C on a 12 h
648 light : 12 h dark photoperiod for at least two weeks before the start of experiments.

649 Choice of olfactory odorants

650 We tested the effect of acutely elevated CO₂ on the olfactory sensitivity of sea
651 bass to amino acids, as odorants principally mediating food detection²¹. Bile acids
652 are cholesterol derivatives secreted in the bile fluid of vertebrates that aid in the
653 digestion and absorption of lipids and lipid-soluble vitamins^{61, 62}. In fish they are also
654 potent olfactory stimuli, which may be involved in chemically-mediated interactions
655 both within and between species⁶³. Cyprinol sulphate is a widely produced bile acid
656 in cyprinids but also found in other teleosts⁶¹ and scymnol sulphate is a major
657 component of elasmobranch bile⁶¹. The body fluids of fish are complex mixtures that
658 can contain potent chemical signals^{1,64,65}, including alarm cues released from
659 damaged skin of some teleost fish that elicit behavioral responses vital for nearby
660 conspecifics to escape from, or be aware of, predators^{23,24}. Therefore, we measured
661 the effect of elevated CO₂/H⁺ on the olfactory responses of sea bass to conspecific
662 bile, intestinal fluid and damage-released alarm cue.

663 Collection of body fluids and alarm cue

664 Body fluids (bile and intestinal fluid) were collected from three male and three
665 female sea bass (146-205 g) that were being sampled for other experiments. Once
666 collected, the samples were kept on ice until return to the lab, then centrifuged, all
667 were pooled and frozen (-20°C) into aliquots until use. For alarm cue four sea bass
668 were killed by a sharp blow to the head. Care was used not to touch the flanks of the
669 fish. Using a sharp scalpel, twelve 1 cm superficial cuts were made to one flank of
670 each fish. The flank was rinsed with charcoal-filtered sea water. This was then either
671 immediately frozen in aliquots for later use (no longer than 48 h later) or diluted for
672 immediate use.

673 Olfactory nerve recording

674 Sea bass were anesthetized in aerated natural seawater containing 300 mg l⁻¹
675 MS222 (ethyl-3-aminobenzoate methanesulphonate salt; Sigma-Aldrich, Spain)
676 buffered with 0.6 g.l⁻¹ sodium bicarbonate, until response to tail pinch had stopped.
677 An intramuscular injection of gallamine triethiodide (Sigma-Aldrich, Spain; 3 mg kg⁻¹
678 in 0.9% NaCl) was then given^{66,67,68}. The sea bass were then placed in a padded V-
679 support and the gills flushed with aerated natural sea water containing 150 mg l⁻¹
680 MS222 (also buffered as above). The electrodes were placed in the olfactory nerve
681 (near the bulb) in a place that gave maximal response to 10⁻⁴ M L-cysteine.
682 Olfactory nerve activity was recorded using tungsten micro-electrodes (0.1 MΩ,
683 World Precision Instruments, UK) as previously described²⁶ (Supplementary Fig S5).
684 The fish was connected to earth via a copper wire inserted in the flank. The raw
685 signal was amplified (20,000x; AC pre-amplifier, Neurolog NL104; Digitimer Ltd.,
686 Welwyn Garden City, UK), filtered (high pass: 200 Hz, low pass: 3,000 Hz; Neurolog
687 NL125, Digitimer Ltd.) and integrated (time constant 1 s; Neurolog NL703, Digitimer

688 Ltd.). Raw and integrated signals were digitized (Digidata 1440A, Molecular Devices,
689 Sunnyvale, CA, USA) and recorded on a PC running AxoScope™ software (version
690 10.2, Molecular Devices).

691 At all times, the gills of the fish were flushed with sea water equilibrated with
692 atmospheric air (i.e. control conditions, pH 8.15) containing anesthetic (see above);
693 therefore, only the olfactory epithelium (also known as the rosette) experienced
694 elevated CO₂ conditions. Charcoal-filtered natural sea water was either bubbled with
695 air (control, pH 8.15 ± 0.01, 476 ± 14 μatm) or CO₂ (pH 7.82 ± 0.01, 1122 ± 19 μatm)
696 until the desired pH_{NBS} was reached (see Water Chemistry section for details;
697 Supplementary Table S12). The charcoal-filtered sea water (control or elevated CO₂)
698 was used to make up the odorant solutions and to superfuse the olfactory epithelium
699 during experiments. Amino acid and bile acid solutions were prepared from frozen
700 aliquots of 10⁻² M, while body fluids were prepared from frozen aliquots and all were
701 diluted in charcoal-filtered sea water (either control or elevated CO₂) immediately
702 prior to being used. The order of testing of the odorants was randomized, but each
703 odorant was always given from lowest to highest concentration (10⁻⁸ M to 10⁻³ M).

