1 Title

2 The persistence of carbon in the African forest understory

3

4 Authors

5	Wannes Hubau ^{1,2,3} *, Tom De Mil ^{1,2} *, Jan Van den Bulcke ^{2,4} , Oliver L. Phillips ³ , Bhély Angoboy Ilondea
6	^{1,5,6} , Joris Van Acker ^{2,4} , Martin J. P. Sullivan ³ , Laurent Nsenga ¹ , Benjamin Toirambe ¹ , Camille Couralet ¹ ,
7	Lindsay F. Banin ⁷ , Serge K. Begne ^{8,3} , Timothy R. Baker ³ , Nils Bourland ^{1,9,10,11} , Eric Chezeaux ¹² , Connie
8	J. Clark ¹³ , Murray Collins ¹⁴ , James A. Comiskey ^{15,16} , Aida Cuni-Sanchez ^{17,18} , Victor Deklerck ^{2,4} , Sofie
9	Dierickx ¹ , Jean-Louis Doucet ¹⁰ , Corneille E. N. Ewango ^{19,20,21} , Ted R. Feldpausch ²² , Martin Gilpin ³ ,
10	Christelle Gonmadje ²³ , Jefferson S. Hall ²⁴ , David J. Harris ²⁵ , Olivier J. Hardy ²⁶ , Marie-Noel D. Kamdem
11	^{8,27} , Emmanuel Kasongo Yakusu ^{1,21,2} , Gabriela Lopez-Gonzalez ³ , Jean-Remy Makana ¹⁹ , Yadvinder Malhi
12	²⁸ , Faustin M. Mbayu ²¹ , Sam Moore ²⁸ , Jacques Mukinzi ^{19,29} , Georgia Pickavance ³ , John R. Poulsen ¹³ , Jan
13	Reitsma ³⁰ , Mélissa Rousseau ^{1,11} , Bonaventure Sonké ⁸ , Terry Sunderland ^{9,31} , Hermann Taedoumg ⁸ , Joey
14	Talbot ³ , John Tshibamba Mukendi ^{1,21,32} , Peter M. Umunay ³³ , Jason Vleminckx ^{26,34} , Lee J. T. White ^{35,36,37} ,
15	Lise Zemagho ⁸ , Simon L. Lewis ^{3,17} , Hans Beeckman ¹
16	
17	* These authors contributed equally to this work.

18

19 Contact information

- 20 <u>whubau@gmail.com, wannes.hubau@africamuseum.be, tom.demil@ugent.be</u>
- 21 Service of Wood Biology, Royal Museum for Central Africa
- 22 Tervuren, Belgium, Leuvensesteenweg 13, 3080 Tervuren
- 23

24 Affiliations

- 25 1 Service of Wood Biology, Royal Museum for Central Africa, Tervuren, Belgium
- 26 2 UGent-Woodlab, Laboratory of Wood Technology, Department of Environment, Ghent University,
- 27 Ghent, Belgium

- 28 3 School of Geography, University of Leeds, Leeds, UK
- 29 4 Centre for X-ray Tomography (UGCT), Ghent University, Ghent, Belgium
- Institut National pour l'Etude et la Recherche Agronomique, Kinshasa I, Democratic Republic of
 Congo
- 32 6 Ecole Régionale Postuniversitaire d'Aménagement et de Gestion intégrés des Forêts et Territoires
- 33 tropicaux (ERAIFT), Kinshasa, Democratic Republic of Congo
- 34 7 Centre for Ecology and Hydrology, Penicuik, UK
- 35 8 Plant Systematic and Ecology Laboratory, Higher Teachers' Training College, University of
- 36 Yaounde I, Cameroon
- 37 9 CIFOR, Bogor, Indonesia
- 38 10 Forest Resources Management, Gembloux Agro-Bio Tech, University of Liege, Belgium
- 39 11 Resources and Synergies Development, Singapore, Singapore
- 40 12 Rougier-Gabon, Libreville, Gabon
- 41 13 Nicholas School of the Environment, Duke University, Durham, NC, USA
- 42 14 Grantham Research Institute on Climate Change and the Environment, London, UK
- 43 15 Inventory & Monitoring Program, National Park Service, Fredericksburg, VA, USA
- 44 16 Smithsonian Institution, Washington, DC, USA
- 45 17 Department of Geography, University College London, London, UK
- 46 18 Department of Geography and Environment, University of York, York, UK
- 47 19 Wildlife Conservation Society-DR Congo, Kinshasa I, Democratic Republic of Congo
- 48 20 Centre de Formation et de Recherche en Conservation Forestiere (CEFRECOF), Epulu, Democratic
 49 Republic of Congo
- 50 21 Faculté de Gestion de Ressources Naturelles Renouvelables, Université de Kisangani, Kisangani,
 51 Democratic Republic of Congo
- 52 22 Geography, College of Life and Environmental Sciences, University of Exeter, Exeter, UK
- 53 23 National Herbarium, Yaounde, Cameroon
- 54 24 ForestGEO, Smithsonian Tropical Research Institute, Panamá, Republic of Panama
- 55 25 Royal Botanic Garden Edinburgh, Edinburgh, UK

56	26	Service d'Évolution Biologique et écologie, Faculté des Sciences, Université Libre de Bruxelles,
57		Brussels, Belgium
58	27	Faculty of Science, Department of Botany and Plant Physiology, University of Buea, Buea,
59		Cameroon
60	28	Environmental Change Institute, School of Geography and the Environment, University of Oxford,
61		Oxford, UK
62	29	Salonga National Park, Kinshasa I, DR Congo
63	30	Bureau Waardenburg, The Netherlands
64	31	Faculty of Forestry, University of British Columbia, Vancouver, Canada
65	32	Faculté des Sciences Appliquées, Université de Mbujimayi, Mbujimayi, Democratic Republic of
66		Congo
67	33	Yale School of Forestry & Environmental Studies, New Haven, CT, USA
68	34	Department of Biological Sciences, Florida International University, FL, USA
69	35	Agence Nationale des Parcs Nationaux, Libreville, Gabon
70	36	Institut de Recherche en Ecologie Tropicale, Libreville, Gabon
71	37	School of Natural Sciences, University of Stirling, Stirling, UK
72		
73		
74		
75		
76		
77		
78		
79		
80		
81		
82		
83		

84 Main text

85

Quantifying carbon dynamics in forests is critical for understanding their role in long-term climate 86 regulation^{1,2,3,4}. Yet little is known about tree longevity in tropical forests^{3,5,6,7,8}, a factor that is vital for 87 estimating carbon persistence^{3,4}. Here we calculate mean carbon age (the period that carbon is fixed in 88 trees⁷) in different strata of African tropical forests using (i) growth-ring records with a unique 89 90 timestamp accurately demarcating 66 years of growth in one site and (ii) measurements of diameter 91 increments from the African Tropical Rainforest Observation Network (23 sites). We find that in spite 92 of their much smaller size, in understory trees mean carbon age (74 years) is greater than in sub-93 canopy (54 years) and canopy (57 years) trees, and similar to carbon age in emergent trees (66 years). 94 The remarkable carbon longevity in the understory results from slow and aperiodic growth as an adaptation to limited resource availability^{9,10,11}. Our analysis also reveals that while the understory 95 represents a small share (11%) of the carbon stock^{12,13}, it contributes disproportionally to the forest 96 97 carbon sink (20%). We conclude that accounting for the diversity of carbon age and carbon sequestration among different forest strata is critical for effective conservation management^{14,15,16}, and 98 for accurate modelling of carbon cycling⁴. 99

100

101 Investing in carbon storage and sequestration represent important climate change mitigation strategies³. Forests have a potential to provide both long-lived carbon stocks^{7,17} and long-term carbon sinks^{1,2}. To 102 103 successfully conserve forest carbon stock and increase forest carbon uptake, we must conserve carbon-rich forests and extend the forested land area³, but decision makers and managers also need to understand the 104 long-term behaviour of carbon within forests^{1,2,3}. Critical questions are: (i) how long does the carbon persist, 105 106 and (ii) where does it stay longest in the system? Carbon residence time is a direct function of tree longevity^{3,7,17}, but attempts to estimate tree age in tropical rainforests are relatively scarce and often highly 107 contested^{5,6,7,8}. Estimating the ages of the oldest trees in tropical forest stands is particularly subject to debate. 108 While some authors claim that broadleaved trees in the tropics may reach ages of 1000 years or older^{8,18}, 109 others estimate maximum ages of not more than 600 years^{5,6}. Furthermore, the oldest carbon in the system is 110 often assumed to be located in large trees⁸. Canopy and emergent trees contain a large proportion of the 111

stand-level biomass^{12,14} but large trees alone may not represent well the entire forest in terms of growth rates, tree lifespan and carbon persistence⁷. Canopy species grow faster¹⁹, but there is a general trade-off between growth and lifespan in organisms^{9,20}. Therefore, long-term carbon storage and sequestration in tropical rainforests may substantially depend on smaller trees too.

116

117 Here, we take advantage of a remarkable rediscovery of a historic forest inventory plot to probe the age 118 structure of African rainforests in a way that has not been possible to date. The Nkulapark plot was 119 established in 1948 in the southwest of the Democratic Republic of the Congo (Supplementary Fig.1). A 120 total of 6315 trees with diameter at breast height (DBH) >=5cm were tagged and DBH was measured 121 annually for 9 years. In 2014, we rediscovered and measured 450 surviving tagged trees, of which 55 were selected to measure growth ring series. We used the grown-in iron nail as a 1948 timestamp, giving accurate 122 estimates of the DBH growth rate (in mm yr⁻¹) and the rate of growth-ring formation (number of rings per 123 124 year) over a 66 year period for each tree (Fig.1, Supplementary Table 1). The age of each individual tree (in 125 years) was calculated as the total number of rings from pith to bark, divided by the rate of growth-ring formation (number of rings per year) (Fig.1, Fig.2). We used the five-class Crown Illumination Index of 126 Dawkins & Field (hereafter CII)^{21,22} to compare growth patterns and tree age among the different rainforest 127 strata: understory (CII=1), sub-canopy (CII=2,3), canopy (CII=4) and emergent stratum (CII=5). Understory 128 129 trees receive no direct light, sub-canopy trees receive lateral or restricted vertical light, canopy and emergent trees receive vertical light (Fig.3a).Supplementary analyses show that the rediscovered Nkulapark tree 130 131 dataset is adequate to compare age differences between forest strata (Supplementary Fig.2, Supplementary 132 Fig.3, Supplementary Discussion).

133

We find that the age of the 55 Nkulapark trees with growth-ring series ranges between 129 and 452 years (Fig.2, Supplementary Table 1). There is no clear linear relationship between tree age and their DBH (p=0.082, Fig.2a). Understory trees (CII=1) do not differ significantly in age from canopy and emergent trees (CII=4 or 5) (p=0.254, boxplot at the right of Fig.2b), while trees in sub-canopy classes (CII=2 or 3) are slightly younger than trees in both the understory and the canopy. Despite their small size, understory trees 139 (CII=1) can be surprisingly old. One *Microdesmis puberula* (TreeID=3545) has an estimated age of 329

140 years, with a DBH of just 156 mm (Supplementary Table 1).