704 All responses to a given stimulus (except alarm cue) were paired: responses
705 were first recorded while the olfactory epithelium was superfused with control sea
706 water, followed by responses to the odorant made up in elevated CO₂ while the
707 olfactory epithelium was also superfused with elevated CO₂ sea water. Initial
708 measurements were performed in control conditions, followed by elevated CO₂, and
709 then control again and the response was reversible (see Supplementary Fig S6).
710 Therefore, multiple stimuli were tested on the same fish under both control and
711 elevated CO₂ conditions (see Supplementary Fig S7). For testing the responses to
712 alarm cue, half of the responses were measured first under control and then under

713 elevated CO₂ sea water and the other half were measured in the reverse order to
714 account for the rapid degradation of the alarm cue (see Supplementary Fig S8).
715 Each odorant was added through a gravity fed three-way-valve (at a rate of 6 ml min⁻¹)
716 ¹). A blank stimulus, treated in the same way as the odorants, was recorded at the
717 beginning and end of each group of samples. All integrated response amplitudes
718 were normalized to the amplitude of the integrated response to 10⁻⁴ M L-cysteine
719 (the 'standard') made up in control charcoal-filtered sea water. Responses to blanks
720 and standards were recorded regularly at the beginning and end of each group of
721 samples (every 3 - 5 samples) throughout the recording session. Each stimulus was
722 applied for 4 seconds, with at least 1 minute between odorants to allow complete
723 recovery of the receptors²⁶.

724 Water chemistry measurements

725 Salinity and temperature were measured using a Thermo Scientific 'Orion Star
726 A329' (Thermo Fischer Scientific, Inc., Waltham, MA, USA) probe. Charcoal-filtered
727 sea water was previously equilibrated with atmospheric air, and then for the
728 experiments this seawater supply was either bubbled with air (control) or elevated
729 CO₂ until the desired pH_{NBS} was reached (pH meter Model HI 8314, Hanna
730 Instruments, Leighton Buzzard, UK; pH probe Model pHC2401, Radiometer
731 Analytical, Lyon, France). Sea water samples (12 ml) from both treatments were
732 collected and were preserved using standard methods⁶⁹, which were then stored at
733 4°C for dissolved inorganic carbon (DIC) analysis as previously described.

734 Calculations

735 The 'active space' of an odorant was calculated by dividing the release rate of
736 that odorant by the predicted detection threshold for it²⁵

737
$$Active\ space = \frac{Release\ rate}{detection\ threshold} \text{ or } V = \frac{X}{Y} \quad (1)$$

738 For a 5 fold increase in detection threshold (e.g. glutamate)

739 $V(CO_2) = \frac{X}{5Y}$ or $V(CO_2) = \frac{V}{5}$ (2)

740 where, $V(CO_2)$ is the active space in elevated CO_2 assuming no change in release
741 rate.

742 We assumed a homogeneous distribution of the odorant in the water. Although this
743 assumption might be representative of natural environments (due to tidal
744 movements, wave action etc.), it represents a pragmatic and best estimate of the
745 average impact. Based on this assumption we calculated the maximum distance
746 from source for odorant detection:

747 $V = \frac{4}{3}\pi r^3$ (3),

748 where V is the active space and r is the distance from the odorant source.

749 $r = \sqrt[3]{\frac{3V}{4\pi}}$ in control or $r = \sqrt[3]{\frac{3V}{4\pi}}$ or $r = 0.62 * \sqrt[3]{V}$ and

750 $r = \sqrt[3]{\frac{3V/5}{4\pi}}$ in high CO_2 or $r = \sqrt[3]{\frac{3V}{20\pi}}$ or $r = 0.36 * \sqrt[3]{V}$ (4)

751 Therefore, r will decrease to 58% in elevated CO_2 and fish would have to be 42%
752 closer.

753 Data and statistical analysis

754 All statistical analyses and calculations were carried out on normalized data.
755 Detection thresholds were determined from the intercept with the x-axis of linear
756 regression fit to the linear part of individual dose response curves for each odorant
757 and each treatment^{64, 70} (SigmaPlot, version 9.2). To test for the effects of
758 concentration and treatment a two-way repeated measured analysis of variance was
759 performed for all the odorants (SigmaPlot, version 9.2). A Holm-Sidak *post hoc* test
760 was used to test for differences from control and between different concentrations.