141

To test if our findings hold true in a wider geographic context, we compared growth and age patterns among 142 the different forest strata in 23 Central African permanent forest inventory plots¹ (Fig.3, Table 1, 143 144 Supplementary Fig.1 and Supplementary Table 2). Selected plots had a similar species composition as the 145 Nkulapark. Plots are demarcated rectangles or squares of median size 1 ha where each tree is mapped, tagged and measured according to standard protocols^{1,2}. DBH of each tree with DBH>= 10 cm was measured at 146 147 least twice. Small trees that grew larger than 10cm during the monitoring period were recorded as recruits. 148 Trees that died were recorded. We used repeated diameter measurements to estimate the growth rate of each individual tree. We estimated tree age by dividing the final diameter (mm) by the growth rate (mm yr⁻¹), 149 assuming a constant growth rate over the lifetime of a tree⁷. We evaluated the robustness of this age 150 151 estimation method using the rediscovered Nkulapark trees as a reference (Supplementary Fig.4). Uncertainties are due to relatively short plot monitoring periods (average 9 years), yielding negative or zero 152 growth rates for 9.7% of the trees. To avoid unrealistic tree age estimates, we replaced slow growth rates by 153 a 'minimum allowed growth rate', defined as the xth percentile of the growth rate distribution within each CII 154 class. Sensitivity analysis showed that x=25 returned a realistic tree age distribution for our dataset 155 (Supplementary Table 3, Supplementary Discussion). Further analysis suggests that x may be lower if 156 monitoring periods are longer. Finally, we estimated tree-level mean carbon age as the average age of each 157 year ring, weighted by the carbon content of the ring⁷, with a year ring sequence deduced from the growth 158 159 rate (see equation 2 in the Methods).

160

The mean tree age for the 23 plots across Central Africa ranged between 131 and 284 years, with an overall mean of 229 years (95% bootstrapped confidence intervals: 212-244 years) (Table 1). Mean tree age in the understory (CII=1) is estimated to be 262 years, which is significantly older than the overall mean (p<0.001) and older than the mean age of the sub-canopy (CII=3, 187 years, p<0.001), the canopy (CII=4, 194 years, p<0.001) and emergent classes (CII=5, 221 years, p=0.021) (Table 1). Furthermore, mean carbon age at the tree level is 65 years (95% CI: 61-70). Carbon stored in the understory trees (CII=1) is estimated to be on average 74 years (69-79), which is significantly older than the overall mean (p<0.001) and older than carbon stored in the sub-canopy (CII=3, 54 years, p<0.001) and the canopy (CII=4, 57 years, p=0.001). However, the difference between mean carbon age in the understory (CII=1) versus the emergent class (CII=5, 66 years) is not significant (p=0.086) (Fig.3b and Table 1).

171

172 For each plot, we calculated the contribution of each forest stratum (in %) to the total above-ground biomass-carbon stock (AGC-stock) and to the total AGC-sink using standard methods^{1,2} (Table 1). The 173 AGC-stock (in Mg C ha⁻¹) represents the carbon reservoir in the system while the AGC-sink (in Mg C ha⁻¹ 174 yr⁻¹) represents the net change, calculated as AGC-productivity (additions to the system from tree growth) 175 minus AGC-mortality (losses)¹⁷. The understory (CII 1) contributes 11% to the total plot-level AGC-stock 176 177 (Table 1) and 20% to the plot-level AGC-sink. In contrast, the sub-canopy classes (CII 2 and 3) together 178 contribute about 25% to the AGC-stock, but nothing to the AGC-sink. The understory thus contributes 179 disproportionally to carbon sequestration, considering its relatively small share in the stock and compared 180

181 Discussion

182

Results from both the Nkulapark dataset (Fig.2) and the 23 inventory plots (Fig.3, Table 1) show that 183 understory (CII=1) and emergent (CII=5) trees are on average older than sub-canopy trees (CII=3). This 184 pattern can be explained by differences in life-history strategies¹⁵. Understory specialist species maintain low 185 growth rates for long periods, resulting in relatively small DBH at older age. Their adaptations allow them to 186 survive in the understory without the need to invest in rapid growth. The sub-canopy classes (CII=2 and 3) 187 188 are dominated by suppressed canopy specialists that survived a recruitment stage but didn't reach the canopy 189 yet. These trees are relatively young (Fig.3) and they experience limited light conditions, resulting in mortality rates equaling productivity rates. Canopy and emergent trees (CII= 4 and 5 respectively) are 190 canopy specialists that have been able to grow rapidly for long periods²³. This is in line with ref.¹⁹ who show 191 192 that growth rates in emergent trees are high and increase continuously. To test the assumption that high tree 193 and carbon age in the understory may be due to a difference in species composition, we classified species as 194 (i) understory specialists, (ii) non-specialists and (iii) canopy specialists (Supplementary Discussion). This

195 confirmed that understory specialists are on average smaller (p<0.001) but older (p<0.001) than canopy 196 specialists (Supplementary Table 4).

197

Furthermore, the Nkulapark dataset shows that there is a significant negative relationship between tree age 198 199 and growth-ring formation rate (Fig.2b). 91% of the trees in the dataset did not form a growth-ring every 200 year, suggesting an aperiodic growth pattern (shifts of growth to dormancy and back do not occur annually). 201 This aperiodic growth pattern is more prominent in understory trees, which formed fewer growth rings (0.36 202 rings per year) than canopy and emergent trees (0.61 rings per year) (p=0.01; boxplots at the top of Fig.2b). Aperiodic growth patterns translate into significantly slower long-term growth rates: understory trees (CII=1) 203 have a mean DBH growth rate of 1.51 mm yr⁻¹ versus 4.99 mm yr⁻¹ in emergent trees (p < 0.001) (Table 1). 204 The observed differences in growth periodicity (Fig.2b), growth rates (Table 1) and age patterns (Fig.3) 205 206 among the different forest strata are most likely a result of differences in survival mechanisms which are a function of resource availability^{11,24, 25} (Supplementary Discussion). 207

208

Our data show that growth and age distribution of tropical trees in mixed lowland rainforests is complex. 209 Large canopy trees are among the oldest in the rainforest, as suggested by several authors⁸, but as they 210 obtained their size and position by maintaining fast growth rates¹⁹, they are not significantly older than slow-211 212 growing small understory specialists (Fig.2, Fig.3, Table 1). Furthermore, while large canopy trees represent the largest share of the carbon stock^{12,13,14}, they suffer most during drought periods²⁶. In comparison, 213 understory trees represent a smaller carbon capital but they are less vulnerable to drought and contribute 214 disproportionally to carbon sequestration. As such, the understory provides long-term stability in forest 215 carbon cycling. Furthermore, the understory is more diverse than the canopy in terms of species 216 composition^{14,15,16}. Therefore, we recommend quantifying forest ecosystem services by considering forests as 217 a whole, with all forest strata providing specific services¹⁶. This is important in Central Africa, where the 218 219 demand for fuelwood and charcoal could severely affect the understory if only large trees were to be protected^{14,27}. 220

222	Final	ly, our results suggest that care is required with large-scale modelling of forest carbon accumulation
223	poten	tial and responses to different climate change scenarios ⁴ . There are two hypothesised responses to
224	increa	asing atmospheric CO ₂ concentrations, possibly explaining the long-term observed AGC-sink in tropical
225	forest	es ^{1,2} : (i) big trees increasing their asymmetric competition to the detriment of the rest of the stand, or (ii)
226	suppr	essed trees do best, as they are closer to their light compensation point ^{17,28} . Our results show that both
227	scena	rios occur in forest stands, with the understory (CII=1), the canopy (CII=4) and the emergent (CII=5)
228	classe	es contributing to carbon sequestration, while the sub-canopy classes (CII=2 and 3) do not contribute.
229	Fores	t and carbon cycle models will need to account for the diversity of carbon age and carbon sequestration
230	poten	tial among the different forest strata. Recent studies have found that forest structure can be predicted
231	from	the characteristics of canopy trees ^{13,14} , but our results suggest that temporal dynamics differ between
232	forest	strata. The long-term persistence of function depends on smaller trees too, which compared to their
233	statur	e contribute disproportionately heavily to long-term carbon storage, sequestration, and climate
234	resili	ence.
235		
236	Ref	erences
237	1.	Lewis, S. L. et al. Increasing carbon storage in intact African tropical forests. Nature 457, 1003–1006
238		(2009).
239	2.	Brienen, R. J. W. et al. Long-term decline of the Amazon carbon sink. Nature 519, 344–348 (2015).
240	3.	Körner, C. A matter of tree longevity. Science 355, 130–131 (2017).
241	4.	Galbraith, D. et al. Residence times of woody biomass in tropical forests. Plant Ecol. Divers. 6, 139-
242		157 (2013).
243	5.	Brienen, R. J. W., Schöngart, J. & Zuidema, P. A. Tree Rings in the Tropics : Insights into the
244		Ecology and Climate Sensitivity of Tropical trees. in Tropical Tree Physiology (eds. Goldstein, G. &
245		Santiago, L. S.) 6, (2016).
246	6.	Worbes, M. One hundred years of tree-ring research in the tropics – a brief history and an outlook to
247		future challenges. <i>Dendrochronologia</i> 20, 217–231 (2002).
248	7.	Vieira, S. et al. Slow growth rates of Amazonian trees: Consequences for carbon cycling. Proc. Natl.
249		Acad. Sci. 102, 18502–18507 (2005).
		9

- Chambers, J. Q., Higuchi, N. & Schimel, J. P. Ancient trees in amazonia. *Nature* **391**, 135–136
 (1998).
- Bigler, C. Trade-Offs between growth rate, tree size and lifespan of mountain pine (Pinus Montana)
 in the swiss national park. *PLoS One* 11, 1–18 (2016).
- 10. Kleczewski, N. M., Herms, D. A. & Bonello, P. Effects of soil type, fertilization and drought on
 carbon allocation to root growth and partitioning between secondary metabolism and ectomycorrhizae
 of Betula papyrifera. *Tree Physiol.* **30**, 807–817 (2010).
- 11. Sass-Klaassen, U. Tree physiology: Tracking tree carbon gain. *Nat. Plants* 1, 15175 (2015).
- 258 12. Bastin, J.-F. et al. Seeing Central African forests through their largest trees. Sci. Rep. 5, 1–8 (2015).
- Bastin, J.-F. *et al.* Pan-tropical prediction of forest structure from the largest trees. *Glob. Ecol. Biogeogr.* in press, (2018).
- 14. Lutz, J. A. *et al.* Global importance of large-diameter trees. *Glob. Ecol. Biogeogr.* 27, 849–864
 (2018).
- 15. Memiaghe, H. R., Lutz, J. A., Korte, L., Alonso, A. & Kenfack, D. Ecological Importance of SmallDiameter Trees to the Structure, Diversity and Biomass of a Tropical Evergreen Forest at Rabi,
 Gabon. *PLoS One* 11, 1–15 (2016).
- Burton, J. I., Ares, A., Olson, D. H. & Puettmann, K. J. Management trade-off between aboveground
 carbon storage and understory plant species richness in temperate forests. *Ecol. Appl.* 23, 1297–1310
 (2013).
- 17. Lloyd, J. & Farquhar, G. D. The CO2 dependence of photosynthesis, plant growth responses to
 elevated atmospheric CO2 concentrations and their interaction with soil nutrient status . I . General
 principles and forest ecosystems. *Funct. Ecol.* 10, 4–32 (1996).
- Laurance, W. F. *et al.* Inferred longevity of Amazonian rainforest trees based on a long-term
 demographic study. *For. Ecol. Manage.* **190**, 131–143 (2004).
- Stephenson, N. L. *et al.* Rate of tree carbon accumulation increases continuously with tree size. *Nature* 507, 90–93 (2014).
- 276 20. Wright, S. J. *et al.* Functional traits and the growth mortality trade-off in tropical trees. *Ecology*277 91, 3664–3674 (2013).