761 Paired Student's *t*-tests were used to test for differences between treatments in
762 detection thresholds using Microsoft Excel. A significance level of $P < 0.05$ was used
763 throughout the analysis.

764

765 Data Availability

766 Sequence datasets are available through the NCBI database
767 (https://www.ncbi.nlm.nih.gov/Traces/sra_sub/sub.cgi?subid2311090) and include
768 the individual gene expression sample files. Water chemistry, behaviour and
769 electrophysiology data are available through Pangaea
770 (<https://doi.pangaea.de/10.1594/PANGAEA.884674>).

771 Gene expression results

772 After 2 days of exposure to elevated CO₂ in both the olfactory bulb and the olfactory
773 epithelium there were 23 and 24 differentially expressed genes, respectively
774 (Supplementary Fig S2). After 7 days of exposure to elevated CO₂ there were 73
775 and 71 differentially expressed genes in the olfactory epithelium and bulb,
776 respectively (Supplementary Fig S2). Expression patterns of all five genes used for
777 validation using qRT-PCR were in agreement with those obtained with RNAseq
778 (Table S12), with three of these being significant ($p < 0.05$) and two being marginally
779 not significant ($p = 0.05 - 0.12$, *tmp4* and *klc8*).

780 **References**

781 48. Poulton DA, Porteus CS, Simpson SD. Combined impacts of elevated CO₂
782 and anthropogenic noise on European sea bass (*Dicentrarchus labrax*). *ICES*
783 *Journal of Marine Science* **74**, 1230–1236 (2016).

- 784 49. Chivers D, Puttlitz M, Blaustein A. Chemical alarm signaling by reticulate
785 sculpins, *Cottus perplexus*. *Environ Biol Fish* **57**, 347-352 (2000).
- 786 50. Field D, Tiwari B, Booth T, Houten S, Swan D, Bertrand N, *et al*. Open
787 software for biologists: from famine to feast. *Nat Biotechnol* **24**, 801-803
788 (2006).
- 789 51. Bolger AM, Lohse M, Usadel B. Trimmomatic: A flexible trimmer for Illumina
790 Sequence Data. *Bioinformatics* **15**, 2114–2120 (2014).
- 791 52. Brown C, Howe A, Zhang Q, Pyrkosz A, Brom T. A reference-free algorithm
792 for computational normalization of shotgun sequencing data arXiv:1203.4802
793 [q-bio.GN] (2012).
- 794 53. Haas BJ, Papanicolaou A, Yassour M, Grabherr M, Blood PD, Bowden J, *et*
795 *al*. De novo transcript sequence reconstruction from RNA-seq using the Trinity
796 platform for reference generation and analysis. *Nat Protocols* **8**, 1494-1512
797 (2013).
- 798 54. Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ. Basic local alignment
799 search tool. *J Mol Bio* **215**, 403-410 (1990).
- 800 55. Yates A, Akanni W, Amode MR, Barrell D, Billis K, Carvalho-Silva D, *et al*.
801 Ensembl 2016. *Nucleic Acids Res* **44**, D710-716 (2016).
- 802 56. Buchfink B, Xie C, Huson DH. Fast and sensitive protein alignment using
803 DIAMOND. *Nat Meth* **12**, 59-60 (2015).

- 804 57. Huson DH, Mitra S, Ruscheweyh H-J, Weber N, Schuster SC. Integrative
805 analysis of environmental sequences using MEGAN4. *Genome Res* **21**, 1552-
806 1560 (2011).
- 807 58. Huang DW, Sherman BT, Lempicki RA. Systematic and integrative analysis of
808 large gene lists using DAVID bioinformatics resources. *Nat Protocols* **4**, 44-57
809 (2008).
- 810 59. Pfaffl MW. A new mathematical model for relative quantification in real-time
811 RT-PCR. *Nucleic Acids Res* **29**, e45-e45 (2001).
- 812 60. Filby AL, Tyler CR. Appropriate 'housekeeping' genes for use in expression
813 profiling the effects of environmental estrogens in fish. *BMC Molecular Biology*
814 **8**, 10 (2007).
- 815 61. Hagey Lee R, Møller Peter R, Hofmann AF, Krasowski Matthew D. Diversity
816 of bile salts in fish and amphibians: Evolution of a complex biochemical
817 pathway. *Physiol Biochem Zool* **83**, 308-321 (2010).
- 818 62. Ballantyne JS. Jaws: The Inside Story. The Metabolism of Elasmobranch
819 Fishes. *Comp Biochem Physiol B Biochem Mol Biol* **118**, 703-742 (1997).
- 820 63. Buchinger TJ, Li W, Johnson NS. Bile salts as semiochemicals in fish. *Chem*
821 *Senses* **8**, 647-654 (2014).
- 822 64. Huertas M, Hagey L, Hofmann AF, Cerdà J, Canário AVM, Hubbard PC.
823 Olfactory sensitivity to bile fluid and bile salts in the European eel (*Anguilla*
824 *anguilla*), goldfish (*Carassius auratus*) and Mozambique tilapia (*Oreochromis*
825 *mossambicus*) suggests a 'broad range' sensitivity not confined to those
826 produced by conspecifics alone. *J Exp Biol* **213**, 308-317 (2010).