- 278 21. Synnott, T. J. A manual of permanent plot procedures for tropical rainforests. Tropical forestry
 279 papers 14, (1979).
- 280 22. Dawkins, H. C. & Field, D. R. B. A long-term surveillance system for British woodland vegetation.
 281 Oxford Forestry Institute Occasional Papers 1, (1978).
- 282 23. Hall, J. S., Harris, D. J., Medjibe, V. P. & Ashton, M. S. The effects of selective logging on forest
 283 structure and tree species composition in a Central African forest: Implications for management of
 284 conservation areas. *For. Ecol. Manag.* 183, 249–264 (2003).
- 285 24. Couralet, C., Van den Bulcke, J., Ngoma, L. M., Van Acker, J. & Beeckman, H. Phenology in
 286 functional groups of Central African trees. *J. Trop. For. Sci.* 25, 361–374 (2013).
- 287 25. Vico, G., Dralle, D., Feng, X., Thompson, S. & Manzoni, S. How competitive is drought
- 288 deciduousness in tropical forests? A combined eco-hydrological and eco-evolutionary approach.
- 289 Environ. Res. Lett. 12, 65006 (2017).
- 26. Bennett, A. C., McDowell, N. G., Allen, C. D. & Anderson-Teixeira, K. J. Larger trees suffer most
 during drought in forests worldwide. *Nat. Plants* 1, 15139 (2015).
- 292 27. FAO. *The charcoal transition : greening the charcoal value chain to mitigate climate change and* 293 *improve local livelihoods*. (Food and Agriculture Organization of the United Nations, 2017).
- 28. Lewis, S. L., Malhi, Y. & Phillips, O. L. Fingerprinting the impacts of global change on tropical
 forests. *Philos. Trans. R. Soc. B Biol. Sci.* **359**, 437–462 (2004).
- 296 29. Hietz, P. A simple program to measure and analyse tree rings using Excel, R and SigmaScan.
 297 *Dendrochronologia* 29, 245–250 (2011).
- 30. Benjamini, Y. & Hochberg, Y. Controlling the False Discovery Rate: A Practical and Powerful
 Approach to Multiple Testing. *J. R. Stat. Soc. Ser. B* 57, 289–300 (1995).

301 Correspondence

302 Correspondence and requests for materials should be addressed to W.H. (<u>whubau@gmail.com</u>).303

304 Acknowledgments

305 Nkulapark: W.H. and T.D.M. were both funded by the Brain program of the Belgian Federal Government 306 (BR/132/A1/AFRIFORD and BR/143/A3/HERBAXYLAREDD). The PhD project of T.D.M and the tenure 307 track of J.V.d.B. were supported by Ghent University Special Research Fund (BOF). Fieldwork was 308 sponsored by the King Leopold III fund for nature exploration and conservation. B.A.I. is supported by the 309 Institut National pour l'Etude et la Recherche Agronomiques en R.D.Congo (INERA- RDC- Luki) and the 310 Ecole Régionale Postuniversitaire d'Aménagement et de Gestion intégrés des Forêts et Territoires tropicaux 311 (ERAIFT Kinshasa). We thank WWF-RDC (Geert Lejeune), INERA and ERAIFT for facilitating fieldwork 312 in the Luki Reserve. We thank the INERA employees (Jean-Baptiste Ndunga, Jean-Maron, Fils Mbungu 313 Phaka, Leonard Ngoma, Noble, Plaside), the WWF ecoguards and the students of the Universities of 314 Kinshasa (UNIKIN) and Boma for assistance in the field. For assistance with datasets we thank Marlène De Groot, Kévin Lievens, Piet Dekeyser, Stijn Willen and José Kempenaers. The 23 permanent inventory plots: 315 this paper is also a product of the AfriTRON network, for which we are indebted to hundreds of institutions, 316 317 field assistants and local communities for establishing and maintaining the plots. This network has been 318 supported by the European Research Council (291585, "T-FORCES"- Tropical Forests in the Changing 319 Earth System, Advanced Grant to O.L.P. and S.L.L.), the Gordon and Betty Moore Foundation, the David and Lucile Packard Foundation, the European Union's Seventh Framework Programme (283080, 320 'GEOCARBON'), and Natural Environment Research Council (NERC) Consortium Grant 'TROBIT' 321 322 (NE/D005590/1), 'BIO-RED' (NE/N012542/1) and a NERC New Investigators Grant, the Royal Society, the 323 Centre for International Forestry (CIFOR) and Gabon's National Parks Agency (ANPN). We are indebted to 324 the University of Yaounde I, the National Herbarium of Yaounde, Rougier-Gabon, the Marien Ngouabi 325 University of Brazzaville, WCS-Congo, Salonga National Park, WCS-D.R.Congo, and the University of 326 Kisangani for logistical support in Africa.

327

328 Author contributions

- 329 W.H., T.D.M., J.V.d.B., J.V.A. and H.B. conceived and designed the Nkulapark study and S.L.L. conceived
- the AfriTRON plot network. T.D.M. and B.A.I. coordinated collection of Nkulapark data and wood cores.
- 331 T.D.M. and J.V.d.B. measured growth ring series. W.H. carried out the data analysis and wrote the paper.
- 332 S.L.L., O.L.P., T.R.B. and Y.M. conceived the Forestplots.net database, and most co-authors helped
- 333 collecting AfriTRON forest census data. S.L.L., B.S., S.B., A.C.S., W.H., T.S. and L.W.W. coordinated
- forest plots data collection. M.J.P.S., G.L.G., S.L.L., O.L.P., T.R.B. and G.P. contributed tools to analyze
- and curate data. All co-authors commented on or approved the manuscript.

336

337 Competing financial interests

338 The authors declare no competing financial interests.

341 Figures



Figure 1 Example of a wood core (Greenwayodendron suaveolens, TreeID=765) showing the 1948 nail 343 344 trace. The image at the top shows the full core. White lines indicate growth-ring boundaries, numbers 345 indicate growth-rings (counted from bark to pith), black arrows indicate important years. The bark to the 346 right of the figure indicates the year of sampling (2014). The dark discoloration in growth-rings 26 to 35 was caused by oxidates from the iron nail that were transported up and down in damaged vessels and fibers. The 347 348 right border of the discoloration accurately marks the start of the year 1948. There are 25 rings between the 349 bark and the 1948 nail trace, suggesting that this individual needed on average 2.6 years to form a ring. 350 Using this rate for the 53 rings that were formed before 1948, we find that the first ring in the core was 351 probably formed around 1811. The location of the pith is indicated by the black lines to the left, which follow the direction of the wood rays²⁹. This shows that the distance from the pith to the first ring boundary 352 353 in the core is about 11 mm. When using the average ring width of rings 78 to 68, we estimate that 7 rings are 354 missing. As such, this individual would be about 224 years old. The three close-ups at the bottom illustrate 355 wood anatomical details used to identify growth-ring boundaries (indicated by white triangles). Ring 356 boundaries in this species are demarcated by distended wood rays and flattening of the fibers. Black scale 357 bars cover 0.2 mm.

358



Figure 2 Variation in tree age inferred from growth-ring patterns in 55 trees where nail traces accurately mark the year 1948. Panel (a) shows the relation between tree age and DBH. Panel (b) shows the relation between tree age and the growth-ring formation rate (number of rings per year) between 1948-2014. In both panels, p-values of simple linear regression models are given in red. Brown dots represent understory individuals (Crown Illumination Index =1), grey dots represent sub-canopy individuals (CII=2 and 3), blue dots represent canopy and emergent trees (CII=4 and 5). Boxplots show the first quartile, the median value and the third quartile of the tree age distribution (vertical axis, boxplots to the right) and of the variables in the x-axis (horizontal boxplots at the top of the panels). P-values under boxplots resulted from two-sided Wilcoxon rank-sum tests. Outliers are marked with open circles.



373

374 Figure 3 Estimation of mean carbon age and contribution to aboveground carbon (AGC) sink per 375 crown illumination category for 23 permanent inventory plots in Central Africa. The 450 rediscovered Nkulapark trees were treated as an additional plot to estimate distributions of mean carbon age. Panel (a) 376 illustrates the Crown Illumination Index (CII) developed by Dawkins and Field (ref.²²), with yellow arrows 377 indicating reception of sunlight (drawing modified from ref.²¹). Understory trees (brown, CII=1) receive no 378 379 direct sunlight, sub-canopy trees (grey) receive lateral (CII=2) or restricted vertical (CII=3) light, canopy 380 trees (blue, CII=4) are almost completely exposed to vertical light and emergent trees (blue, CII=5) have a 381 crown that is completely exposed to vertical and lateral light. Boxplots in panels (b) and (c) show the 382 distribution of plot-level mean carbon age and the contribution to the total plot-level AGC-sink per CII class. 383 Boxplots represent the 25 % quartile, the median value and the 75 % quartile of the plot-level average ages. 384 Outliers are marked with open circles. Comparison among CII classes was performed with Dunn's rank-sum test using the Benjamini-Hochberg adjustment for multiple comparisons ³⁰. The grey dotted line shows the 385 386 overall (plot-level) average age.