- 827 65. Zhang C, Brown S, Hara T. Biochemical and physiological evidence that bile
828 acids produced and released by lake char (*Salvelinus namaycush*) function as
829 chemical signals. *J Comp Physiol B* **171**, 161-171 (2001).
- 830 66. Hubbard P, Barata E, Ozório RA, Valente LP, Canário AM. Olfactory
831 sensitivity to amino acids in the blackspot sea bream (*Pagellus bogaraveo*): a
832 comparison between olfactory receptor recording techniques in seawater. *J*
833 *Comp Physiol A* **197**, 839-849 (2011).
- 834 67. Hubbard PC, Barata EN, Canario AV. Olfactory sensitivity to changes in
835 environmental $[Ca^{2+}]$ in the marine teleost *Sparus aurata*. *J Exp Biol* **203**,
836 3821-3829 (2000).
- 837 68. Velez Z, Hubbard PC, Barata EN, Canário AVM. Olfactory transduction
838 pathways in the Senegalese sole *Solea senegalensis*. *J Fish Biol* **83**, 501-514
839 (2013).
- 840 69. Dickson AG, Sabine CL, Christian JR. Guide to best practices for ocean CO₂
841 measurements. (North Pacific Marine Science Organization, Sidney, British
842 Columbia, Canada, 2007).
- 843 70. Meredith TL, Caprio J, Kajiura SM. Sensitivity and specificity of the olfactory
844 epithelia of two elasmobranch species to bile salts. *J Exp Biol* **215**, 2660-2667
845 (2012).

846 **Figure Captions**

847 **Figure 1. Behavior responses of European seabass (*Dicentrarchus labrax*) to a**
848 **5 min exposure to a predator odor (monkfish bile).** Behaviour recorded after
849 exposure to control (~420 μatm) or elevated CO_2 (~950 μatm) for 2, 7, and 14 days.
850 **a**, baseline activity after 2, 7, and 14 days exposure to control and elevated CO_2 . **b**,
851 change in activity before and during the first minute of exposure to predator odor
852 (dilution 1:1,000,000). Period of time spent freezing before and during 5 min
853 exposure to predator odor after 2 (**c**), 7 (**d**), and 14 (**e**) days of exposure. Values are
854 means \pm s.e.m. Asterisks indicate statistically significant differences compared to
855 control data obtained before exposure to predator odor. Crosses indicate statistical
856 significance between control and elevated CO_2 treatments ($p < 0.05$).

Figure 2. Elevated CO_2 decreases the olfactory sensitivity of European sea bass to amino acids, bile acids and body fluids. **a**, L-serine (N=6). **b**, L-cysteine (N=6). **c**, L-alanine (N=8). **d**, L-arginine (N=6). **e**, L-glutamate (N=6). **f**, cyprinol sulphate (N=10). **g**, scymnol sulphate (N=10). **h**, intestinal fluid dilutions (N=6). **i**, conspecific bile dilutions (N=6). **j**, alarm cue dilutions (N=8). Responses measured under control (blue) ($\text{pH } 8.15 \pm 0.01$, $476 \pm 14 \mu\text{atm}$) and elevated CO_2 (orange) ($\text{pH } 7.82 \pm 0.01$, $1122 \pm 19 \mu\text{atm}$). Values are expressed as % of the response to 10^{-4} M L-cysteine and represented as mean \pm s.e.m. Different letters indicate significant differences between the response to different concentrations of odorants ($p < 0.01$). † denotes differences between treatments ($p < 0.05$). See online Supplementary Fig S7 for raw traces of these responses.