387 Tables

388

metric	All trees	CII 1	CII 2	CII 3	CII 4	CII5	p-value
Tree age (yr)	229 (212-244)	262 (243-282)	210 (196-223)	187 (170-204)	194 (175-212)	221 (192-250)	0.021
Mean carbon age (yr)	65 (61-70)	74 (69-79)	60 (55-64)	54 (49-59)	57 (51-63)	66 (57-75)	0.086
Contribution to total AGC-sink	100	20 (9-32)	-8 (-21-5)	7 (-5-19)	28 (0-56)	52 (22-76)	<0.001
(%)							
Contribution to total AGC-stock	100	11 (8-14)	15 (12-18)	10 (7-13)	31 (25-39)	33 (27-39)	<0.001
(%)							
stem density (stems ha-1)	428 (392-465)	194 (165-228)	112 (94-130)	38 (33-44)	59 (46-70)	25 (19-31)	<0.001
DBH (mm)	313 (291-338)	161 (154-168)	235 (222-250)	299 (273-326)	435 (397-469)	700 (644-760)	<0.001
DBH growth (mm yr-1)	2.38 (2.17-2.63)	1.51 (1.35-1.69)	2.16 (1.96-2.38)	2.98 (2.69-3.31)	4.03 (3.71-4.34)	4.99 (4.32-5.68)	<0.001
wood density (g cm-3)	0.59 (0.56-0.61)	0.64 (0.62-0.65)	0.61 (0.6-0.63)	0.59 (0.56-0.61)	0.59 (0.56-0.61)	0.57 (0.53-0.61)	0.006
ratio evergreen:deciduous	3.27 (2.85-3.71)	4.72 (3.86-5.59)	4.61 (2.76-7.85)	3.62 (2.78-4.6)	3.03 (2.09-4.22)	1.78 (0.88-3.43)	<0.001
proportion evergreen trees (%)	55 (52-57)	56 (52-59)	55 (51-59)	56 (52-61)	51 (46-56)	44 (37-52)	0.01
proportion deciduous trees (%)	18 (16-20)	14 (12-17)	19 (16-22)	21 (17-25)	26 (21-31)	35 (28-41)	<0.001

³⁸⁹

390 Table 1 Estimation of tree age, mean carbon age, contribution to total aboveground carbon sink

391 (AGC-sink), contribution to total AGC-stock and leaf habit per crown illumination category, for 23

392 permanent forest plots. The 450 rediscovered Nkulapark trees were treated as an additional plot for
393 estimation of tree age and mean carbon age (first two rows). All metrics were averaged per plot and per CII
394 class. Reported values are means, 95% confidence intervals are given between brackets. Components of the
395 AGC-sink are AGC-productivity and AGC-mortality. Components of AGC-stock are diameter (DBH), wood
396 density and stem density. Comparison among CII classes was performed with Dunn's rank-sum test using the
397 Benjamini-Hochberg adjustment for multiple comparisons³⁰; the reported p-values compare CII1 and CII5.

399 Online Methods

400

401 Site description. The Nkulapark is a phenology and tree-growth monitoring plot covering 174 ha within the Luki Man and Biosphere reserve, located in the southern Mayumbe mountains in the Democratic Republic of 402 the Congo²⁴ (Supplementary Fig.1). The region is situated around 13.10°E and 5.62°S and experiences a 403 404 humid tropical climate with a dry season between mid-May and mid-October and a short dry season from 405 mid-December to mid-February. Yearly precipitation ranges from 649 mm to 1853 mm with a mean precipitation of 1173 mm. Temperature ranges between 19 °C and 30 °C with a mean temperature of 25.5 °C 406 ²⁴. The Nkulapark is situated almost entirely in a catchment with a valley and a ridge and includes several 407 408 microclimatic conditions. The semi-deciduous lowland forest consists of (i) mature forest dominated by 409 Prioria balsamifera, (ii) old regenerating forest dominated by Terminalia superba, (iii) mixed-species mature forest and (iv) modified forest patches 31,32 . 410

411

412 Nkulapark plot design in 1948. The Nkulapark was established and managed by the Institut National pour 413 l'Étude Agronomique du Congo Belge (INEAC), which was later renamed Institut National pour l'Étude et 414 la Recherche Agronomique en R.D. Congo (INERA, http://www.inera-drc.org). The person in charge of the tagging and the measurements was Léon Toussaint, who worked as a botanist in the Luki reserve between 415 1946 and 1952³³. The planning of the plot was first announced in the INEAC-Luki annual report of 1946³⁴. A 416 417 total of 29.2 km of observation paths were cut in the forest in 1947, following the contour lines of the Nkula river valley (Supplementary Fig.1)³¹. A total number of 6315 trees were tagged by the end of 1947^{31} , so we 418 assume that first wood formation after tagging occurred during the wet season that started in October 1948. 419 Tree selection was performed by randomly selecting trees from the pool of trees > 5 cm DBH in 1948. The 420 421 Nkulapark area was mapped, showing the locations of the largest tagged trees. From 1948 to 1957 yearly diameter measurements were performed on all tagged trees³⁵. Mortality events were recorded in the 422 423 datasheets. Trees were measured at a height of 1.3 m and the point of measurement (POM) was indicated on 424 the tree with a horizontal line of lead-based paint. For trees with buttresses or deformities, the POM was 425 raised to a point high enough to avoid the irregularities interfering with diameter measurements at 426 subsequent censuses. For trees with extremely high buttresses, diameters were estimated. For the same 6315

- 427 trees, weekly phenology observations were recorded. Phenological observations were done for leaf habit,
- flowering, fruiting and seed dissemination. The plot was abandoned in 1957 but the datasheets were kept in
 the library of the INERA station in Luki and digitized in 2008²⁴ and 2014.
- 430

431 Rediscovering Nkulapark trees in 2014. Each of the original 6315 Nkulapark trees were labelled during the 432 first census in 1948 with a zinc number tag that was attached to the tree using an iron nail of 8 cm long. A 433 part of these trees were indicated on the original 1948 map. During a first prospective field campaign in 434 august 2014, this map was digitized and georeferenced with QGIS (QGIS development team, 2016) using landmarks such as easily rediscovered trees, contour lines and observation paths that were still visible and 435 436 could be tracked with a Garmin 64s GPSmap (see Supplementary Fig.1). Based on this map, we pinpointed the approximate location of 1521 individuals that were recorded as alive during the last census in 1957. 437 438 During a second field campaign in September- October 2014, these 1521 individuals were searched for. 450 439 of them were found alive, 16 were found dead and the remaining 1057 could not be relocated and were assumed to be dead and rotten away, albeit some may have been missed (see Supplementary Fig.2, 440 441 Supplementary Fig.3, Supplementary Discussion for an in-depth analysis of survivorship rates). The original 442 1948 tree tags and nails of the rediscovered trees were either still present outside the trunk or detected inside 443 the tree using a metal detector (BHJS, Bounty Hunter, USA). Scars on the trunk indicated the presence of an 444 overgrown nail, and repelled number tags were sometimes found on or in the ground nearby the tree using 445 the metal detector. In most cases, the numbers on the tags were still readable. On 95% of the rediscovered 446 trees, the lead-based paint of the POM was still visible, allowing a representative DBH measurement. 447 Comparison of DBH, DBH growth rates and tree age distribution in the original dataset of Nkulapark trees versus the dataset of rediscovered trees, showed that the rediscovered dataset is slightly biased towards 448 449 discovering slower-growing trees, but this bias affected both the classes of understory and canopy 'specialist' 450 species (Supplementary Fig.3). Hence, the rediscovered tree dataset is representative to compare growth and 451 age patterns among the different forest strata in the Nkulapark area.

452

453 Sampling Nkulapark trees in 2014. Wood samples for growth-ring analysis were taken from rediscovered
454 trees if following criteria were met : (a) the nail was still present in the wood, either totally grown-in or

partly sticking out of the trunk, (b) the exact position of the nail could be identified visually or with the metal 455 detector, (c) the nail was not overgrown by excessive wound tissue, buttresses or other deformities. As such, 456 457 increment cores or stem discs were taken near the nail for 58 of the rediscovered trees. For each sampled tree, increment cores were taken a few centimetres above, below, to the left and to the right of the nail using 458 459 a 40cm Pressler bore. For each tree, two additional cores were taken at 120° from the nail trace along the 460 circumference of the tree. As such, 6 increment cores were available for each tree. This maximised the chance of sampling the pith of the tree. To study and describe the reaction of the wood after tagging, 461 462 additional larger wood samples containing the nail were extracted from 30 trees using a saw and a chisel. 463

Visualizing and measuring growth-ring series. For each wood core, growth-ring series were visualized using 464 two imaging methods as described by ref.³⁶: (i) first, density profiles were calculated from X-ray CT scans of 465 entire wood cores, then (ii) the cores were surfaced with a core microtome³⁷ and scanned using a flatbed 466 scanner (EPSON Perfection 4990 PHOTO). To obtain X-ray CT volumes, cores were scanned at 110 µm 467 resolution with the Nanowood CT facility from the Centre for X-ray Computed Tomography of Ghent 468 University (UGCT, <u>www.ugct.ugent.be</u>)³⁸, developed in collaboration with XRE (<u>www.xre.be</u>; now part of 469 470 the TESCAN ORSAY HOLDING a.s.). Reconstruction was performed with the Octopus software package on a GeForce GPX 770 4GB GPU^{38,39}. X-ray and flatbed scans were analysed using the toolchain for tree-471 ring analysis described by De Mil et al. (ref.³⁶). This toolchain semi-automatically indicates the growth-ring 472 boundaries and calculates growth-ring width series. Depending on the visibility of the growth-ring patterns, 473 either the X-ray or the flatbed scans were used to check growth-ring boundaries and measure growth-ring 474 475 widths. Growth-ring boundaries were distinguished using visual wood anatomical characteristics such as distended rays, flattened fibers and terminal parenchyma bands^{5,40,41}. For unclear ring boundaries, 476 477 microscopic thin sections were taken to study wood anatomy at high resolution using an Olympus BX60 478 microscope (Fig.1).

479

480 *Detecting the 1948 nail trace.* Discolorations or wound tissue formed as a reaction on the tagging were 481 visible in the cores taken near the nail. The surface of these stem discs was sanded to a few millimetres 482 above the nail and the anatomy was observed using an Olympus BX60 microscope. The nails were

remarkably well preserved inside the trees, probably due to cathodic protection of the iron by the zinc of the 483 tags. Hence, discolorations or wound tissue were visible in each of these samples. Discolorations were 484 485 recognizable as darkened tissue in otherwise light-coloured wood (Fig.1). Discolorations occurred in cells 486 that were formed before the nail was inserted, due to oxidation processes between the wood and the iron nail. 487 Water in vessels and fibers in the neighbourhood of the nail (especially those damaged by the nail) spread 488 these oxidates upwards and downwards. Therefore, the discolorations in these cells are also detectable on 489 samples that were taken a few centimetres under or above the nails. Vessels and fibers that were formed after 490 the tagging were not damaged, hence did not show discolorations. As such, the boundary of the discoloration 491 accurately serves as a timestamp indicating the year of tagging (1948). Furthermore, wound-induced 492 deformities occurred in the wood that was formed after the nail was inserted. This wound tissue is 493 characterized by increased woody productivity around the nail, forming a lump in the growth-rings that were 494 formed just after the nail was inserted. This lump formation is not present in the wood that was formed 495 before tagging.

496

497 Estimating tree age using growth rings and nail traces. Six cores were assessed for each tree. For some trees, 498 none of the cores contained the pith because (i) the tree radius exceeded the borer length or (ii) the pith was 499 eccentric and missed. In these cases, the core with the largest number of visible rings was used to estimate the total number of growth-rings. The missing core length from the end of this core to the pith was estimated 500 using the intersection of three lines of ring boundaries marked along the rays, as described by ref.²⁹ (see 501 502 Fig.1 for an example). We estimated the number of rings in the missing core part by dividing the missing 503 core length by the average ring width of the 5 oldest rings of the sampled core. We tested the robustness of 504 this method by rerunning the analysis using the average growth rates of the 5, 10, 15 and 20 oldest rings. We 505 found that the overall average tree age and the trends observed in Fig.2 are not sensitive to varying the 506 method used to estimate the number of rings in the missing part of the core. The core with the clearest nail 507 trace was used to count the number of rings formed after the year of tagging (1948). We used the number of 508 growth-rings formed between 1948 (nail mark) and 2014 (cambium) as a reference to calculate the number 509 of years the individual needed to form one ring (years per ring). We then multiplied this ratio with the total 510 number of rings formed before 1948 to estimate the age of the tree (Fig.1 and Fig.2). Three individuals were

511 not used for analysis because the estimated missing core length of each core exceeded 20 cm. As such,

512 growth-ring series of 55 trees were retained (Supplementary Table 1).

- 513
- 514 *Permanent forest inventory plots.* To test if our findings hold true in a wider geographic context, we
- 515 estimated mean tree age and carbon age of the different forest strata using plots from the
- 516 African Tropical Rainforest Observation Network (AfriTRON;
- 517 www.afritron.org). We selected 23 permanent forest plots located in four different Central African countries
- 518 (Cameroon, Gabon, Congo Brazzaville, D.R.Congo) (Supplementary Fig.1 and Supplementary Table 2).
- 519 Plots were selected if at least 65% of the trees belonged to species that also occur in the Nkulapark.
- 520 Furthermore, plots selected for analysis conformed to the following criteria^{1,2,42}: (1) plots had an actual plot
- 521 area of ≥ 0.2 ha, (2) plots were georeferenced, (3) all trees with DBH ≥ 100 mm were measured, (4) the
- 522 majority of stems were identified to species level, (5) plots had at least 2 censuses, (6) plots had a total
- 523 monitoring length of \geq 3 years, (7) plots were situated within structurally intact, apparently mature forest
- 524 (excluding young or open forests), (8) plots were free from major human impacts, (9) plots were located at
- \geq 50m from the anthropogenic forest edge, (10) altitude was below 1500 m.a.s.l., (11) mean annual air
- 526 temperature was $\geq 20.0^{\circ}$ C, (12) mean annual precipitation was ≥ 1000 mm yr⁻¹, (13) plots were located in
- 527 terra firme forest and (14) the CI index was recorded for each tree in the plot.. For analysis purposes, plots
- 528 smaller than 0.5 ha that were within 1 km of each other and located in similar forest types were merged (i.e.
- 529 the LME cluster). AfriTRON plot data are curated in the ForestPlots.net database⁴³, and subject to identical
- 530 quality control and quality assurance procedures. All calculations of plot data metrics described hereafter
- 531 were performed using the R statistical platform⁴⁴, version 3.2.1.
- 532
- 533 *Estimating AGC-stock.* For each tree and each census, Aboveground Biomass at the tree level (AGB, Mg 534 stem⁻¹) was estimated using a published allometric equation for moist forests including terms for diameter 535 (DBH, mm), dry wood density (ρ , g cm⁻³) and total tree height (H, m)⁴⁵ :

536
$$AGB = \frac{0.0673 \times (\rho \times \left(\frac{DBH}{10}\right)^2 \times H)^{0.976}}{1000}$$
 (formula 1).

Wood density values were derived from the dryad database (www.datadryad.org). Stems were matched to 537 species-specific wood density values or the mean values for the genus or family, following ref.¹ and⁴³. 538 Heights were calculated using a single height-diameter model (Weibull) for central African lowland terra 539 firme forests published by ref.⁴⁶, using commands implemented in the R-package BiomasaFP⁴⁷. 540 Aboveground Biomass-Carbon (referred to as AGC) is considered as 47% of the AGB following IPCC 541 recommendations⁴⁸. For each individual tree in the plot dataset, AGC-stock was calculated as the mean of the 542 543 first and last censuses. Finally, we calculated the AGC-stock in each CII class in each plot, divided by the total plot-level AGC-stock and multiplied by 100 to express the results as % of the total plot-level AGC-544 545 stock.

546

Estimating AGC-sink. For the calculation of AGC-sink, only the first and the last censuses were used for 547 each plot. First, AGC- productivity (Mg C stem⁻¹ yr⁻¹) for each stem surviving the monitoring period, was 548 549 calculated as the difference between its total AGC at the end census minus the total AGC at the start census 550 of the interval, divided by the census interval length (yr). AGC-productivity for stems recruited during the monitoring period (i.e. reaching DBH ≥100 mm), was calculated in the same way, assuming DBH=0 mm at 551 the start of the interval. AGC-mortality for each tree that died during the monitoring period (Mg C stem⁻¹ yr⁻¹ 552 ¹) was calculated as the AGC at the start of the monitoring period, divided by the total monitoring length 553 (yr). AGC-productivity at the stand level (Mg C ha⁻¹ yr⁻¹) was then calculated as the sum of tree-level 554 productivity estimates of all survivors and recruits. AGC-mortality at the stand level (Mg C ha⁻¹ yr⁻¹) was 555 556 calculated as the sum of tree-level mortality estimates of all dead trees in the subset. We corrected for 557 unobserved components of biomass growth and mortality due to census interval length effects, as discussed by ref.², ref.⁴⁹ and ref.⁵⁰. A method to correct productivity and mortality rates for these uncertainties was 558 developed by ref.⁴⁹. This correction accounts for (i) trees that recruit and die within the same interval (i.e. 559 560 unobserved recruits) and (ii) growth of trees that grow and die within the interval (i.e. unobserved biomass 561 growth from mortality, which is not recorded because dead trees are not measured). As such, for each census interval, we calculated unobserved recruitment and unobserved mortality components (Mg ha^{-1} yr⁻¹) using 562 the formulas proposed by ref.⁴⁹ and both components were added to both the AGC-productivity and AGC-563 564 mortality estimates. Estimates of the unobserved biomass components usually accounted for less than 3% of

the total AGC-productivity and AGC-mortality. The AGC-sink was calculated as stand-level AGCproductivity minus AGC-mortality. Finally, we calculated the AGC-sink for each CII class in each
plot, divided by the total plot-level AGC-sink and multiplied by 100 to express the results as % of the total
plot-level AGC-sink. Absolute sink values are not reported because the total sink is likely to be
overestimated due to a relatively small sample size (23 plots), possibly capturing occasional natural
disturbance events. Commands to calculate AGC-stock and AGC-sink are implemented in the BiomasaFP R
package ⁴⁷.

572

Tree age inferred from DBH growth rates in permanent inventory plots. For each tree within the permanent 573 forest inventory plots, we estimated the age by dividing the DBH (mm) in the final census with the DBH 574 growth rate $(mm yr^{-1})$ of the tree itself. For each tree, we averaged the DBH growth rate over all census 575 576 intervals preceding the last census. This method uses the actual growth rate of each tree, which is accurate for healthy trees but returns unrealistic age estimates for trees with (i) slightly negative growth rates, (ii) zero 577 growth rates or (iii) very slow growth rates. Such slow growth rates may be recorded in all DBH classes. 578 First, slow growth rates in large DBH classes may occur when a tree is diseased or at the end of its life 579 580 (senescence). These growth rates may not be representative for the total lifespan of these trees. Secondly, 581 small growth rates may be recorded for small suppressed trees if their growth is so slow that it cannot be 582 recorded with sufficient precision using standard census procedures. Hence, these growth rates are replaced by growth rates that yield a realistic tree age estimate. As such, we chose a 'minimum allowed growth rate' 583 per CII class, following ref.⁷. We calculated the minimum allowed growth rate for each CII class as the 25th 584 585 percentile (=first quartile) of the growth rate distribution within the CII class. For each tree with a growth rate slower than the minimum allowed rate, we replaced the growth rate by the minimum allowed growth 586 587 rate of the CII class. We conducted a sensitivity analysis to check how results vary when varying the minimum allowed growth rate: we reran the analysis using the 10th, 15th, 20th, 25th and 30th percentile of the 588 growth rate distribution within each CII class as a minimum allowed growth rate (Supplementary Discussion 589 590 and Supplementary Table 3). We used the average tree age in the dataset of 450 rediscovered Nkulapark 591 trees as a reference to evaluate the tree-age estimation method based on DBH growth rates (Supplementary 592 Figure 4).

593

Mean carbon age. As a tree grows, it increasingly stores more carbon. Carbon near the bark of the tree is younger than carbon in the pith. As such, mean carbon age of a tree does not equal total tree age. For each tree in the selected 23 plots, we deduced a year-ring series using its final DBH and the DBH growth rate (mm yr⁻¹), assuming a constant growth rate over the lifetime of the tree . We then used this deduced year-ring series to calculate mean tree-level carbon age using the same formula as ref.⁷:

599 mean carbon age =
$$\frac{\sum_{i=1}^{n} (C_i \times A_i)}{\sum_{i=1}^{n} (C_i)}$$
 (formula 2),

600 with:

- 601 C_i = carbon content of the ith ring (kg),
- $602 \qquad A_i \qquad = age \ of \ the \ i^{th} ring \ (yr),$

603 n = nr of rings.

604

The carbon content of the ith ring is calculated as the carbon content of a tree with the DBH of ring i minus the carbon content of a tree with the DBH of ring i-1.