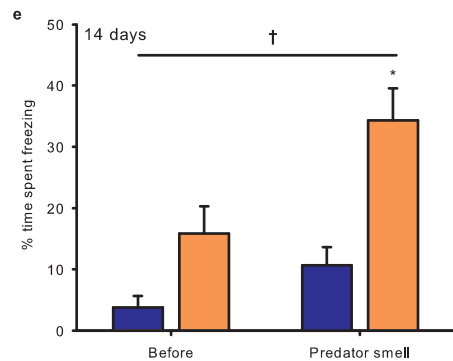
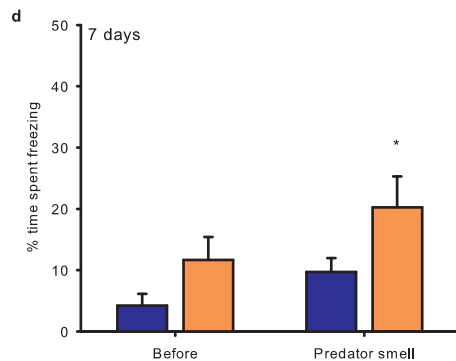
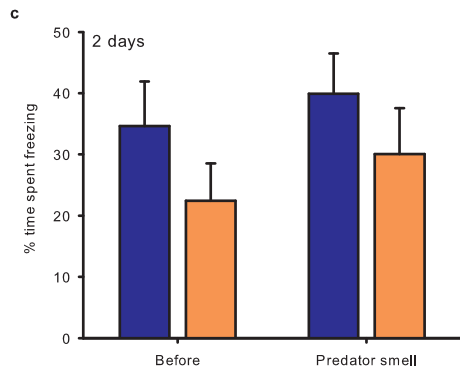
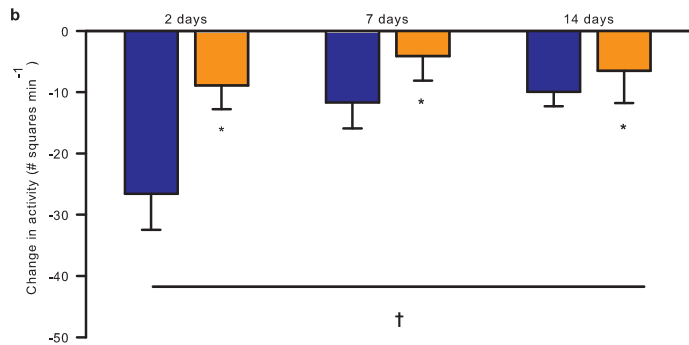
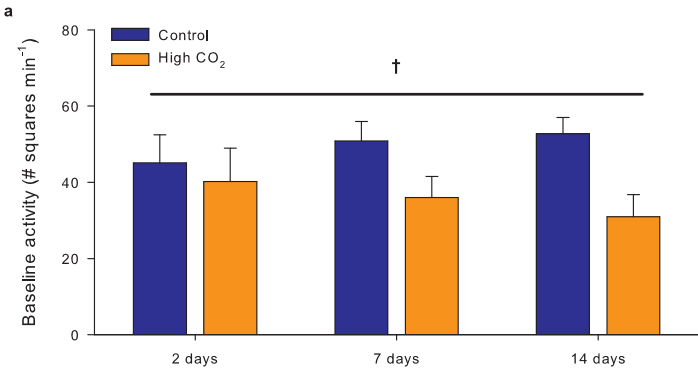
Figure 3. Acute exposure of European seabass to elevated CO₂ (~1000 µatm) decreases the amplitude of the olfactory response and increases the detection threshold of several odorants tested. Elevated CO₂ reduces the active space (represented by the blue sphere) of an odor by up to 80% (represented by the yellow dashed line) and the distance to a detectable odor source (arrow) by up to 42% in European sea bass. This suggests potentially drastic consequences on their ecology and survival (see Methods for calculations and assumptions, and Fig. 2 for amplitude response curves). Asterisks and crosses indicate statistically significant differences from the control group ($p < 0.05$) in detection threshold and amplitude, respectively. NS, not significant. Sea bass image Kovalevska/shutterstock.com