607

Classification of tree species and statistical analysis. To distinguish between understory, sub-canopy, canopy 608 609 and emergent trees in the Nkulapark and the permanent forest plots, we used the Crown Illumination Index (CII) of Dawkins & Field (ref.²²). Fig.3 illustrates the 5 classes with a drawing modified from ref.²¹. In each 610 plot, each tree was attributed to one of the CII classes. The index was attributed in the field, mostly during 611 612 one census and mostly by a single person. We estimated mean tree age, mean carbon age, mean AGC-stock 613 and mean AGC-sink for each of the CII classes. None of the metrics reported in Table 1 meet the criterion of 614 homogeneity of variances (Bartlett test). Therefore, differences among the CII classes were tested using the 615 non-parametric Dunn's rank-sum test. To avoid the multiple comparison problem, we used the Benjamini-Hochberg p-value adjustment³⁰ ('dunn.test' package in R^{44}). 616

617

Data availability. The input data and R-scripts to generate the figures and tables are available for download
 using the following private link : https://figshare.com/s/06c793575d3b52ef5574. Images of wood cores are

620	availa	ble using the following link : <u>https://figshare.com/s/e6101fe7d330f8ea140a</u> . This folder also contains
621	all anı	notation documents needed to visualize growth ring boundaries on the wood samples (please consult
622	the RI	EADME document for guidelines). Wood samples used to conduct this analysis are stored in the
623	Tervu	ren xylarium
624	(<u>http:/</u>	/www.africamuseum.be/collections/browsecollections/naturalsciences/earth/xylarium). These samples
625	may b	e studied, within the Tervuren xylarium, upon request addressed to the curator H.B.
626	(<u>hans.</u>	beeckman@africamuseum.be) or the corresponding author W.H. (<u>whubau@gmail.com</u>).
627		
628	Refe	rences (methods only)
629	31.	INEAC. Rapport Annuel INEAC-Luki. (1947).
630	32.	Coppieters, G. Inventaris van het archief van de Rijksplantages en de Regie der Plantages van de
631		Kolonie, het Nationaal Instituut voor de Landbouwkunde in Belgisch-Congo en de
632		Documentatiedienst voor Tropische Landbouwkunde en Plattelandsontwikkeling 1901-1999. (2013).
633	33.	Académie Royale des Sciences d'outre-mer. Biographie Coloniale Belge/Biographie Belge d'Outre-
634		Mer, Tome IX. (2015).
635	34.	INEAC. Rapport Annuel INEAC-Luki. (1946).
636	35.	INEAC. Rapport Annuel INEAC-Luki. (1948).
637	36.	De Mil, T., Vannoppen, A., Beeckman, H., Van Acker, J. & Van Den Bulcke, J. A field-to-desktop
638		toolchain for X-ray CT densitometry enables tree ring analysis. Ann. Bot. 117, 1187–1196 (2016).
639	37.	Gärtner, H. & Nievergelt, D. The core-microtome: A new tool for surface preparation on cores and
640		time series analysis of varying cell parameters. Dendrochronologia 28, 85–92 (2010).
641	38.	Dierick, M. et al. Recent micro-CT scanner developments at UGCT. Nucl. Instruments Methods
642		Phys. Res. Sect. B Beam Interact. with Mater. Atoms 324, 35–40 (2014).
643	39.	Vlassenbroeck, J. et al. Software tools for quantification of X-ray microtomography at the UGCT.
644		Nucl. Instruments Methods Phys. Res. Sect. A Accel. Spectrometers, Detect. Assoc. Equip. 580, 442–
645		445 (2007).

- 646 40. Worbes, M. Tree-Ring Analysis. *Encycl. For. Sci.* 586–599 (2004).
- 647 41. Tarelkin, Y. et al. Growth-ring distinctness and boundary anatomy variability in tropical trees. IAWA

- 648 *J.* **37**, 275–294 (2016).
- 42. Phillips, O. & Baker, T. Field manual for plot establishment and remeasurement RAINFOR. *Rainfor*22 (2009). doi:10.13140/RG.2.1.1735.7202
- 43. Lopez-Gonzalez, G., Lewis, S. L., Burkitt, M. & Phillips, O. L. ForestPlots.net: A web application and research tool to manage and analyse tropical forest plot data. *J. Veg. Sci.* **22**, 610–613 (2011).
- 44. R Development Core Team. R: A language and environment for statistical computing. (2008).
- 654 45. Chave, J. *et al.* Improved allometric models to estimate the aboveground biomass of tropical trees.
 655 *Glob. Chang. Biol.* 20, 3177–3190 (2014).
- 46. Feldpausch, T. R. *et al.* Tree height integrated into pantropical forest biomass estimates.
- 657 Biogeosciences 9, 3381–3403 (2012).
- 47. Lopez-Gonzalez, G., Sullivan, M. & Baker, T. BiomasaFP package. Tools for analysing data
- downloaded from ForestPlots.net. R package version 1.1. Available at
- 660 http://www.forestplots.net/en/resources. (2015).
- 48. Aalde, H. *et al.* IPCC Guidelines for National Greenhouse Gas Inventories. Volume 4: Agriculture,
 Forestry and Other Land Use. Chapter 4: Forest Land. *Forestry* 4, 1–29 (2006).
- 49. Talbot, J. *et al.* Methods to estimate aboveground wood productivity from long-term forest inventory
 plots. *For. Ecol. Manage.* 320, 30–38 (2014).
- 665 50. Clark, D. a *et al.* Measuring net primary production in forest concepts and field methods. *Ecol. Appl.*666 11, 356–370 (2001).
- 667

668

. .

- 669
- 670
- 671

Supplementary Figures



Supplementary Fig.1. Map showing the locations of the Nkulapark forest and the selected forest plots. Panel (a) is a map of tropical West-Central Africa showing the position of the Nkulapark and the 23 selected permanent sampling plots, grouped into 11 clusters. Red dots indicate position of the clusters, dot size corresponds to the square root of total plot size (ranging from 0.4 to 9 ha). The grey area shows the cover of all closed, lowland and submontane (<1500 masl) forests, swamps and forest mosaics, according to Global Land Cover map 2000¹. Boxes show histograms of average tree age per class for each plot cluster. The upper limit of each box corresponds to 400 years. The Crown Illumination Index was recorded in each of the 23 plots. For each cluster, the boxes contain 5 bars representing the average age per CII class (brown bars: CII 1; grey bars: CII 2 or 3; blue bars: CII 4 or 5 respectively). The three-letter codes refer to cluster codes (see **Supplementary Table 2** for a full plot list), with the number of plots indicated between brackets. Panel (b) shows a map of the Nkulapark plot, reconstructed from the original map (1948) and showing the contour lines (370 masl in dark red, 140 masl in dark green), the Nkula river (blue) and the position of the 450 rediscovered trees (red dots).



Supplementary Fig.2. Evolution of survivorship in the dataset of Nkulapark trees along the observation paths that were visited during rediscovery fieldwork in 2014. Panel (a) shows the repartition of all trees per class at the start of the monitoring period in 1948 (und. spec. = understory specialist species, non-spec. = non-specialists; can. spec. = canopy specialists). Panel (b) shows the evolution of the stem count over the subsequent censuses between 1948-1957, including an exponential relation between the number of survivors and time for each group of species. Panel (c) shows the repartition of the survivors at the end of the annual monitoring period in 1957. Panel (d) shows the repartition of the expected surviving trees in 2014, assuming that the exponential relations parameterized for 1948-1957 were maintained. Panel (e) shows the repartition of the actually rediscovered trees in 2014. Numbers on top of each bar represent the exact stem count for each group of trees, and the percentage of the total dataset. Note the different y-axis in panels a-c versus panels d-e.



Supplementary Fig.3. Comparison of the full Nkulapark dataset (a-d) with the dataset of rediscovered trees (e-h) and the dataset of rediscovered trees with growth ring data (i-l). All metrics are presented for the year 1948 (the year of the first census). Boxplots represent distribution of metrics for populations of trees belonging to three different classes. Boxplots show the 25 % quartile, the median value and the 75 % quartile. Outliers are displayed in grey. Comparison between understory, canopy and non-specialists was performed with Dunn's rank-sum test using the Benjamini-Hochberg adjustment for multiple comparisons².



Supplementary Fig.4. Testing the performance of the age estimation method used in permanent sampling plots. Panel (a) compares tree age estimated from DBH growth rates (used for age estimation in plots) with tree age estimated from growth-ring counts (reference age), using the 55 trees where nail traces accurately mark the year 1948 (Fig.1 and Fig.2). The grey line represents the ideal 1:1 relation, the red line represents a simple linear regression model, with p-value and slope given in red. Boxplots show the first quartile, the median value and the third quartile of the total tree age population. p-values under boxplots resulted from two-sided Wilcoxon rank-sum tests. Outliers are marked with open circles. Panel (b) shows the difference between the reference age versus the estimated age. In both panels, brown dots represent understory individuals (Crown Illumination Index =1), grey dots represent sub-canopy individuals (CII=2 and 3), blue dots represent canopy and emergent individuals (CII=4 and 5).

Supplementary Tables

Supplementary Table 1. Number of growth rings, growth-ring formation rate, tree age and mean carbon age of all 55 Nkulapark trees with 1948 nail traces used for this analysis.

Supplementary Table 2. List of all 23 plots included in this analysis, with geographic coordinates, plot size, dates of first and last census, and the main researchers for each plot.

Supplementary Table 3. Sensitivity analysis for the growth correction method used to estimate tree age based on the individual DBH growth rates. This table shows how mean tree age in the understory (CII 1) and the emergent class (CII 5) varies when varying the minimum allowed growth rate used for correction. The minimum allowed growth rate is defined as the 10th, 15th, 20th, 25th and 30th percentile of the age distribution within each CII class. As a reference, we use long-term growth rates (spanning 66 years) and age estimations (at the reference year 1948) of 450 rediscovered trees in the Nkulapark.

	Understory trees (CII = 1)			Emergent trees (CII = 5)					All trees					Nkulapark		
	10pc 1	5pc	20pc	25pc	30pc	10pc	15pc	20pc	25pc	30pc	10pc	15pc	20pc	25pc	30pc	rediscovered trees
mean age (yr)	360 3	360	326	265	238	297	276	238	214	198	314	296	264	222	197	224
minimum DBH growth rate (mm yr ⁻¹)	0.09 0).09	0.16	0.26	0.32	0.47	0.8	1.26	1.61	1.89	0.09	0.09	0.16	0.26	0.32	0.27
% trees > 1000 years old	17.9 1	17.9	6.05	1.12	0.43	9.3	3.86	0.35	0	0	13.2	9.34	3.2	0.6	0.19	1.04

Supplementary Table 4. Comparison of tree age, mean carbon age, DBH, DBH growth and leaf habit between understory specialists, non-specialists and canopy specialists in the 23 selected plots. Mean values are given in bold, bootstrapped 95% confidence intervals are given between brackets. Comparison between understory, canopy and non-specialists was performed with Dunn's rank-sum test using the Benjamini-Hochberg adjustment for multiple comparisons².

metric	understory specialists	non- specialists	canopy specialists	p-value		
tree age (yr)	228 (225-233)	203 (193-211)	203 (182-223)	<0.001		
mean carbon age (yr)	65 (64-66)	59 (56-62)	60 (54-66)	<0.001		
DBH (mm)	205 (203-208)	311 (302-320)	488 (459-515)	< 0.001		
DBH growth (mm yr ⁻¹)	2.08 (2.02-2.13)	3.56 (3.4-3.74)	7.12 (6.29-8.04)	<0.001		
ratio evergreen:deciduous	6.49 (4.73-8.48)	1.64 (1.11-2.26)	0.46 (0.21-0.79)	<0.001		
proportion evergreen trees (%)	60 (57-64)	47 (41-54)	20 (10-31)	< 0.001		
proportion deciduous trees (%)	13 (10-16)	43 (35-52)	56 (41-72)	<0.001		

Supplementary Discussion

Understory vs canopy 'specialist' species

To test the assumption that high tree and carbon age in the understory (CII 1) (**Fig.3, Table 1**) may be due to a larger proportion of understory 'specialist' species, we classified all species in the 23 selected permanent forest plots in three classes : understory specialists, non-specialists and canopy specialists. This classification was based on the frequency of reported Crown Illumination Index (CII) for each species. Understory specialists were defined as those species from which >80% of the individuals received a CII of 1 or 2 in the selected plots where CII was reported. Canopy specialists were defined as those species from which >50% of the individuals received a CII of 4 or 5. Non-specialists are all species that were not classified as an understory or canopy specialist using the criteria described above.

We used the selected permanent forest plots to calculate mean tree and carbon age for each of the three species classes (**Supplementary Table 2** and **Supplementary Fig.1**). We find that tree age and mean carbon age are significantly largest in the understory specialists (p<0.001, **Supplementary Table 4**), while DBH of the understory specialists (205 mm) is less than half the DBH of the canopy specialists (488 mm)(p<0.001). Furthermore, DBH growth rates of the understory specialists (2.04 mm yr^{-1}) are significantly smaller than for the canopy specialists (5.31 mm yr^{-1}). This confirms that canopy specialists are large and fast-growing trees which especially occupy intermediate and higher forest strata. Furthermore, **Supplementary Table 4** shows that the species classified as understory specialists are mostly evergreen (ratio evergreen:deciduous = 6.49) while canopy specialists are mostly deciduous (ratio evergreen:deciduous = 0.46). This confirms that evergreenness may be an adaptation to living in the understory^{3,4}. Similar relations between resource limitation, adaptation strategies and longevity are reported in the arctic⁵ and alpine⁶ regions, where growth is limited due to temperature constraints.

Uncertainties in estimating tree age in rediscovered trees

Representativeness of the dataset of 450 rediscovered trees

Our dataset of 450 rediscovered trees is a subset of the original 6315 Nkulapark trees that were randomly selected in 1948. The dataset of rediscovered trees is possibly biased towards sampling the more robust trees that survived the 66 year period between tagging and sampling (survivorship bias, described by ref.⁷). Furthermore, there may have been a tendency towards discovering slow-growing trees, with nails and tags present near the bark, or a bias by failing to discover small understory trees along the observation paths (discovery biases). We tested these possible sampling biases by comparing the full original Nkulapark dataset with the dataset of rediscovered trees. The CI index was not developed yet during the period when the full dataset was monitored. Therefore, we compared the differences between the understory 'specialist' species versus the canopy 'specialist' species in both datasets (cfr. species classification in **Supplementary Table 4**).

First, we tested the survivorship bias by comparing survivorship rates of the understory specialists with survivorship rates of the canopy specialists (**Supplementary Fig.2**). For this, we used the dataset of trees that were located along the observation paths that were visited during the rediscovery campaign in 2014. Along the visited paths, 1731 trees were tagged and monitored during the 1948-1957 period (**Supplementary Fig.2a**). 1521 of these trees were still alive in 1957 (**Supplementary Fig.2b and 2c**), and were searched during the 2014 rediscovery campaign. The ratio between understory specialists and canopy specialists changed slightly in favor of the understory trees during the 1948-1957 monitoring period (53% versus 55% understory specialists, **Supplementary Fig.2c**). We tested an exponential relation expressing the long-term survivorship rate using the 1948-1957 data (**Supplementary Fig.2b**). If these exponential relations would have been maintained during the 1957-2014 period, then we would expect a total number of 677 survivors, which would be 39% of the initial 1948 population (**Supplementary Fig.2d**). Of these, we would expect 432 understory survivors (64%), versus 133 canopy survivors (20%). As such, the ratio between understory and canopy survivors would have continued to change in favor of the understory trees, which could have caused a survivorship bias in the rediscovered tree dataset. However, the ratio between understory (54%) versus canopy specialists (29%) in the actual rediscovered tree dataset is close to the initial

1948 ratio (**Supplementary Fig.2e**). This is because the actual rediscovery campaign yielded almost exactly the expected number of canopy specialists (130 trees), while we probably missed a substantial number of the understory specialists because small understory trees are harder to find (**Supplementary Fig.2e**). We conclude that the expected survivorship bias in our dataset of rediscovered trees may have been compensated by a discovery bias, yielding a final dataset of rediscovered survivors (**Supplementary Fig.2e**) that is comparable to the initial dataset of tagged trees (**Supplementary Fig.2a**) in terms of ratio between understory and canopy specialists.

Secondly, we further tested the rediscovery bias by comparing following metrics in **Supplementary Fig.3**: (i) DBH, (ii) DBH growth rate and (iii) tree age. We calculated DBH distributions and tree ages at the reference year 1948 for both the original and the rediscovered tree datasets (1948 was the start of the monitoring period in the Nkulapark). From the original dataset of 6315 tagged trees, 1935 trees could not be used for these calculations because (i) the diameter was never measured between 1948-1957, or (ii) measured growth rates were unreliable due to insufficiently recorded changes in point of measurement (POM)(see Methods for more information on POM). We find that the DBH distribution in both the full and the rediscovered tree datasets are comparable, with about 55% of the trees in the lowest DBH classes (0-20 cm) (Supplementary Figures 3a and 3e) and comparable overall mean DBH (347 mm in the original dataset versus 359 mm in the rediscovered dataset; Supplementary Figures 3b and 3f). Furthermore, the proportions of evergreen trees, deciduous trees and understory specialists are comparable, while the proportion of canopy specialists is slightly higher in the dataset of rediscovered trees. Also, average tree age is comparable (199 years in the original dataset versus 211 in the rediscovered dataset; Supplementary Figures 3d and 3h). However, growth rates are smaller in the rediscovered tree dataset than in the original dataset, both in the understory (2.3 vs 2.7) and in the canopy class (3.4 vs 4.4) (Supplementary Figures 3c and 3g). In summary, results in Supplementary Fig.3 suggest that the rediscovered tree dataset may be biased slightly towards slow-growing trees (discovery bias), but the dataset is still representative because (i) the bias affected both the understory and the canopy species, (ii) the bias did not alter the DBH distribution and species composition, and (iii) the bias did not increase mean tree age. Therefore, we conclude that our dataset of 450 rediscovered trees is representative to compare growth and age patterns among the different forest strata in the Nkulapark area (**Fig.3**).

Representativeness of the dataset of 55 rediscovered trees with growth-ring records

The 55 trees with prominently visible nail traces (**Fig.1**) and accurate growth-ring records are a subsample of the 450 rediscovered trees (see **Methods**). Average tree age of these 55 trees is 163 years in 1948 (**Supplementary Fig.3l**), which is younger than the average tree age in the full rediscovered tree dataset (211 years; **Supplementary Fig.3d**). This is because the length of the core borer used for sampling was 50cm and only trees with DBH<105 cm could be sampled (**Supplementary Table 1**). However, this sampling bias affected both the understory and the canopy classes (**Supplementary Fig.3**). Furthermore, the small difference in age between understory and canopy observed in the full dataset (**Supplementary Fig.3d**) and in the rediscovered dataset (**Supplementary Fig.3h**), is still maintained in the dataset of 55 trees with growth ring series (**Supplementary Figure3i**). As such, the dataset used for **Fig.2** is biased towards sampling younger trees, but the dataset is representative to compare age differences between different forest strata.

Assumption of constant rate of growth-ring formation

Our method of estimating tree age using a combination of tree-rings and grown-in nails is unique in the tropics. We build on the methods and assumptions of classic tree-ring analysis, but with the advantage of a very old reference point provided by the nail trace. Our assumption that the rate of growth-ring formation (rings per year) remains constant over the lifetime of a tree can be criticized. However, a similar assumption is made by most studies using growth-rings to estimate tree age. Occurrence of missing rings, false rings and wedging rings is a well-known problem in tropical dendrochronology^{8,9}. Therefore, species are carefully selected. Cambial pinning experiments^{10,11} and bomb-peak radiocarbon dating¹² are used to assess whether growth ring formation may be an annual process. If annuality of growth-ring formation can be confirmed for a certain amount of individual trees of a certain species, growth-rings are assumed to be annual for each individual of the same species^{13,14}. Our assumption is robust in this respect because our dataset represents an accurate pinning experiment covering 66 years, while classic cambial pinning experiments mostly cover only

a few years^{10,11}. Furthermore, as the average age of the 55 trees in our dataset is 229 years, our cambial pinning experiment on average covers 29% of the lifetime of a tree.

Uncertainties in estimating tree age based on growth-rate measurements

Assumption of constant growth rate

Our age estimation method using DBH divided by DBH growth rates, is based on the assumption of a constant growth rate over the lifetime of each individual. Yet, growth rates most likely vary over the lifetime of a tree¹⁵. In mature rainforests, shade-tolerant trees may survive as suppressed trees in the understory, then start growing more rapidly when light conditions change (gap formation). The age of these trees may be underestimated when growth rates are measured at the end of their life. In contrast, light-demanding gap colonizers may grow rapidly at the start of their life, then reach an optimum and slow down after that (senescence). The age of these trees may be overestimated when growth rates are measured at the end of their life. In both cases, errors occur especially in the largest DBH classes.

To test the performance of our method, we compared the age estimations based on DBH growth rates with the age estimations based on growth-ring series, using the 55 trees with nail traces as a reference dataset. **Supplementary Fig.4a** suggests that using the DBH growth rate of each tree is a robust method to infer tree age from DBH growth rates, as there is an almost 1:1 relationship between the reference age (from growth-rings) and the estimated age (from DBH growth rates) (slope=0.978). Furthermore, **Supplementary Fig.4b** shows that the absolute difference between the reference age and the age estimated using DBH growth rates increases with tree age. However, both underestimations and overestimations occur, so there is no significant trend between the difference and the reference age (p=0.826, **Supplementary Fig.4b**). As such, we suggest that over- and underestimation errors may compensate one another when estimating average tree age on the population level using the DBH growth rate method. A similar conclusion was drawn by ref.¹⁶ who used radiocarbon dates to test the robustness of a DBH growth rate method for age estimation.

Sensitivity to growth rate correction

Estimating tree age by dividing DBH by the DBH growth rate returns unrealistic age estimates for trees with (i) slightly negative growth rates, (ii) zero growth rates or (iii) very slow growth rates. Hence, these growth rates are replaced by a 'minimum allowed growth rate' per CII class, following ref.¹⁶. We calculated possible minimum allowed growth rates for each CII class as the 10th, 15th, 20th, 25th, and 30th percentile of the growth rate distribution within the CII class. Mean tree age decreases when using higher thresholds for minimum allowed growth rate (314 years when using 10th pc; 197 years when using 30th pc) (**Supplementary Table 3**). To select a realistic minimum allowed growth rate, we (i) compared the average tree age with a reference dataset and (ii) we evaluated the number of trees exceeding 1000 years.