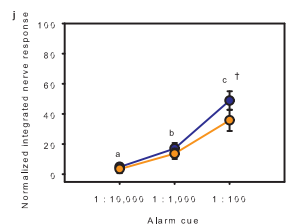
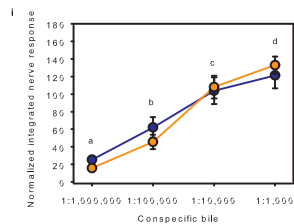
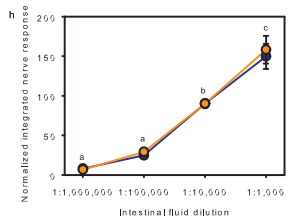
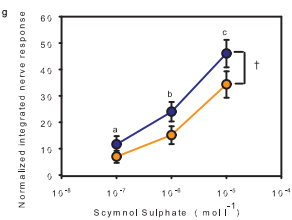
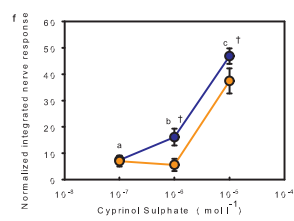
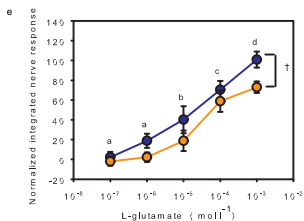
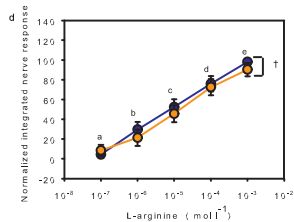
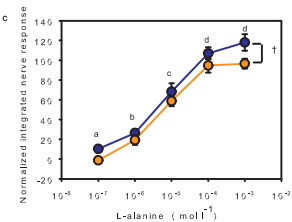
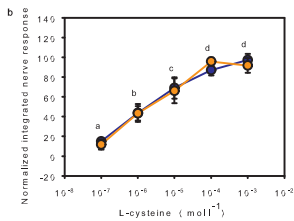
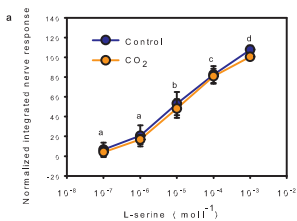
Figure 4. Differential regulation of genes in the olfactory epithelium and olfactory lobe of European sea bass exposed to control and high CO₂. Genes involved in neuronal growth (*efnb2a*) and development (*zak*) were significantly down-regulated in the olfactory epithelium. Additionally genes encoding for ion channels (*scn4*, *cacna2*, *chrna7* and *kcnn3*) responsible for maintaining cell excitability were also down-regulated in both the olfactory epithelium and the bulb. In the olfactory bulb there was also down regulation of glutamate ionotropic receptors (AMPA), mitogen activated protein kinase kinase (*map2k2*) and CAMKII indicative of long term depression (process involved in decreased synaptic plasticity). Moreover, olfactory receptor genes were downregulated in both the olfactory epithelium and the bulb, indicating no compensatory mechanism for loss of olfactory function and changes in the wiring of the olfactory system in juvenile sea bass. Arrows represent direct pathways of activation, and T bars represent direct pathways of repression.

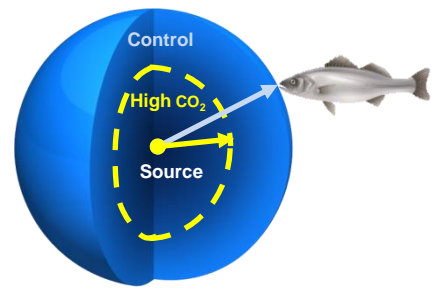
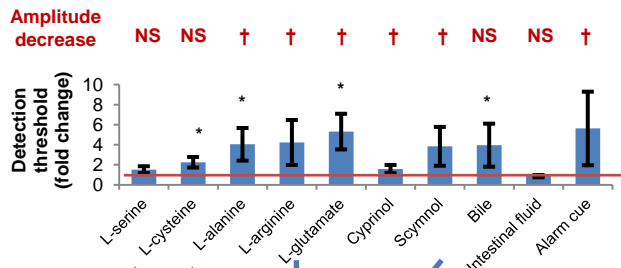
Note that the axons of the olfactory sensory neurons in the epithelium synapse with neurons in the olfactory bulb.

Figure 5. Proposed mechanism of action of CO₂-induced ocean acidification on fish behavior via the olfactory pathway. Ocean acidification has an acute effect on the binding of odorants to their receptors, decreasing both detection threshold and amplitude of the response. Long term exposure to high CO₂ decreases cell and neuron excitability, indicating less olfactory information is being transmitted from the olfactory epithelium to higher brain centers. In combination with a decrease in synaptic plasticity, this altered gene expression can affect behavior and learning in fish.

857







Gene expression

- ↑ 2 days
- ↑ 7 days
- ↓ 2 days
- ↓ 7 days