First, we calculated average tree age in the dataset of 450 rediscovered Nkulapark trees as a reference. For each of these trees, we were able to calculate a growth rate over an interval spanning 66 years. Because this is a very long growth interval, none of the growth rates were negative, zero, or unrealistically small (i.e. minimum 0.27 mm yr⁻¹; **Supplementary Table 3**). We calculated tree age at the start of the monitoring period, by dividing the DBH in 1948 by the 1948-2014 growth rate. Average tree age is 224 years, which is most comparable to the average tree age in the 23 forest plots when using a 25th percentile threshold (222 years; **Supplementary Table 3**). Secondly, we counted the number of old trees. While maximum tree age in the tropics is a matter of debate^{17,18}, a wide assessment of tree-ring based age determinations suggested that maximum ages of 1000 years or more are very rare¹⁷. When using a 10th pc threshold, 13.2 percent of the trees exceeded 1000 years, which is probably not realistic. When using higher thresholds, the number of trees exceeding 1000 years decreased, and dropped below 1% when a 25th percentile for each CII class returns the more realistic average tree age estimates, with only 0.6% of the trees exceeding 1000 years and without overcorrecting the growth rates.

Effect of monitoring period

For each tree in the 23 permanent sampling plots, we used the total monitoring period (first and last census) to estimate a growth rate representative for the entire tree lifespan. The average monitoring length of these

trees is 7.8 years. However, each tree has been measured several times over the monitoring period, yielding growth rate estimates over shorter census intervals, with an average census interval length of 3.0 years. A two sample t-test shows that the average tree-level growth rate calculated over the entire monitoring period (2.44 mm yr⁻¹) is significantly higher than the average tree-level growth rate calculated over the shorter census intervals (2.28 mm yr⁻¹) (p<0.001), using the full dataset of trees in the 23 permanent inventory plots. This is because the method of measuring DBH with a DBH tape is sensitive to slow growth, which is often recorded as a zero or negative growth rate if the interval is very short. In extreme cases where the census interval is below 1 year, subsequent censuses may be performed during different seasons. When the first census is performed during the wet season and the second census during the dry season, growth rates can be smaller than the shrinkage of the wood.

As such, 25.1% of the growth rates in the dataset using short census intervals are zero or negative. In contrast, 9.7% of the growth rates in the dataset using the entire monitoring period are zero or negative. Furthermore, none of the growth rates estimated using the 66 year interval in the Nkulapark dataset (450 rediscovered trees) were below 0.27 mm yr⁻¹ (**Supplementary Table 3**). As a result, none of the Nkulapark growth rates needed to be corrected to get meaningful age estimates, while 25% of the trees in the permanent inventory plots needed a DBH growth rate correction (**Supplementary Table 3**). Our analysis suggests that if the monitoring periods of the permanent inventory plots would have been shorter, a larger share of the trees would have needed a DBH growth rate correction. We conclude that our method of using the entire monitoring period for growth rate estimation yields a more reliable estimate of long-term growth rates than using specific (shorter) census intervals.

Diversification of survival mechanisms

The observed differences in growth periodicity, growth rates and age patterns among the different forest strata (**Fig.2, Fig.3, Table 1**) are most likely a result of differences in survival mechanisms. Cambial activity and wood formation are directly associated with leaf activity, which is in turn a function of resource availability^{19,20}. Small trees in the understory are facing challenging environments, characterized by light, water and nutrient limitation and the constant threat of mechanical injuries (e.g. by falling branches).

Therefore, understory trees form disproportionally large root systems to capture enough nutrients and water^{3,21,22}. Hence, photosynthetic assimilates are more often allocated to roots than to stems. Furthermore, understory trees form significantly heavier wood (0.635 versus 0.571 g cm⁻³, **Table 1**), providing mechanical support and possibly reducing cavitation risk^{23,24}. However, investing in solid wood comes at the cost of reduced volumetric growth. More importantly, most understory trees keep their leaves throughout the year²⁵. This allows them to capture light at any time, and it reduces nutrient loss³. As such, evergreenness allows them to survive under conditions of limited light and nutrient availability⁴. Hence, they lack a clear annual periodicity in leaf turnover, which in turn translates into aperiodic and slow growth patterns.

In contrast, large canopy (CII 4) and emergent (CII 5) trees grow in an environment with less light constraints during most of their lifespan, allowing them to grow rapidly. However, they face more challenging hydraulic requirements and higher evaporative demands due to higher solar radiation and a more exposed canopy^{26,27}. An important drought-avoidance strategy is deciduousness. Large canopy and emergent trees often shed their leaves, especially in areas with pronounced dry seasons^{27,28}. Deciduous canopy and emergent trees lose their leaves (almost) every year and experience an annual period of cambial dormancy, producing a clear annual growth-ring pattern.

We used the Nkulapark phenology data published by ref.²⁵ to calculate the number of evergreen and deciduous trees in the selected permanent forest plots. This shows that evergreen trees are more abundant in the understory (p=0.01), while deciduous trees are more abundant in the canopy (p<0.001) (**Table 1**), which is in line with ref.²⁷. The mean ratio between evergreen and deciduous trees is 4.72 in the understory (CII 1). This ratio is much lower in the emergent class (CII 5, ratio=1.78, p<0.001, **Table 1**). Also, the individuals in the canopy and emergent class in the dataset of 55 trees with nail traces (**Fig.2**) mostly belong to prominent deciduous species (especially *Prioria balsamiferum*), while the understory trees with aperiodic growth patterns belong to species that are mostly evergreen (e.g. *Microdesmis puberula, Greenwayodendron suaveolens, Hexalobus crispiflorus*). As such, diversification of survival mechanisms may explain patterns observed in mixed lowland rainforests on terra firme soils. However, this may not apply for monodominant

forests (e.g. dominated by *Gilbertiodendron dewevrei*) or edaphic forest types (e.g. swamps), which were excluded from this analysis.

References

- 1. Mayaux, P. et al. A land cover map of Africa. European Commission, Joint Research Center (2003).
- 2. Benjamini, Y. & Hochberg, Y. Controlling the False Discovery Rate: A Practical and Powerful Approach to Multiple Testing. *J. R. Stat. Soc. Ser. B* **57**, 289–300 (1995).
- Vico, G., Dralle, D., Feng, X., Thompson, S. & Manzoni, S. How competitive is drought deciduousness in tropical forests? A combined eco-hydrological and eco-evolutionary approach. *Environ. Res. Lett.* 12, 65006 (2017).
- 4. Aerts, R. The advantages of being evergreen. *Tree* **10**, 402–407 (1995).
- Wilmking, M. *et al.* Continuously missing outer rings in woody plants at their distributional margins. *Dendrochronologia* 30, 213–222 (2012).
- Lu, X., Camarero, J. J., Wang, Y., Liang, E. & Eckstein, D. Up to 400-year-old Rhododendron shrubs on the southeastern Tibetan Plateau: Prospects for shrub-based dendrochronology. *Boreas* 44, 760– 768 (2015).
- 7. Brienen, R. J. W., Gloor, E. & Zuidema, P. A. Detecting evidence for CO 2 fertilization from tree ring studies: The potential role of sampling biases. *Global Biogeochem. Cycles* **26**, (2012).
- Brienen, R. J. W., Schöngart, J. & Zuidema, P. A. Tree Rings in the Tropics: Insights into the Ecology and Climate Sensitivity of Tropical trees. in *Tropical Tree Physiology* (eds. Goldstein, G. & Santiago, L. S.) 6, (2016).
- 9. Worbes, M. Tree-Ring Analysis. *Encycl. For. Sci.* 586–599 (2004).
- Sass, U., Killmann, W. & Eckstein, D. Wood Formation in Two Species of Dipterocarpaceae in Peninsular Malaysia 1. *IAWA J.* 16, 37–1 (1995).
- De Mil, T. *et al.* Cambial activity in the understory of the Mayombe forest, DR Congo. *Trees Struct. Funct.* 31, 49–61 (2017).
- Hua, Q., Barbetti, M. & Rakowski, A. Z. Atmospheric Radiocarbon for the Period 1950–2010.
 Radiocarbon 55, 2059–2072 (2013).

- Groenendijk, P., Sass-Klaassen, U., Bongers, F. & Zuidema, P. A. Potential of tree-ring analysis in a wet tropical forest: A case study on 22 commercial tree species in Central Africa. *For. Ecol. Manage.* 323, 65–68 (2014).
- 14. Fichtler, E., Goettingen, U. & Clark, D. A. Age and Long-term Growth of Trees in an Old-growth Tropical Rain Forest, Based on Analyses of Tree Rings and 14 C. *Biotropica* **35**, 306–317 (2003).
- 15. Stephenson, N. L. *et al.* Rate of tree carbon accumulation increases continuously with tree size. *Nature* **507**, 90–93 (2014).
- Vieira, S. *et al.* Slow growth rates of Amazonian trees: Consequences for carbon cycling. *Proc. Natl. Acad. Sci.* 102, 18502–18507 (2005).
- Worbes, M. One hundred years of tree-ring research in the tropics a brief history and an outlook to future challenges. *Dendrochronologia* 20, 217–231 (2002).
- Chambers, J. Q., Higuchi, N. & Schimel, J. P. Ancient trees in amazonia. *Nature* 391, 135–136 (1998).
- 19. Sass-Klaassen, U. Tree physiology: Tracking tree carbon gain. *Nat. Plants* 1, 15175 (2015).
- 20. Fonti, P. *et al.* Studying global change through investigation of the plastic responses of xylem anatomy in tree rings. *New Phytol.* **185**, 42–53 (2010).
- 21. Bigler, C. Trade-Offs between growth rate, tree size and lifespan of mountain pine (Pinus Montana) in the swiss national park. *PLoS One* **11**, 1–18 (2016).
- Kleczewski, N. M., Herms, D. A. & Bonello, P. Effects of soil type, fertilization and drought on carbon allocation to root growth and partitioning between secondary metabolism and ectomycorrhizae of Betula papyrifera. *Tree Physiol.* **30**, 807–817 (2010).
- Chao, K. J. *et al.* Growth and wood density predict tree mortality in Amazon forests. *J. Ecol.* 96, 281–292 (2008).
- 24. Chave, J. et al. Towards a worldwide wood economics spectrum. Ecol. Lett. 12, 351–366 (2009).
- 25. Couralet, C., Van den Bulcke, J., Ngoma, L. M., Van Acker, J. & Beeckman, H. Phenology in functional groups of Central African trees. *J. Trop. For. Sci.* **25**, 361–374 (2013).
- 26. Bennett, A. C., McDowell, N. G., Allen, C. D. & Anderson-Teixeira, K. J. Larger trees suffer most during drought in forests worldwide. *Nat. Plants* **1**, 15139 (2015).

- 27. Condit, R. *et al.* Quantifying the deciduousness of tropical forest canopies under varying climates. *J. Veg. Sci.* 11, 649–658 (2000).
- 28. Fauset, S. *et al.* Drought-induced shifts in the floristic and functional composition of tropical forests in Ghana. *Ecol. Lett.* **15**, 1120–1129 (2012).